# SCIENTIFIC REPORT PREPARED IN PARTIAL FULFILLMENT OF FISHERIES RESTORATION GRANT Grantee Agreement No: P1210323 

Results of regional spawning ground surveys and estimates of total salmonid redd construction in the Humboldt Bay, Humboldt County California, 2015-2016.

## Prepared by

Humboldt State University Sponsored Programs Foundation In partnership with State of California Department of Fish and Wildlife

Author
Colin W. Anderson ${ }^{1}$

Principle Investigator
Darren M. Ward ${ }^{2}$

[^0]
## Acknowledgements:

Thanks to the many research assistants, Americorps Watershed Stewards Program members, and California Conservation Corps Veterans Program members who worked on this project. Thanks to the staff at Humboldt State University Sponsored Programs. Thanks to Seth Ricker and Dave Hankin for their insights into study design considerations and the ecology of salmonids. Thanks to Sharon Powers for help with GIS analysis and maps. For keeping the machine running we would like to thank Mary Kuehner. Without access to the watershed none of this would be possible. Thanks to the many landowners especially Green Diamond Resource Company, Humboldt Redwood Company, and Bureau of Land Management .

This project would not exist without the groundwork laid by the California Monitoring Program project developers and partners.

The majority of this work was funded by the California Department of Fish and Wildlife's Fisheries Restoration Grants Program.


#### Abstract

Field staff conducted spawning ground surveys in 23 reaches within the Humboldt Bay watershed between November 17, 2015 and April 21, 2016. Individual stream reaches were surveyed an average of 4.1 times at an average return interval of 29 days. We observed 2 Chinook salmon, 740 coho salmon, 34 steelhead trout, and 23 unidentified live fish. A total of 2 Chinook salmon, 339 coho salmon, 2 steelhead trout, and 53 unidentified carcasses were found. We identified 454 individual redds of which 171 were assigned to a species. We estimate 617 (430-849, 95\% Confidence Interval) coho redds and 97 (21-184, 95\% Confidence Interval) steelhead redds in Humboldt Bay as expanded from the sampled reaches. We estimate 323 (251-394, 95\% Confidence Interval) coho redds within the Freshwater Creek Life Cycling Monitoring Station as the sum of a complete census of all reach estimates.


## CONTENTS

1 INTRODUCTION ..... 1
1.1 Background ..... 1
1.2 Study Area ..... 3
2 METHODS ..... 4
2.1 Sample Frame Construction ..... 4
2.2 Reach Selection ..... 4
2.3 Reach Survey Protocol ..... 4
2.4 Assigning Species to Unknown Redds ..... 4
2.5 Estimation of Within-Reach Redd Abundance ..... 5
2.6 Estimation of Total Redd Abundance Within the Sample Frame ..... 5
3 RESULTS ..... 6
3.1 Reach Survey Frequency ..... 6
3.2 Fish Observations ..... 6
3.3 Redd Observations ..... 11
3.4 Total Redd Abundance ..... 11
4 DISCUSSION ..... 14
5 REFERENCES ..... 17
6 APPENDIX ..... 19

## List of Figures

Figure 1. Humboldt Bay sampling frame. Reaches are depicted with alternating colors, and reach codes are presented as table references. ............................................................. 3

Figure 2. Stacked Bar Plot of live fish (A), individual redd, and carcass (C) observations by week. Please note redds don't include those from extended surveys in March and April.

Figure 3. Discharge in Little River, used as a proxy for Humboldt Bay, December 2015 through February 2016.15

## List of Tables

Table 1. Summary of spawning ground survey statistics of the mean number of days
between reach surveys and the maximum number of days between surveys.................... 7
Table 2. Live fish observations by calendar week. .............................................................. 8
Table 3. Carcass observations by calendar week............................................................... 9
Table 4. Descriptive statistics of live fish observations (L) and carcasses (C)................. 11
Table 5. Confusion matrix of the known species redds in columns crossed with LOOCV
redd predictions in rows.................................................................................................. 11
Table 6. Counts of newly constructed known redds by calendar week. ........................... 12
Table 7. Counts of anadromous salmonid redds within each survey reach. Survey reach numbers are labled in figure 1. ........................................................................................ 13
Table 8. Estimated number of redds in surveyed reaches in Humboldt Bay and in the Freshwater Creek Life Cycle Monitoring Station (LCS) where all reaches were surveyed.
Estimated Descriptive statistics are between sample reaches........................................... 13
Table 9. Reach survey frequency by year in Humboldt Bay tributaries........................... 14
Table 10. Redd estimates with $95 \%$ confidence intervals for three species during six years
in Humboldt Bay.............................................................................................................. 16
Table 11. Redd estimates with $95 \%$ confidence intervals for three species during six years in Freshwater Creek Life Cycle Monitoring Station......................................................... 16

Table 12. Counts of newly constructed redds by calendar week for non-target species. 19

## 1 INTRODUCTION

### 1.1 Background

Pacific Salmon (Oncorhynchus sp.) have experienced marked decline in abundance over the last 60 years. Due to this decline, coho salmon (Oncorhynchus kisutch) in the Southern Oregon and Northern California Coasts (SONCC) Evolutionary Significant Unit (ESU) were federally listed as threatened pursuant to the Endangered Species Act (ESA) in 1997 (NMFS 1997). This federal listing status was reviewed and reaffirmed in 2005 (NMFS 2005) and is currently under another review. The California Fish and Game Commission found coho salmon populations within the SONCC warranted listing as threatened species under the California Endangered Species Act (CESA) (CDFG 2002). All California steelhead (O. mykiss) south of the Klamath River are Federally ESA listed (NMFS 2006) and coastal Chinook salmon (O. tshawytscha) south of the Klamath River to the Russian River are federally ESA listed (NMFS 1999). In 2004 the California Department of Fish and Game developed a recovery strategy for coho salmon populations within California (CDFG 2004). This recovery strategy is intended to direct management and restoration actions needed to recover the species, and provides basin by basin threat assessments and attempts to prioritize management and restoration actions needed to recover the species. The Federal government requires that listed species have recovery plans developed that require objective, measurable criteria which when met, would result in the species being removed from the listing (16 USC 1531, Endangered Species Act 1973). Recovery of salmon and steelhead listed under the Federal and California ESAs can be measured in part on the increase in abundance of spawning adults (Good et al. 2005). Delisting will depend on abundance thresholds and the connectivity of populations to one another (Spence et al. 2008, Williams et al. 2008).

The California Department of Fish and Wildlife and the National Oceanic and Atmospheric Administration ~ Fisheries recognize four key parameters for assessing the long term viability of salmonid populations. These viable salmonid population (VSP) parameters are population size, population growth rate (productivity), population spatial structure, and life history diversity (McElhany et al. 2000). Monitoring these population parameters is essential to evaluating the success of recovery efforts.

To address data needs for viability assessment, the California Department of Fish and Wildlife (CDFW) and the National Oceanic and Atmospheric Administration ~ Fisheries (NOAA~Fisheries) cooperatively developed the draft Coastal California Salmonid Monitoring Plan (CMP)(Adams et al. 2011). Two complimentary tasks are considered high priority in the northern monitoring area and form the foundation of the CMP approach. The first task consists of probabilistic sampling of stream reaches within a defined region using spawning ground surveys (SGS) to establish the regional status and trends of adult salmonid abundance. The second task develops intensively monitored Life Cycle monitoring Stations (LCS) nested within the regional sample frame of the SGS.

Freshwater Creek is a LCS nested within the regional sample frame of the Humboldt Bay SGS. The nesting of the SGS within the LCS investigates the relationship between SGS observations and adult escapement.

This report summarizes the results of yearly abundance and survival monitoring efforts from November 2015 to June 2016, as well as integrates all years of project data to make inference on population trajectories.
1.2 Study Area


Figure 1. Humboldt Bay sampling frame. Reaches are depicted with alternating colors, and reach codes are presented as table references.

## 2 METHODS

### 2.1 Sample Frame Construction

Sampling frames were constructed with several factors presented by Garwood and Ricker 2011 including; documented historical salmonid distributions, documented barriers to anadromy, stream gradient, and field reconnaissance. Sample frames were constructed around coho salmon and may exclude habitat which is used by Chinook salmon and steelhead trout.

Each sampling frame was further divided into streams and then into survey reaches. From each sampling frame a portion of reaches was selected to survey. Reaches were assigned numbers in a fashion that ensured that selected survey reaches were balanced spatially within each stream (Garwood and Ricker 2011). Short sub-reaches less than 1000 meters were combined with the reach into which they flowed into.

### 2.2 Reach Selection

Reaches were selected for sampling from each frame using a General Randomized Tessellation Stratified (GRTS) sampling algorithm (McDonald 2003). GRTS sampling ensures a randomized spatially balanced draw from each stream. We sampled 22 reaches of a total possible 33 reaches to achieve a sampling rate of approximately $66 \%$.

### 2.3 Reach Survey Protocol

Spawning ground surveys protocol closely followed Gallagher et al. (2007). Teams of two walked upstream on small streams or boated downstream in large stream reaches. Observations include carcasses, live fish, and redds. Live and dead fish were identified to species when possible. Carcasses were marked with tags to prevent counting during subsequent surveys. All redds were flagged, measured, and assigned a unique record number. Redds were assigned a species if a fish was observed constructing, defending, or holding on a redd. Newly observed redds were aged as catergory one-new to the survey and redds recaptured on subsequent surveys received categorical ages ranging from age two as still visible and measurable, age three as still visible but not measurable, and age four redds were no longer visible (flag only).

### 2.4 Assigning Species to Unknown Redds

In order to assign species to unknown redds Ricker et al. (2014) applied a K-nearest neighbors (kNN) algorithm. The kNN uses distance in X-Y space and time to nearest 3 known species red or live fish to make a prediction of species based on majority vote. Each unknown redd was assigned to a species for furtheranalysis of species specific red abundance.

We used Leave-one-out-cross-validation (LOOCV) to evaluate the KNN predictions. The LOOCV process involved removing each known redd sequentially from the data, predicting the red from the remaining data, and comparing the prediction to truth. Errors in prediction are not however propogated into the total error of within reach abundance. All kNN and LOOCV analysis were executed in program R with the "class" package (Venables and Ripley 2002) and "caret" package (Kuhn 2013).

### 2.5 Estimation of Within-Reach Redd Abundance

Estimation of the number of redds in each sample reach is derived by dividing the total redd count by the square root of the seasonally pooled redd survival rate. The redd survival rate is estimated as fraction of re-observed age 2 and 3 flagged redds (still visible) to age 4 (no longer visible) (Section 2.3) (Ricker et al. 2014, Schwarz et al 1993).

### 2.6 Estimation of Total Redd Abundance Within the Sample Frame

Total redd abundance expansion to the frame is estimated with a Simple Random Sample estimator (Adams et al. 2011). See Ricker et al. 2014 for standard error equations and calculation methods and Adams et al. 2011 for correction factors. Bootstrap re-sampling was implemented to estimate between-reach variance and within-reach variance (Ricker et al. 2014).

## 3 RESULTS

### 3.1 Reach Survey Frequency

Field staff conducted spawning ground surveys in twenty two reaches, in five streams, within the Humboldt Bay watershed between November 7, 2015 and June 14, 2016. Visitation rates to stream reaches averaged 4.1 times and the average number of days between visits was twenty nine days (Table 1).

### 3.2 Fish Observations

Live fish observations include 3 Chinook salmon, 509 coho salmon, 15 steelhead trout, and 22 unidentified live salmonid adults. Table 2 and Table 3 summarize and Figure 2 displays counts of live fish and carcass observations by week in Humboldt Bay. Non target live fish included 10 cutthroat trout and 14 pacific lamprey. Carcass observations include 2 Chinook salmon, 69 coho salmon, 0 steelhead trout, and 40 unidentified species. A total of six pacific lamprey carcasses were observed.

Sex ratios of male to female live fish and carcass observations were $0: 1,1: 0.84$, and 1:0.4 for Chinook salmon, coho salmon, and steelhead trout respectively. Descriptive statistics for live fish and carcasses are presented in (Table 4). Out of 111 carcass recoveries 44 were whole and measured for fork length.

Table 1. Summary of spawning ground survey statistics of the mean number of days between reach surveys and the maximum number of days between surveys.

| Reach Code | Stream | Mean | Max | N |
| :---: | :---: | :---: | :---: | :---: |
| 1039 | Elk River | 34.66 | 59 | 4 |
| 1041 | Elk River | 30.66 | 46 | 4 |
| 1043 | Elk River | 27.00 | 30 | 3 |
| 1062 | Elk River | - | 30 | 2 |
| 1065 | Elk River | 24.00 | 36 | 3 |
| 1100 | Elk River | 51.00 | 60 | 3 |
| 1101 | Elk River | - | 17 | 2 |
| 1132 | Salmon Creek | - | 90 | 2 |
| 1133 | Salmon Creek | - | 25 | 2 |
| 1135 | Salmon Creek | - | 37 | 2 |
| 905 | Jacoby Creek | - | 26 | 2 |
| 907 | Jacoby Creek | - | 22 | 2 |
| 916 | Jacoby Creek | 9.16 | 16 | 7 |
| 945 | Ryan Creek | 49 | 60 | 3 |
| $* 979$ | Freshwater Creek | 26.28 | 47 | 8 |
| $* 980$ | Freshwater Creek | 21.6 | 42 | 11 |
| $* 981$ | Freshwater Creek | 16.11 | 45 | 10 |
| $* 982$ | Freshwater Creek | - | 33 | 2 |
| 990 | Freshwater Creek | 20.5 | 30 | 3 |
| 994 | Freshwater Creek | - | 23 | 2 |
| $* 1004$ | Freshwater Creek | 18.42 | 41 | 8 |
| 1009 | Freshwater Creek | 18.8 | 45 | 7 |
| $* 1014$ | Freshwater Creek | 17.8 | 44 | 6 |

*Denotes reaches in which surveys continued through late May or early June to capture the entire steelhead spawning run.

Table 2. Live fish observations by calendar week.

| Week Beginning | Chinook | coho | steelhead | unidentified | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $2015-11-16$ | 0 | 0 | 0 | 0 | 0 |
| $2015-11-23$ | 0 | 0 | 0 | 0 | 0 |
| $2015-11-30$ | 0 | 0 | 0 | 0 | 0 |
| $2015-12-07$ | 0 | 0 | 0 | 0 | 0 |
| $2015-12-14$ | 0 | 36 | 0 | 0 | 36 |
| $2015-12-21$ | 0 | 7 | 0 | 0 | 7 |
| $2015-12-28$ | 1 | 83 | 0 | 1 | 85 |
| $2016-01-04$ | 0 | 45 | 0 | 2 | 47 |
| $2016-01-11$ | 1 | 207 | 0 | 8 | 216 |
| $2016-01-18$ | 0 | 89 | 0 | 3 | 92 |
| $2016-01-25$ | 1 | 15 | 0 | 0 | 16 |
| $2016-02-01$ | 0 | 18 | 11 | 5 | 34 |
| $2016-02-08$ | 0 | 7 | 3 | 3 | 13 |
| $2016-02-15$ | 0 | 0 | 0 | 0 | 0 |
| $2016-02-22$ | 0 | 0 | 0 | 0 | 0 |
| $2016-02-29$ | 0 | 2 | 1 | 0 | 3 |
| $* 2016-03-07$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-03-14$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-03-21$ | 0 | 0 | 0 | 0 | 0 |
| $2016-03-28$ | 0 | 0 | 0 | 0 | 0 |
| $2016-04-04$ | 0 | 0 | 0 | 0 | 0 |
| $2016-04-11$ | 0 | 0 | 0 | 0 | 0 |
| $2016-04-18$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-04-25$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-05-02$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-05-09$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-05-16$ | 0 | 0 | 0 | 0 | 0 |
| $2016-05-23$ | 0 | 0 | 0 | 0 | 0 |
| $2016-05-30$ | 0 | 3 | 0 | 0 | 0 |
| Total | 3 | 0 | 0 | 0 | 0 |

[^1]Table 3. Carcass observations by calendar week.

| Week Beginning | Chinook | coho | steelhead | unidentified | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $2015-11-16$ | 0 | 0 | 0 | 0 | 0 |
| $2015-11-23$ | 0 | 0 | 0 | 0 | 0 |
| $2015-11-30$ | 0 | 0 | 0 | 0 | 0 |
| $2015-12-07$ | 0 | 0 | 0 | 0 | 0 |
| $2015-12-14$ | 0 | 0 | 0 | 0 | 0 |
| $2015-12-21$ | 0 | 0 | 0 | 0 | 0 |
| $2015-12-28$ | 0 | 1 | 0 | 1 | 2 |
| $2016-01-04$ | 1 | 1 | 0 | 0 | 2 |
| $2016-01-11$ | 0 | 12 | 0 | 1 | 13 |
| $2016-01-18$ | 0 | 10 | 0 | 5 | 15 |
| $2016-01-25$ | 0 | 4 | 0 | 2 | 6 |
| $2016-02-01$ | 0 | 22 | 0 | 10 | 32 |
| $2016-02-08$ | 1 | 11 | 0 | 15 | 29 |
| $2016-02-15$ | 0 | 7 | 0 | 2 | 7 |
| $2016-02-22$ | 0 | 0 | 0 | 3 | 3 |
| $2016-02-29$ | 0 | 1 | 0 | 1 | 2 |
| $* 2016-03-07$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-03-14$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-03-21$ | 0 | 0 | 0 | 0 | 0 |
| $2016-03-28$ | 0 | 0 | 0 | 0 | 0 |
| $2016-04-04$ | 0 | 0 | 0 | 0 | 0 |
| $2016-04-11$ | 0 | 0 | 0 | 0 | 0 |
| $2016-04-18$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-04-25$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-05-02$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-05-09$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-05-16$ | 0 | 0 | 0 | 0 | 0 |
| $2016-05-23$ | 0 | 0 | 0 | 0 | 0 |
| $2016-05-30$ | 0 | 0 | 0 | 0 | 0 |
| Total | 2 | 0 | 0 | 0 | 0 |

[^2]

Figure 2. Stacked Bar Plot of live fish (A), individual redd, and carcass (C) observations by week. Please note redds don't include those from extended surveys in March and April.

Table 4. Descriptive statistics of live fish observations (L) and carcasses (C).

| Species | Sex | Mean |  |  | Median |  |  | Min |  |  | Max |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | L | C | L | C | L | C | L | C | L | C | L | C |
| Chinook | F | 98 | 95 | 95 | 95 | 90 | 94 | 110 | 95 | 10.4 | 0.7 | 3 | 2 |
| Chinook | M | - | - | - | - | - | - | - | - | - | - | - | - |
| Chinook | Unk | - | - | - | - | - | - | - | - | - | - | - | - |
| coho | F | 59 | 61 | 60 | 60 | 40 | 50 | 70 | 70 | 5.3 | 4.8 | 192 | 18 |
| coho | M | 57 | 64 | 60 | 66 | 25 | 44 | 80 | 74 | 11.9 | 7.2 | 229 | 19 |
| coho | Unk | 51 | 61 | 55 | 61 | 25 | 52 | 70 | 69 | 14.1 | 12.2 | 17 | 2 |
| steelhead | F | 63 | 73 | 63 | 73 | 60 | 73 | 65 | 73 | 3.5 | N/A | 2 | 1 |
| steelhead | M | 66 | - | 70 | - | 55 | - | 75 | - | 8.2 | - | 5 | - |
| steelhead | Unk | 58 | - | 60 | - | 15 | - | 75 | - | 18.6 | - | 8 | - |
| Unknown | F | 60 | - | 60 | - | 55 | - | 65 | - | 5.0 | - | 3 | - |
| Unknown | M | 53 | - | 53 | - | 30 | - | 75 | - | 31.8 | - | 2 | - |
| Unknown | Unk | 48 | 67 | 50 | 67 | 20 | 65 | 65 | 68 | 12.9 | 2.12 | 16 | 2 |

### 3.3 Redd Observations

Out of 454 redds observed 171 redds were identified to species. The known species observations compared to the LOOCV predictions are displayed in Table 5. The KNN predictions, correctly predicted redds divided by the total known redds, were $94.1 \%$ accurate.

Table 5. Confusion matrix of the known species redds in columns crossed with LOOCV redd predictions in rows.

| Chinook | Chinook | coho | steelhead |
| :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 |
|  | 3 | 161 | 5 |
|  | 0 | 2 | 0 |

Newly constructed redd observations for coho and unidentified species peaked during the second week of January when 135 anadromous salmonid redds were counted (Table 6, Figure 2). Steelhead redd observations peaked during the first week in February when 5 redds were tallied. Reach 1039, in Elk river, had the highest count of redds (Table 7). Non-target redd totals were 14 cutthroat redds and 608 pacific lamprey redds (See Appendix, Table 12).

### 3.4 Total Redd Abundance

Total redd abundance estimates for Chinook salmon in Humboldt Bay tributaries in 2015 were not generated due to low sample sizes. Total redd abundance estimates, in all 31 reaches, for coho salmon and steelhead trout with 95\% confidence intervals was 617 $(430,849)$ and $96(21,185)$ respectively. Estimates and descriptive statistics of surveyed reaches are presented in Table 8.

Table 6. Counts of newly constructed known redds by calendar week.

| Week Beginning | Chinook | coho | steelhead | unidentified | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $2015-11-16$ | 0 | 0 | 0 | 1 | 1 |
| $2015-11-23$ | 0 | 0 | 0 | 0 | 0 |
| $2015-11-30$ | 0 | 0 | 0 | 1 | 1 |
| $2015-12-07$ | 0 | 0 | 0 | 0 | 0 |
| $2015-12-14$ | 0 | 14 | 0 | 8 | 22 |
| $2015-12-21$ | 0 | 2 | 0 | 2 | 4 |
| $2015-12-28$ | 1 | 30 | 0 | 49 | 80 |
| $2016-01-04$ | 0 | 19 | 0 | 27 | 46 |
| $2016-01-11$ | 1 | 69 | 0 | 66 | 135 |
| $2016-01-18$ | 0 | 15 | 0 | 10 | 25 |
| $2016-01-25$ | 1 | 2 | 0 | 3 | 6 |
| $2016-02-01$ | 0 | 10 | 5 | 37 | 52 |
| $2016-02-08$ | 0 | 1 | 1 | 25 | 27 |
| $2016-02-15$ | 0 | 0 | 0 | 10 | 10 |
| $2016-02-22$ | 0 | 0 | 0 | 10 | 10 |
| $2016-02-29$ | 0 | 1 | 0 | 20 | 21 |
| $* 2016-03-07$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-03-14$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-03-21$ | 0 | 0 | 0 | 0 | 0 |
| $2016-03-28$ | 0 | 0 | 0 | 0 | 0 |
| $2016-04-04$ | 0 | 0 | 0 | 3 | 3 |
| $2016-04-11$ | 0 | 0 | 0 | 4 | 4 |
| $2016-04-18$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-04-25$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-05-02$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-05-09$ | 0 | 0 | 0 | 0 | 0 |
| $* 2016-05-16$ | 0 | 0 | 0 | 0 | 0 |
| $2016-05-23$ | 0 | 0 | 0 | 3 | 3 |
| $2016-05-30$ | 0 | 0 | 0 | 3 | 3 |
| Total | 3 | 163 | 6 | 282 | 454 |

*No surveys conducted during this week.

Table 7. Counts of anadromous salmonid redds within each survey reach. Survey reach numbers are labled in figure 1.

| Location Code | Chinook | coho | steelhead | unidentified | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 905 | 0 | 6 | 0 | 6 | 12 |
| 907 | 0 | 1 | 0 | 2 | 3 |
| 916 | 0 | 18 | 0 | 14 | 32 |
| 945 | 0 | 1 | 0 | 14 | 15 |
| 979 | 0 | 1 | 0 | 20 | 21 |
| 980 | 0 | 7 | 0 | 44 | 51 |
| 981 | 3 | 9 | 1 | 22 | 36 |
| 990 | 0 | 1 | 0 | 4 | 5 |
| 994 | 0 | 7 | 0 | 10 | 17 |
| 1004 | 0 | 22 | 0 | 35 | 57 |
| 1006 | 0 | 3 | 0 | 9 | 12 |
| 1009 | 0 | 2 | 0 | 6 | 8 |
| 1014 | 0 | 6 | 0 | 9 | 15 |
| 1039 | 0 | 29 | 0 | 31 | 60 |
| 1041 | 0 | 12 | 0 | 13 | 25 |
| 1043 | 0 | 4 | 2 | 2 | 8 |
| 1062 | 0 | 0 | 1 | 0 | 1 |
| 1065 | 0 | 14 | 0 | 7 | 21 |
| 1100 | 0 | 14 | 2 | 23 | 39 |
| 1101 | 0 | 6 | 0 | 4 | 10 |
| 1133 | 0 | 0 | 0 | 6 | 6 |
| 1135 | 0 | 0 | 0 | 1 | 1 |
| Total | 3 | 163 | 6 | 282 | 454 |

Table 8. Estimated number of redds in 23 surveyed reaches in Humboldt Bay and in the Freshwater Creek Life Cycle Monitoring Station (LCS) where all 9 reaches were surveyed. Estimated Descriptive statistics are between sample reaches.

|  | Humboldt Bay |  | LCS |  |
| :---: | :---: | :---: | :---: | :---: |
|  | coho | steelhead | coho | steelhead |
| Redd Estimate | 587 | 59 | 323 | 1 |
| Standard Error | 6.24 | 1.92 | 12.65 | - |
| Variance | 858.1 | 77.0 | 1281.2 | - |
| $95 \%$ Confidence Interval | $(529,629)$ | $(42,77)$ | $(252,395)$ | - |

## 4 DISCUSSION

This year's reach survey frequency is less than in recent years (Table 9) due to heavy rainfall in an El Nino weather cycle. Table 9 below includes randomly selected reaches and additional reaches surveyed in the LCM, for a total of 23 reaches during each year.

Table 9. Reach survey frequency by year in Humboldt Bay tributaries.

| Year | Mean \# surveys per reach | Mean \# days between surveys |
| :---: | :---: | :---: |
| $2010-2011$ | 6.8 | 20 |
| $2011-2012$ | 7.2 | 20 |
| $2012-2013$ | 8.3 | 16 |
| $2013-2014$ | 3.8 | 26 |
| $2014-2015$ | 5.6 | 25 |
| $2015-2016$ | 4.1 | 29 |

The timing of the rainfall events in January (Figure 3) prevented surveys from occurring as frequently as the protocol suggests (every ten days). Normally surveys are delayed for $4-6$ days after a significant rainfall event. Minimum visibility conditions to conduct redd surveys is 50 cm which is usually reached as streams approach median winter base flows (Figure 3). If the storm frequency is once a week only a few reaches in the frame can be surveyed between storm events. Some reaches were visited only two times during the entire spawning season (Table 1). Infrequent reach visitation could decrease the chances of observations of both carcasses and new redds.

High water during the historical peak of the coho spawning season, in January, most likely obscured redds and decreased carcass and live fish observations. Redds constructed during this period may have been obscured and were unavailable for surveyors to observe during a subsequent survey (Jones 2012). Flood events could have moved carcasses into deep pools and high onto the flood plain impeding observations. As the water visibility decreases the chances of viewing live fish decreases. All of these factors could lead to a lower estimate of redds within the Humboldt Bay sample frame during an El Nino water year.


Figure 3. Discharge in Little River, used as a proxy for Humboldt Bay, December 2015 through February 2016.

The Coho redd estimate total for 2015 was lowest when compared to the last five years (Table 10) in the Humboldt Bay estimate. For the Freshwater Creek Life Cycle Monitoring Station coho had the third highest redd estimate when compared to the last five years (Table 11). The high estimate within the Freshwater Creek LCM compared to the Humboldt Bay estimate may be due to the survey frequency in the LCM (Table 1). More frequent surveys in the LCM are required to estimate the adult coho population by recapturing marked carcasses.

Estimates for low density species such as Chinook salmon, in both Humboldt Bay and the Freshwater Creek LCM (2011 to 2015), and Steelhead, in the LCM (2010 to 2013), continue to challenge the protocol and the ability to analyze the data with the tools available in Adams et al. (2011). Steelhead estimates may also be influenced by the lack of surveys later in the season which could capture the spawning activity.

Table 10. Redd estimates with $95 \%$ confidence intervals for three species during six years in Humboldt Bay.

| Year | Chinook | coho | steelhead |
| :---: | :---: | :---: | :---: |
| $2010-2011$ | $19(2,37)$ | $1099(719,1478)$ | $11(3,19)$ |
| $2011-2012$ | 0 | $1738(1014,2463)$ | $19(4,33)$ |
| $2012-2013$ | 0 | $763(511,1016)$ | $172(73,272)$ |
| $2013-2014$ | 0 | $630(123,1138)$ | $35(10,61)$ |
| $2014-2015$ | 0 | $1632(1171,2094)$ | $226(127,326)$ |
| $2015-2016$ | $3(3,3)$ | $617(430,849)$ | $97(21,184)$ |

Table 11. Redd estimates with $95 \%$ confidence intervals for three species during six years in Freshwater Creek Life Cycle Monitoring Station.

| Year | Chinook | coho | steelhead |
| :---: | :---: | :---: | :---: |
| $2010-2011$ | $12(12,12)$ | $231(223,239)$ | $4(4,4)$ |
| $2011-2012$ | 0 | $420(391,449)$ | $7(7,7)$ |
| $2012-2013$ | 0 | $244(240,249)$ | $13(13,13)$ |
| $2013-2014$ | 0 | $127(87,168)$ | $2(2,2)$ |
| $2014-2015$ | 0 | $453(367,538)$ | $72(57,88)$ |
| $2015-2016$ | $3(3,3)$ | $322(251,394)$ | $1(1,1)$ |

## 5 REFERENCES

Adams, P., L. B. Boydstun, S. P. Gallagher, M. K. Lacy, T. McDonald, and K. E. Shaffer. 2011. California Coastal Salmonid Population Monitoring: Strategy, Design, and Methods. Fish Bulletin No. 180. California Department of Fish and Wildlife.

California Department of Fish and Game. 2002. Status review of California coho salmon north of San Francisco. Report to the California Fish and Game Commision, April 2002.

California Department of Fish and Game. 2004. Recovery strategy for California coho salmon. Report to the California Fish and Game Commission. 594 pp.

Gallagher, S. P., P. K. Hahn, and D. H. Johnson. 2007. Redd counts. Pages 197-234 in D. H. Johnson, B. M. Shrier, J. S. O'Neal, J. A. Knutzen, X. Augerot, T. A. O'Neil, and T. N. Pearsons, editors. Salmonid field protocols handbook: techniques for assessing status and trends in salmon and trout populations. American Fisheries Society, Bethesda, Maryland.

Garwood, J., and S. Ricker. 2011. Spawner survey frame developmentfor monitoring adult salmonid populations in California. California Department of Fish and Game. Annual Report. 50 Ericson Ct., Arcata, CA. 18p.

Jones, E.C. 2012. Environmental stochasticity and the reliability of redd count data: a simulation study of redd construction, redd loss, and survey frequency in a small coastal stream in northern California. Master's thesis. Humboldt State University, Arcata, California.

Kuhn, M. Contributions from Jed Wing, Steve Weston, Andre Williams, Chris Keefer, Allan Engelhardt and Tony Cooper. 2013. caret: Classication and Regression Training. R package version 5.17-7. http://CRAN.R-project.org/package=caret.

McDonald, T., 2003. GRTS for the average Joe: A GRTS sampler for Windows. WEST Inc., 2003 Central Ave., Cheyenne, WY 82001. 9 pp.

McElhany, P., M.H. Ruckelshaus, M.J. Ford, T.C. Wainwright, and E.P. Bjorkstedt. 2000. Vaible salmonid populations and the recovery of evolutionarily significant units. U.S. Dept. Commerce, NOAA Tech. Memo. NMFS-NWFSC-42, 156p.

NMFS (National Marine Fisheries Service). 1997. Endangered and threatened species: threatened status for southern Oregon/northern California coast evolutionarily significant unit (ESU) of coho salmon. Federal Register 62: 24588-24609.

NMFS (National Marine Fisheries Service). 1999. Endangered and Threatened Species: Threatened Status for two Chinook salmon Evolutionarily Significant Units (ESUs) in California; final rule. 50 CFR Part 223. Federal Register 64: 50394-50415.

NMFS (National Marine Fisheries Service). 2005. Endangered and Threatened Species: final listing determinations for 16 ESUs of west coast salmon, and Final 4(d) protective regulations for threatened salmonid ESUs. Federal Register 70: 3716037204.

NMFS (National Marine Fisheries Service). 2006. Endangered and Threatened Species: final listing determinations for 10 distinct population segments of west coast steelhead; Final Rule. Federal Register 71: 831-862.

Ricker, S.J., K. Lindke, and C. Anderson. 2014. Results of regional spawning ground surveys and estimates of total salmonid redd construction in Humboldt Bay, Humboldt County California, 2013. California Department of Fish and Game, Anadromous Fisheries Resource Assessment and Monitoring Program, 50 Ericson Ct., Arcata, CA 95521.

Ricker, S.J., Ferreira, S. P. Gallagher, D.McCanne, and S.A. Hayes. 2014. Methods for Classifying Anadromous Salmonid Redds to Species. Coastal Salmonid Population Monitoring Technical Team Report. California Department of Fish and Wildlife, Arcata, California. 24p.

Schwarz, C.J. et al. 1993. Estimating salmon spawning escapement using capturerecapture methods . Canadian Journal of Fisheries and Aquatic Sciences 50 (6), 1181-1197, 72.

Venables, W. N., Ripley, B. D. 2002. Modern Applied Statistics with S. Fourth Edition. Springer, New York. 495 p. ISBN 0-387-95457-0.

## 6 APPENDIX

Table 12. Counts of newly constructed redds by calendar week for non-target species.

| Week Beginning | cutthroat | Pacific Lamprey |
| :---: | :---: | :---: |
| $2015-11-16$ | 0 | 0 |
| $2015-11-23$ | 0 | 0 |
| $2015-11-30$ | 0 | 0 |
| $2015-12-07$ | 0 | 0 |
| $2015-12-14$ | 0 | 0 |
| $2015-12-21$ | 0 | 0 |
| $2015-12-28$ | 0 | 0 |
| $2016-01-04$ | 1 | 0 |
| $2016-01-11$ | 0 | 0 |
| $2016-01-18$ | 0 | 0 |
| $2016-01-25$ | 0 | 0 |
| $2016-02-01$ | 0 | 0 |
| $2016-02-08$ | 10 | 0 |
| $2016-02-15$ | 2 | 1 |
| $2016-02-22$ | 0 | 0 |
| $2016-02-29$ | 0 | 0 |
| $* 2016-03-07$ | 0 | 0 |
| $* 2016-03-14$ | 0 | 0 |
| $* 2016-03-21$ | 0 | 0 |
| $2016-03-28$ | 0 | 3 |
| $2016-04-04$ | 0 | 52 |
| $2016-04-11$ | 0 | 130 |
| $2016-04-18$ | 0 | 165 |
| $* 2016-04-25$ | 0 | 0 |
| $* 2016-05-02$ | 0 | 0 |
| $* 2016-05-09$ | 0 | 0 |
| $* 2016-05-16$ | 0 | 0 |
| $2016-05-23$ | 0 | 91 |
| $2016-05-30$ | 1 | 166 |
| Total | 14 | 608 |

*No surveys conducted during this week.


[^0]:    ${ }^{1}$ Primary author to whom correspondence should be addressed. Humboldt State University Sponsored Programs. ${ }^{2}$ Humboldt State University.

[^1]:    *No surveys conducted during this week.

[^2]:    *No surveys conducted during this week.

