

# **The 2015 Refugio Beach Oil Spill: Assessment of Surfperch (Embiotocidae) Exposure**

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## INTRODUCTION

On May 19, 2015, an underground pipeline (Line 901), owned and operated by Plains All American Pipeline, L.P., sustained a release of crude oil near Refugio State Beach in Santa Barbara County, California. Oil released from the pipeline pooled, then overflowed into a nearby culvert, across land and other drainage systems, and entered the Pacific Ocean in the surf zone. The surf zone in this area supports relatively large populations of fish, such as silversides, surfperches, croakers, flatfishes and rays (Allen and Pondella, 2006). These fish would have been exposed by direct contact with floating or submerged oil, uptake from oil dissolved or suspended in the water column and the food chain. Additionally, the spill occurred during the spawning season of several surf zone fish species, such that sensitive early life stages may have been exposed to oil. For example, the barred surfperch (*Amphistichus argenteus*) and walleye surfperch (*Hyperprosopon argenteum*) give birth to live young from March to July in this area (Carlisle et al., 1960; California Department of Fish and Game, 2001).

Crude oil contains hundreds to thousands of chemicals that are potentially toxic to fish. Exposure to one class of chemicals, polycyclic aromatic hydrocarbons (PAHs), has been associated with developmental abnormalities, immunosuppression, hepatic lesions and altered growth in fish (Myers et al. 1994; Heintz et al. 2000; Arkoosh et al. 2001; Meador et al. 2006; Reynaud and Deschaux 2006; Incardona et al. 2004; Incardona et al. 2012). Fish rapidly take up PAHs present in their food and the environment and quickly metabolize these compounds to more polar compounds. The more polar PAH metabolites are then secreted into fluids such as bile and urine for elimination via the gastroenteric tract or kidneys (Roubal et al., 1977; Varanasi et al., 1989; Krahn et al., 1984). Therefore, assessment of bile for PAH metabolites provides information on recent uptake and exposure to these compounds. Elevated biliary PAH metabolites have been measured in fish following oil spills (Krahn et al., 1986; Sol et al., 2000; Murawski et al, 2014; Snyder et al., 2015). Additionally, fish living near the Coal Oil Point oil seeps in Santa Barbara have been shown to have elevated levels of PAH metabolites in bile, compared to nearby reference locations (Spies et al., 1996; Roy et al., 2003).

The primary objective of this assessment was to obtain a quantitative estimate of PAH exposure in fish by measuring bile and muscle tissue concentrations in an indicator fish, as well as concentrations in the water. Surfperches (Embiotocidae) were selected because they are relatively resident and occupy the surf zone and shallow subtidal areas where significant oiling occurred (Carlisle et al, 1960). Barred surfperch and walleye surfperch were the two species evaluated. Exposure at Refugio State Beach, a heavily oiled area, was compared to a lesser oiled area, Gaviota State Beach, and an area near the Coal Oil Point oil seep, Campus Point, using samples from all three sites collected at both four days and approximately one year after the oil spill.

## METHODS

### Field Sampling Procedures

Sampling locations were selected based on shoreline oiling observations on 22 May 2015 (Figure 1). Gaviota State Beach was not reported as being oiled at that time. Refugio State Beach was adjacent to the spill location and was heavily oiled. Campus



**Figure 1.** Sampling locations for surfperch bile on 23 May 2015 and 18 May 2016.

Point was not reported as oiled but may have had some oil exposure from Line 901 oil and adjacent Coal Oil Point seeps. On 23 May 2015, surfperches were caught by hook and line at Refugio State Beach due to safety limitations of entering oil contaminated water. A beach seine was used at Campus Point and hook and line was used at Gaviota State Beach due to wind and surf conditions. On 18 May 2016, a beach seine was used at the same three locations to collect surfperch. Fish were identified to species and maintained alive until sample processing within one to two hours. Total length was measured and the fish were killed by cervical dislocation. The gall bladder was immediately extracted and bile was collected in 4 milliliter Sun-Sri™ amber vials and stored on ice in the field. Bile samples were then frozen at -20°C until analysis at the Northwest Fisheries Science Center, Seattle, WA. The sex was determined by examining gonadal tissue. The remaining carcass was wrapped in foil, placed in a zip-top bag, stored on ice in the field, and then frozen at -20°C until analyses were conducted at the California Department of Fish and Wildlife, Water Pollution Control Laboratory (CDFW-WPCL), Gold River, CA. As part of the fisheries closure assessment (OEHHA, 2015), barred surfperch were collected by hook and line at Gaviota State Beach and Refugio State Beach on 10 June 2015. No fish were collected from Campus Point. Fish were wrapped in foil, placed in a zip-top bag, stored on ice in the field, and then frozen at -20°C until analyses were conducted at the CDFW-WPCL.

Triplicate surf water samples were collected on 27 May 2015 and 17 May 2016 at Gaviota State Beach, Refugio State Beach and Campus Point. Samples were collected in one-liter amber glass bottles by submerging the bottle in the surf zone until filled with minimal headspace. Samples were immediately placed on ice and transported to the CDFW-WPCL for analysis.

## **Bile Analysis**

Bile samples were analyzed using a high-performance liquid chromatography fluorescence (HPLC-F) method described in Krahn et al. (1984). This method results in the determination of the concentrations of classes of PAH metabolites fluorescing in the regions typified by naphthalene (NPH), phenanthrene (PHN) and benzo[a]pyrene (BaP). Bile was injected directly onto a Waters HPLC-F system equipped with a C-18 reverse-phase column (Phenomenex Synergi Hydro). The PAH metabolites were eluted with a linear gradient from 100% water (containing a trace amount of acetic acid) to 100% methanol at a flow of 1.0 mL/min. Chromatograms were recorded at the following wavelength pairs: 1) 292/335 nm where many 2-3 benzene ring aromatic compounds (e.g., NPH) fluoresce, 2) 260/380 nm where several 3-4 ring compounds (e.g., PHN) fluoresce and 3) 380/430 nm where 4-5 ring compounds (e.g., BaP) fluoresce. Peaks eluting after 9 minutes were integrated and the areas of these peaks were summed. The concentrations of fluorescent PAHs in the bile samples of the fish were determined using NPH, PHN or BaP as external standards and converting the fluorescence response of bile to PHN (ng PHN equivalents/g bile), NPH (ng NPH equivalents/g bile) or BaP (ng BaP equivalents/g bile) equivalents on a wet weight basis. In addition, protein analysis as described in da Silva et al. (2006) was completed for all bile samples as previous laboratory contaminant exposure studies on fish have shown that normalization of biliary PAH metabolite concentrations to protein values may help account for variation in metabolite levels based on feeding status (Collier and Varanasi 1991).

To ensure that the HPLC-F system was operating properly, a NPH/PHN/BaP calibration standard was analyzed numerous times ( $n \geq 5$ ) until a relative standard deviation  $< 15\%$  was obtained for each PAH. As part of the laboratory quality assurance plan (Sloan et al. 2006), a method blank and a fish bile positive control sample (bile of Atlantic salmon exposed to 25 mg/L of Monterey crude oil for 48 hours) were analyzed with each batch of fish bile samples. All sample batches met the laboratory quality assurance criteria.

## **Fish Tissue and Water Analysis**

For the 23 May 2015 collection, skinless filets of individual barred surfperch were composited into one sample for each site: Gaviota ( $n=6$  fish), Refugio ( $n=9$  fish), and Campus Point ( $n=6$  fish). For the 10 June 2015 collection, skinless filets of individual barred surfperch were again composited into one sample for each site: one sample for Gaviota ( $n=9$  fish), and one for Refugio ( $n=4$  fish). Tissues were extracted by pressurized fluid extraction, followed by gel permeation chromatography and silica clean-up. Water samples and tissue extracts were analyzed for PAHs by GC/MS-SIM (USEPA Method 8270 mod). Results for these 45 individual PAHs and alkylated homologue groups were summed to estimate total PAHs (TPAH<sub>45</sub>): naphthalene; C1-naphthalenes; C2-naphthalenes; C3-naphthalenes; C4-naphthalenes; acenaphthylene; acenaphthene; fluorene; C1-fluorenes; C2-fluorenes; C3-fluorenes; phenanthrene; anthracene; C1-phenanthrene/anthracene; C2-phenanthrene/anthracene; C3-phenanthrene/anthracene; C4-phenanthrene/anthracene; pyrene; fluoranthene; C1-fluoranthene/pyrenes; C2-fluoranthene/pyrenes; C3-fluoranthene/pyrenes; C4-fluoranthene/pyrenes; benz[a]anthracene; chrysene; C1-chrysenes; C2-chrysenes; C3-

chrysenes; C4-chrysenes; benzo(a)pyrene; perylene; benzo(e)pyrene; benzo(b)fluoranthene; benzo(k)fluoranthene; benzo(g,h,i)perylene; indeno(1,2,3-c,d)pyrene; dibenz(a,h)anthracene; C1-dibenz(a,h)anthracene; C2-dibenz(a,h)anthracene; C3-dibenz(a,h)anthracene; dibenzothiophene; C1-dibenzothiophenes; C2-dibenzothiophenes; C3-dibenzothiophenes and biphenyl. When calculating TPAH<sub>45</sub>, non-detects were assumed to be zero. Tissue results are reported on a dry weight basis.

## RESULTS

### Field Observations

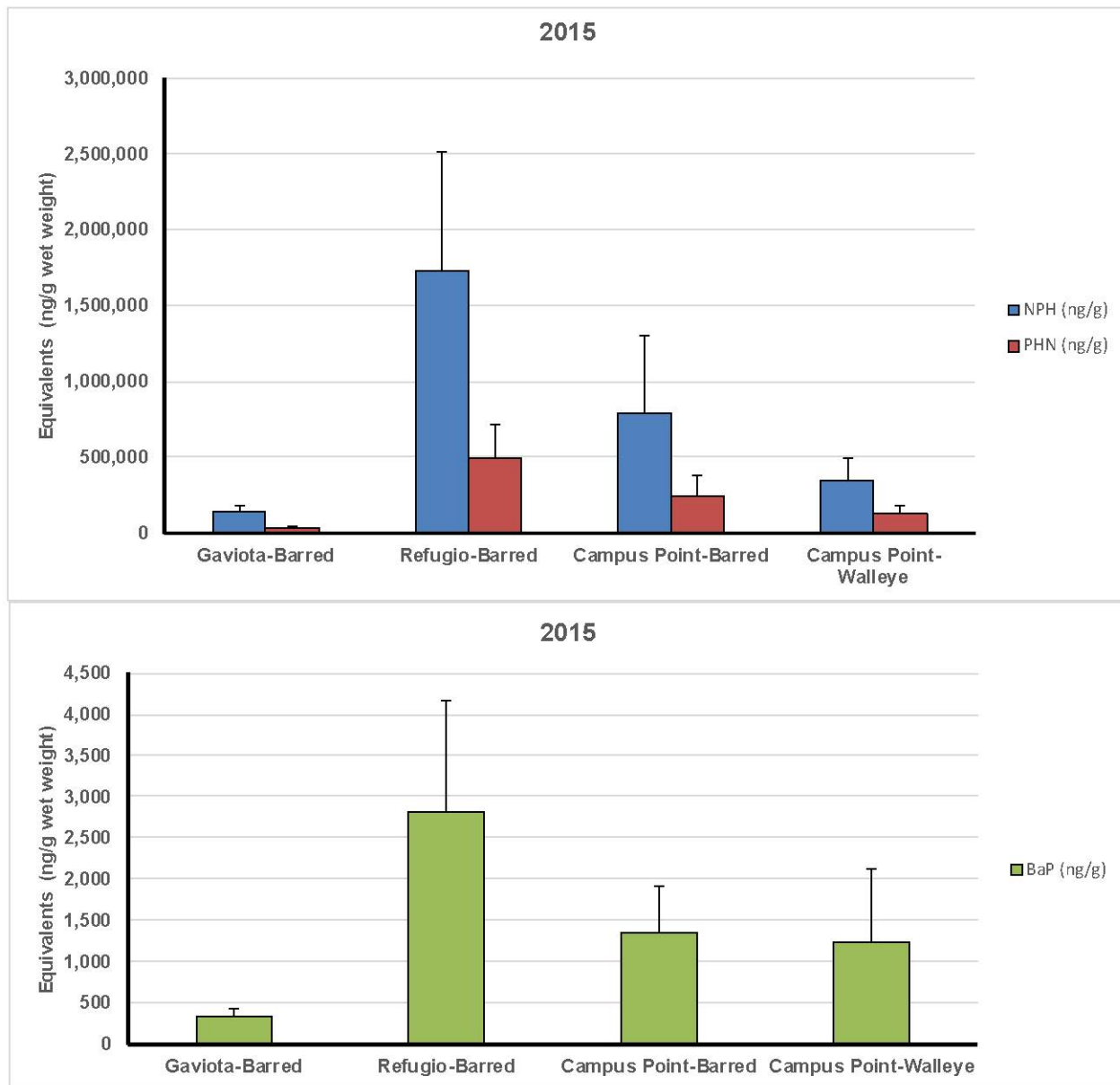
Total lengths of surfperch caught on 23 May 2015 ranged from 153 to 297 mm at Gaviota, 130 to 230 mm at Refugio and 142 to 205 mm at Campus Point. At each location, one female barred surfperch was observed to contain live young upon dissection. For the 18 May 2016 sampling, total lengths of surfperch ranged from 115-190 mm at Gaviota, 145-225 mm at Refugio and 145-195 mm at Campus Point. Two female with live young were observed at Campus Point and Refugio. In 2016, other species caught in the surf zone via beach seine included: shiner surfperch (*Cymatogaster aggregate*), kelp surfperch (*Brachyistius frenatus*), corbina (*Menticirrhus undulates*), topsmelt (*Atherinops affinis*), Pacific sardine (*Sardinops sagax*), jacksmelt (*Atherinopsis californiensis*), black perch (*Embiotoca jacksoni*), sargo (*Anisotremus davidsoni*), opaleye (*Girella nigricans*), white croaker (*Genyonemus lineatus*); giant kelpfish (*Heterostichus rostratus*), and diamond turbot (*Hypsopsetta guttulata*).

### Biliary PAH Metabolite Concentrations

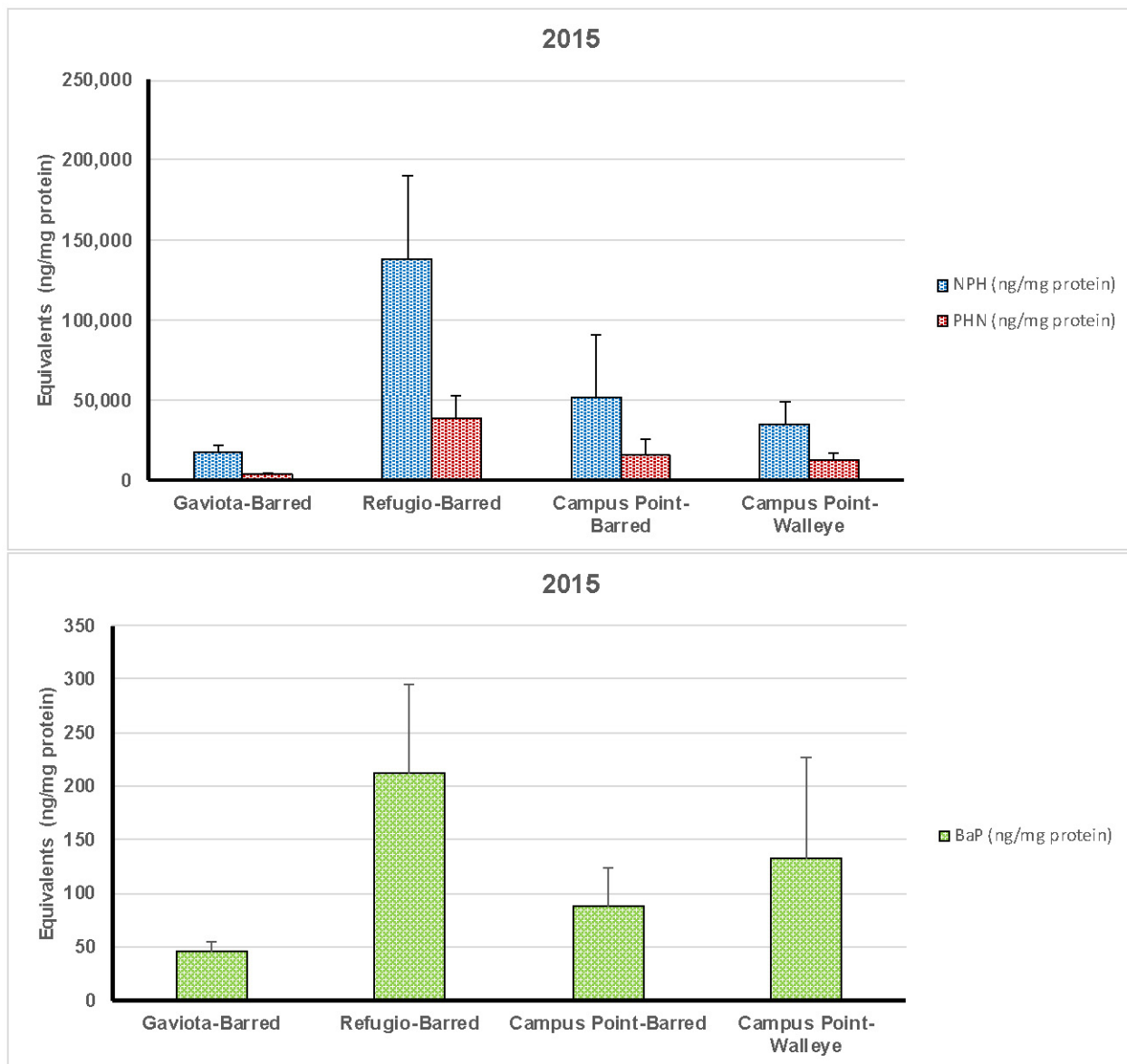
Concentrations of biliary PAH metabolites (based on wet weight or biliary protein) are provided in Appendices 1 and 2. The mean ( $\pm$  SD) biliary NPH, PHN and BaP equivalent concentrations measured in surfperch collected at Gaviota State Beach (n=6 for barred), Refugio State Beach (n=21 for barred) and Campus Point (n=5 for barred; n=9 for walleye) in 2015 are shown in Figure 2 (ng/g bile, wet weight) and Figure 3 (ng/mg protein). Significant differences (ANOVA  $p < 0.05$ ; Tukey-Kramer HSD test) in mean NPH, PHN and BaP equivalent concentrations (based on wet weight or biliary protein) were found among collection sites. For each PAH metabolite, barred surfperch from Refugio State Beach, adjacent to the oil release site, had a significantly higher mean level than those determined in fish from Campus Point or Gaviota. Mean PAH metabolite concentrations measured in bile of barred surfperch from Campus Point, adjacent to offshore oil seeps, were significantly higher than the same metabolites measured in barred surfperch from Gaviota, a lesser oiled site. At the Campus Point site, mean levels of biliary NPH and PHN equivalents (based on wet weight only) were significantly higher (ANOVA  $p < 0.05$ ; t-test) in barred surfperch than in those measured in walleye surfperch. Mean concentrations of NPH and PHN were higher than BaP equivalents.

Mean ( $\pm$  SD) NPH, PHN and BaP equivalent concentrations based on wet weight (Figure 4) or biliary protein (Figure 5) for each species collected at Gaviota (n=7 for barred; n=1 for walleye), Refugio State Beach (n=8 for barred; n=20 for walleye) and

Campus Point (n=1 for barred; n=13 for walleye) were not significantly different (ANOVA  $p > 0.05$ ;  $\log_{10}$  transformed data) in fish collected one year after the spill except PHN equivalent levels (wet weight only) in barred surfperch from the Refugio Beach site and Gaviota (  $p = 0.0487$ ).

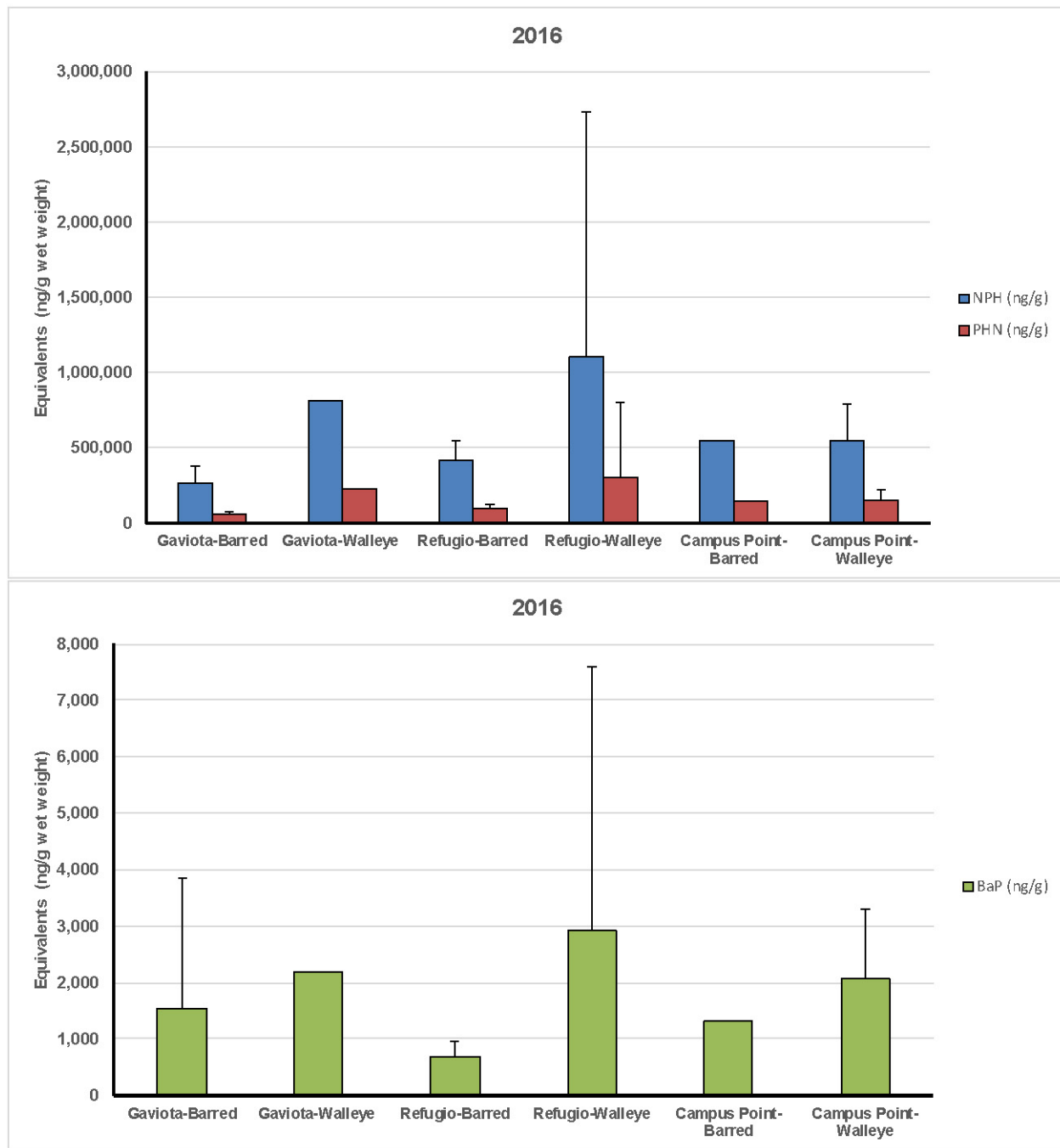


**Figure 2.** Mean ( $\pm$ SD) concentrations of bile equivalents (ng/g bile wet weight) measured in barred and walleye surfperch collected in 2015: naphthalene (NPH), and phenanthrene (PHN; Top) and benzo[a]pyrene (BaP; Bottom).

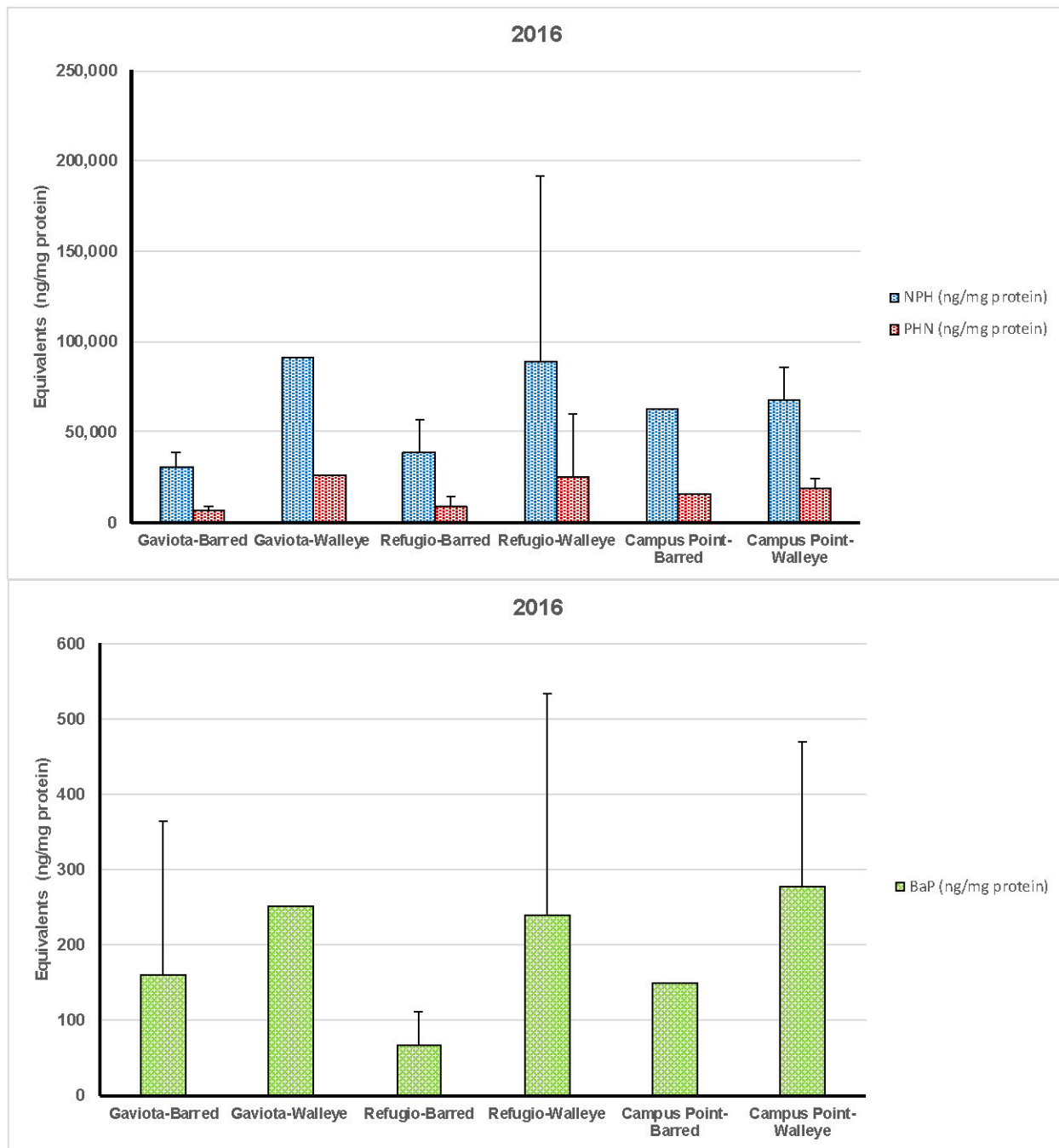


**Figure 3.** Mean ( $\pm$ SD) concentrations of bile equivalents (ng/mg bile protein) measured in barred and walleye surfperch collected in 2015: naphthalene (NPH) and phenanthrene (PHN; Top) and benzo[a]pyrene (BaP; Bottom).





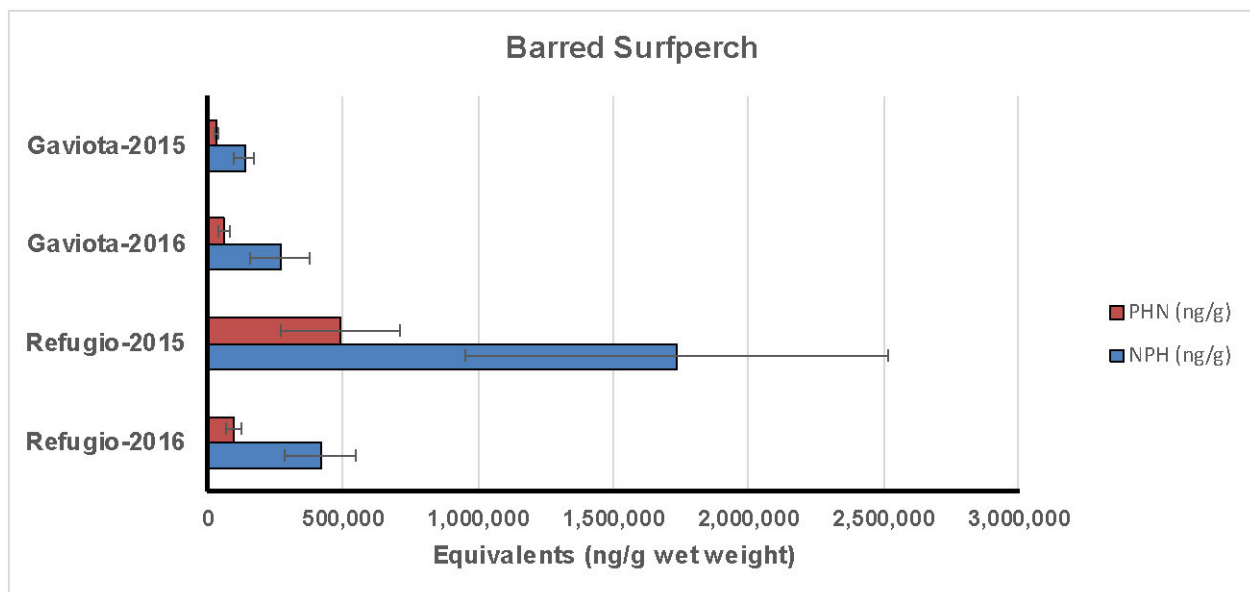
**Figure 4.** Mean ( $\pm$ SD) concentrations of bile equivalents (ng/g bile wet weight) measured in barred and walleye surfperch collected in 2016: naphthalene (NPH) and phenanthrene (PHN; Top); and benzo[a]pyrene (BaP; Bottom).



**Figure 5.** Mean ( $\pm$ SD) concentrations of bile equivalents (ng/mg bile protein) measured in barred and walleye surfperch collected in 2016: naphthalene (NPH) and phenanthrene (PHN; Top); and benzo[a]pyrene (BaP; Bottom).

At the Refugio State Beach site, mean concentrations of all PAH metabolites based on wet weight (Figure 4) in the 2016 collected fish were significantly higher in walleye surfperch compared to barred surfperch (ANOVA  $p < 0.05$ ; t-test;  $\log_{10}$  transformed data). Similarly, mean levels of protein-corrected PAH metabolites were significantly higher (ANOVA  $p < 0.05$ ; t-test) in walleye compared to barred surfperch except protein-corrected NPH ( $p = 0.0505$ ). Comparisons between species at the other two collection sites were not conducted due to inadequate numbers of bile samples.

Differences in mean concentrations of PAH metabolites based on sampling year for each species collected at the same site were examined. Barred surfperch collected at Refugio Beach in 2015 had significantly higher (ANOVA  $p < 0.0001$ ; t-test;  $\log_{10}$  transformed data) mean NPH, PHN and BaP concentrations (wet weight and protein-corrected) than those determined in the 2016 (Figure 6). In contrast, Gaviota barred surfperch collected in 2016 (Figure 6) had significantly higher mean concentrations (ANOVA  $p < 0.05$ ; t-test;  $\log_{10}$  transformed data) of NPH equivalents (wet weight and protein-corrected) and PHN equivalents (wet weight only) than the mean values of the 2015 fish. Walleye surfperch collected from Campus Point in 2016 had significantly higher mean concentrations of protein-corrected NPH, PHN and BaP equivalents, as well as NPH equivalents (wet weight only), compared to the 2015 fish. No other significant differences (ANOVA  $p > 0.05$ ) in mean concentrations of PAH metabolites were found for walleye surfperch from this site.

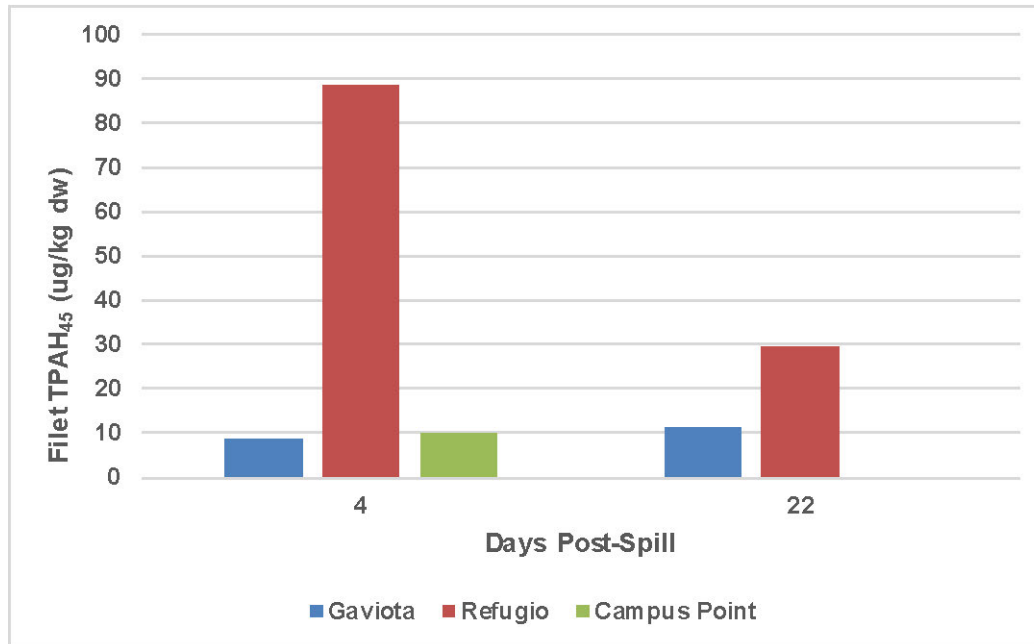


**Figure 6.** Mean ( $\pm$ SD) concentrations of bile naphthalene (NPH) and phenanthrene (PHN) equivalents (ng/g bile wet weight) measured in barred surfperch collected in 2015 and 2016.

### Fish Muscle PAH Concentrations

The TPAH<sub>45</sub> concentration in skinless filets collected four days after the spill followed the pattern seen in bile, with highest concentrations observed at Refugio (88 ug/kg dw; Figure 7). Naphthalenes (parent and C1-C4 alkylated) were the primary PAHs detected at Gaviota (100%), Refugio (91%) and Campus Point (78%). Tricyclic PAHs were also

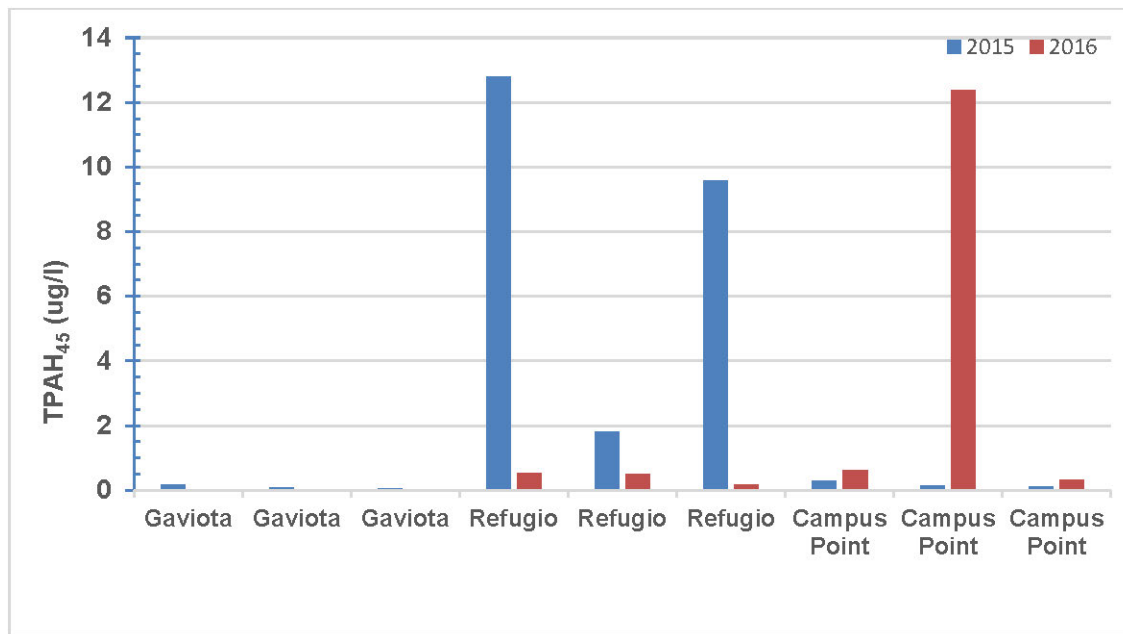
detected at Refugio (e.g., acenaphthene, dibenzothiophenes, and C1-phenanthrene/anthracene) and Campus Point (C1-phenanthrene/anthracene). Concentrations were almost three times lower at Refugio (30 ug/kg dw) 22 days post spill but were similar at Gaviota, consisting only of naphthalenes at both locations.



**Figure 7.** TPAH<sub>45</sub> concentrations (ug/kg dw) in a composite sample of barred surfperch skinless filets 4 and 22 days after the spill from Gaviota (4; n=6; 22; n=9 fish) and Refugio (4; n=9; 22; n= 4 fish) and 4 days after the spill from Campus Point (4; n=6 fish).

### Surf Water PAH Concentrations

TPAH<sub>45</sub> concentrations in surf water 8 days after the spill were highest at Refugio (1.8 – 12.8 µg/l) when compared to Gaviota (0.06 – 0.18 µg/l) and Campus Point (0.12 – 0.30 µg/l; Figure 8). This is consistent with the 2015 spatial pattern observed in fish bile and muscle tissue. One year after the spill, TPAH<sub>45</sub> concentrations were lower at Refugio (0.16 – 0.53 µg/l) and Gaviota (0.0 – 0.04 µg/l), but variable at Campus Point (0.3 – 12.4 µg/l; Figure 8). Based on fingerprinting analysis (Stout, 2016), it was determined that the maximum concentrations at Refugio in 2015 and Campus Point in 2016 contained crude oil micro-droplets or emulsions, due to the presence of minimally soluble 4- to 6-ring PAHs. Further PAH composition analysis revealed that the Refugio 2015 sample was a probable match to the Line 901 oil, due to the high proportion of dibenzothiophenes, but the Campus Point 2016 PAH distribution was consistent with seep oil. The PAH composition in the maximum Campus Point 2015 sample was also consistent with seep oil.



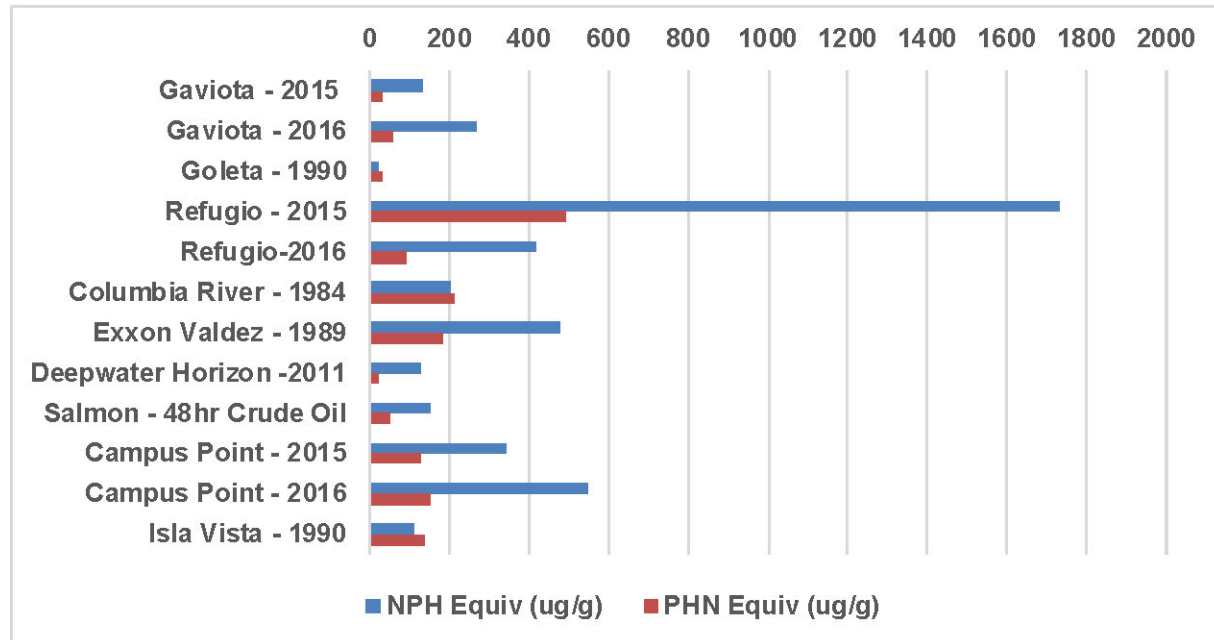
**Figure 8.** TPAH<sub>45</sub> concentrations ( $\mu\text{g/l}$ ) in triplicate surf water samples from Gaviota, Refugio and Campus Point collected 8 days after the spill (27 May 2015) and approximately one year after the spill (17 May 2016).

## DISCUSSION

Four days after the Line 901 oil release, surfperch biliary PAH metabolite concentrations were significantly higher at Refugio State Beach, compared to Campus Point and Gaviota State Beach. TPAH<sub>45</sub> concentrations in surfperch muscle and surf water reflected a similar spatial pattern in 2015. These results indicated surf zone fish exposures to PAHs were higher in the area adjacent to the oil release. One year after the oil spill, mean biliary PAH metabolite concentrations at Refugio declined, such that there was no longer a significant difference between the three sampling locations. Campus Point surfperch, continued to show elevated biliary PAH metabolite levels, compared to Gaviota State Beach, likely due to the presence of nearby natural oil seeps, consistent with elevated TPAH<sub>45</sub> levels in surf water at this location in 2016. Within site species differences between barred and walleye surfperch bile metabolite levels may have resulted from differences in food and habitat preferences (Carlisle et al, 1960; Feder et al, 1974; Hobson and Chess, 1986) but additional research would have to be conducted to further assess interspecies differences.

PAH metabolites in 2015 and 2016 bile samples were predominately naphthalene and phenanthrene derived metabolites, consistent with exposure to fresh crude oils, rather than higher molecular weight PAHs (e.g., BaP) that are associated with pyrogenic sources (Lee and Anderson, 2005). Exposure of fish to oil seep sediment has also resulted in bile PAH metabolites being dominated by NPH and PHN equivalents (Roy et al., 2003). However, levels measured in this study were somewhat higher than previously measured near Santa Barbara oil seeps. Spies et al (1996) sampled rainbow surfperch near the Isla Vista seeps at Coal Oil Point, at depths of 8-15m, and the Goleta

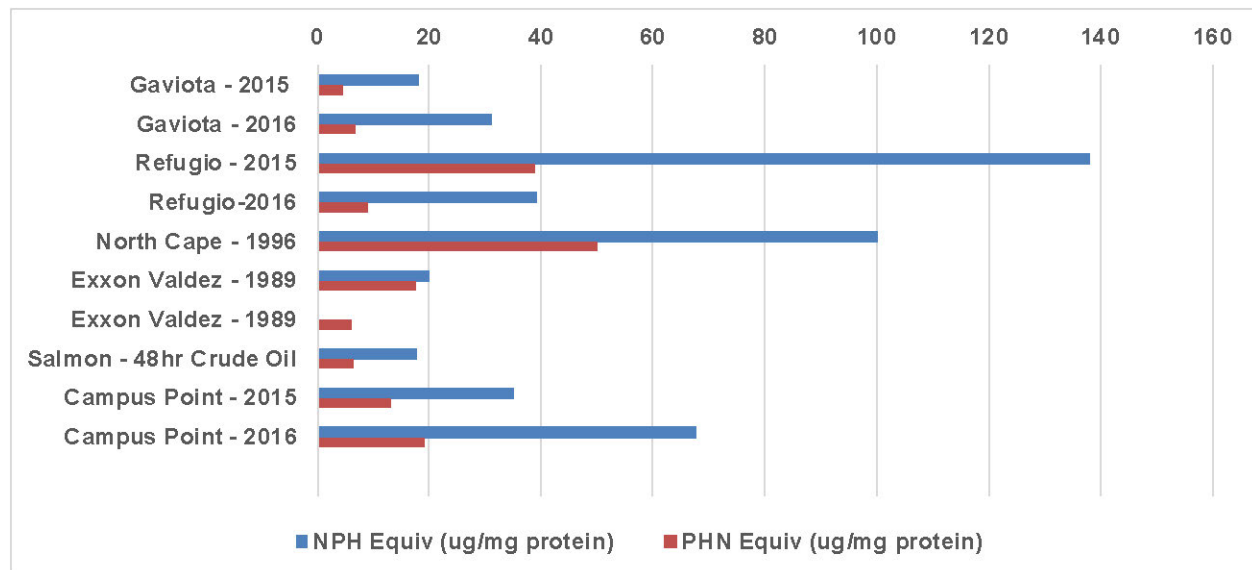
Pier in September 1990. Bile NPH and PHN equivalents were elevated near the seep site, compared to Goleta Pier (Figure 9). Liver cytochrome P-450 enzyme levels and mean gill and liver lesion scores were significantly higher in fish collected from the seep area, compared to Goleta.



**Figure 9.** Comparison of biliary naphthalene (NPH) and phenanthrene (PHN) equivalents (ug/g wet weight) mean concentrations from this study (2015, 2016 and salmon-48hr crude oil standard; barred surfperch at Gaviota and Refugio and walleye surfperch at Campus Point), previous studies in Santa Barbara (Goleta and Isla Vista, 1990; Spies et al., 1996) and following other oil spills (Krahn et al, 1986; Hom et al., 2008; Snyder et al., 2015).

NPH equivalent mean concentrations measured in barred surfperch bile at Refugio State Beach in 2015 were higher than measured following other oil spills (Figures 9 and 10). In 1984, a tanker released more than 170,000 gallons of residual fuel oil into the Columbia River (Krahn et al., 1986). White sturgeon collected 5 days later and 57 miles downriver from the spill had significantly elevated mean concentrations of NPH (200 ug/g) and PHN (210 ug/g) equivalents, compared to the upriver reference site (32 and 9.7 ug/g respectively). Fish downriver of the spill were observed to have oil in their mouths and showed physical signs of stress (e.g., excess mucus secretion; Kennedy and Baca, 1984). In March 1989, 11 million gallons of Prudhoe Bay crude oil from the Exxon Valdez were released into Prince William Sound. Sol et al (2000) collected Dolly Varden (*Salvelinus malma*) 2-3 months after spill and found elevated levels of PAH metabolites in bile, associated with reduced plasma estradiol. Hom et al (1996; 2008) reported elevated PHN equivalents in pink salmon bile collected at an oiled site in 1989, compared to a reference location. Several studies documented that the Exxon Valdez spill adversely effected early life stages resulting in adverse effects on salmonid populations (Geiger et al., 1996; Incardona et al, 2013). One month following the North Cape oil spill of No. 2 fuel oil, winter flounder (*Pleuronectes americanus*) had elevated

levels of NPH and PHN equivalents in bile, compared to a reference site (Collier et al., 1997; Figure 10). It was concluded that exposure levels were sufficient to cause reproductive impairment, associated with reduced plasma estradiol levels. Effects to winter flounder early life stages were also reported (Hughes, 1999). Most recently, elevated levels of fish biliary PAH metabolites were reported following the 2010 Deepwater Horizon oil spill in the Gulf of Mexico (Snyder et al, 2015). An elevated incidence of skin lesions was observed in fish in 2011 and the incidence rate declined in 2012 (Murawski et al., 2014).



**Figure 10.** Comparison of biliary naphthalene (NPH) and phenanthrene (PHN) equivalents (ug/mg bile protein) mean concentrations from this study (2015, 2016 and salmon-48hr crude oil standard; ; barred surfperch at Gaviota and Refugio and walleye surfperch at Campus Point) and following other oil spills (Collier et al., 1997; Hom et al., 1996; Sol et al., 2000).

Biliary PAH metabolites have been shown to indicate uptake to fish from all exposure routes, providing an integrated estimate of recent PAH exposure (Meador et al., 2008). Many studies have used them as a biomarker of exposure for petroleum related contamination. In this assessment, a quantitative estimate of PAH exposure to surfperch was obtained, indicating that elevated exposure occurred in the surf zone at Refugio State Beach following the 2015 oil spill.

## REFERENCES

- Allen, L.G. and Pondella. D.J. 2006. Surf Zone, Coastal Pelagic Zone and Harbors. In: The Ecology of Marine Fishes: California and Adjacent Waters. Allen, L.G., D.J. Pondella and M.H. Horn eds. University of California Press, Berkeley CA. pp. 149-166
- Arkoosh, M.R., Clemons E., Huffman P, Kagley A.N., Casillas E., Adams N., Sanborn H.R., Collier T.K., Stein J.E. 2001. Increased susceptibility of juvenile Chinook salmon

to Vibriosis after exposure to chlorinated and aromatic compounds found in contaminated urban estuaries. *Journal of Aquatic Animal Health* 13(3):257-268.

California Department of Fish and Game. 2001. California's Living Marine Resources: A Status Report. Sacramento, CA.

Carlisle J.G., Schott J.W. and Abramson N.J. 1960. The Barred Surfperch (*Amphistichus argenteus*) in Southern California. State of California Department of Fish and Game, Marine Resources Operations Fish Bulletin No. 109.

Collier, T.K. and Varanasi U..1991. Hepatic activities of xenobiotic metabolizing enzymes and biliary levels of xenobiotics in English sole (*Parophrys vetulus*) exposed to environmental contaminants. *Archives Environmental Contamination and Toxicology* 20:462-473.

Collier, T. K., Johnson L.L., Hom T., Krahn M. M and Stein J. E. 1997. Oil Exposure of Fish in the Salt Ponds of Rhode Island Following the North Cape Oil Spill, and Estimation of the Potential for Biological Injury to Winter Flounder (*Pleuronectes americanus*). National Oceanic and Atmospheric Administration, Seattle WA.

da Silva D.A.M, Buzitis J., Krahn M.M., Bicego M.C., Pires-Vanin A.M.S. 2006. Metabolites in bile of fish from São Sebastião Channel, São Paulo, Brazil as biomarkers of exposure to petrogenic polycyclic aromatic compounds. *Marine Pollution Bulletin* 52:175-183.

Feder H.M., Turner C.H., Limbaugh C. 1974. Observations on fishes associated with kelp beds in southern California. Department of Fish and Game Fish Bulletin 160: 1-144.

Geiger H.J., Bue B.J., Sharr S., Wertheimer A.C., Willette T.M. 1996. A life history approach to estimating damage to Prince William Sound pink salmon caused by the Exxon Valdez oil spill. *American Fisheries Society Symposium* 18:487-498.

Heintz R.A., Rice S.D., Wertheimer A.C., Bradshaw R.F., Thrower F.P., Joyce J.E., Short J.W. 2000. Delayed effects on growth and marine survival of pink salmon *Onchorynchus gorbusha* after exposure to crude oil during embryonic development. 2000. *Marine Ecology Progress Series* 208:205-216.

Hom T, Varanasi U., Stein J. E., Sloan C. A., Tilbury K. L. and Chan S. 1996. Assessment of the exposure of subsistence fish to aromatic compounds after the Exxon Valdez Oil Spill. *American Fisheries Society Symposium* 18:856-866.

Hobson, E.S. and Chess, J.R. 1986. Relationships among fishes and their prey in a nearshore sand community off southern California. *Environmental Biology of Fishes* 17:201-226.

Hom T., Collier T.K., Krahn M. M., Strom M. S., Ylitalo G. M., Nilsson W. B., Paranjpye R. N., Varanasi U. 2008. Assessing Seafood Safety in the Aftermath of Hurricane Katrina. *American Fisheries Society Symposium* 64:73-93.

Hughes J.B. 1999. Cytological-cytogenetic analysis of winter flounder embryos collected from the benthos at the barge North Cape oil spill. *Marine Pollution Bulletin* 38:30-35.



Incardona J.P., Collier T.K., Scholz N.L. 2004. Defects in cardiac function precede morphological abnormalities in fish embryos exposed to polycyclic aromatic hydrocarbons. *Toxicology and Applied Pharmacology* 196:191-205.

Incardona J.P., Vines C.A., Anulacion B.F., Baldwin D.H., Day H.L., French B.L., Labenia J.S., Linbo T.L., Myers M.S., Olson O.P., Sloan C.A., Sol S., Griffin F.J., Menard K., Morgan S.G., West J.E., Collier T.K., Ylitalo G.M., Cherr G.N., Scholz N.L. 2012. Unexpectedly high mortality in Pacific herring embryos exposed to the 2007 Cosco Busan oil spill in San Francisco Bay. *Proceedings of the National Academy of Sciences* 109(2):E51-58.

Incardona J.P., Swarts T.L., Edmunds, R.C., Linbo, T.L., Aquilina-Beck, A., Sloan C.A., Gardner L.D., Block B.A. and Scholz N.L. 2013. Exxon Valdez to Deepwater Horizon: Comparable toxicity of both crude oils to fish early life stages. *Aquatic Toxicology* 142-143:303-316.

Kennedy D.M. and B.J. Baca. 1984. Fate and Effects of the Mobiloil Spill in the Columbia River. Report Submitted to the National Oceanic and Atmospheric Administration, Ocean Assessments Division.

Krahn M.M., Myers M.S., Burrows D.G., Malins D.C. 1984. Determination of xenobiotics in bile of fish from polluted waterways. *Xenobiotica* 14: 633-646.

Krahn M.M., Kittle L.J., MacLeod W.D. 1986. Evidence of exposure of fish to oil spilled into the Columbia River. *Marine Environmental Research* 20:291-298.

Lee, R.F., Anderson, J.W. 2005. Significance of cytochrome P450 system responses and levels of bile fluorescent aromatic compounds in marine wildlife following oil spills. *Marine Pollution Bulletin* 50:705-723.

Meador J.P., Sommers F.C., Ylitalo G.M., Sloan C.A. 2006. Altered growth and related physiological responses in juvenile Chinook salmon (*Oncorhynchus tshawytscha*) from dietary exposure to polycyclic aromatic hydrocarbons (PAHs). *Canadian Journal of Fisheries and Aquatic Sciences* 63: 2364-2376.

Meador, J.P., Buzitis J., Bravo C.F. 2008. Using fluorescent aromatic compounds from bile from juvenile salmonids to predict exposure to polycyclic aromatic hydrocarbons. *Environmental Toxicology and Chemistry* 27:845-853.

Murawski S.A., Hogarth W.T., Peebles E.B., Barbeiri L. 2014. Prevalence of external skin lesions and polycyclic aromatic hydrocarbon concentrations in Gulf of Mexico fishes, post-Deepwater Horizon. *Transactions of the American Fisheries Society* 143:1084-1097.

Myers M.S., Stehr C.M., Olson O.P., Johnson L.L., McCain B.B., Chan S-L., Varanasi U. 1994. Relationships between toxicopathic hepatic lesions and exposure to chemical contaminants in English sole (*Pleuronectes vetulus*), starry flounder (*Platichthys stellatus*), and white croaker (*Genyonemus lineatus*) from selected marine sites on the Pacific Coast, USA. *Environmental Health Perspectives* 102(2): 200-215.

OEHHA. 2015. Risk Assessment of Seafood Consumption Following the Refugio Beach Oil Spill Incident in Santa Barbara County, California. December. Office of

Environmental Health Hazard Assessment, California Environmental Protection Agency, Sacramento, CA.

Reynaud S., Deschaux P. 2006. The effects of polycyclic aromatic hydrocarbons on the immune system of fish: A review. *Aquatic Toxicology* 77: 229-238.

Roubal W.T., Collier T.K., Malins D.C. 1977. Accumulation and metabolism of carbon-14 labeled benzene, naphthalene, and anthracene by young coho salmon (*Oncorhynchus kisutch*). *Archives of Environmental Contamination and Toxicology* 5: 513-529.

Roy L.A., Steinert S., Bay S.M., Greenstein D., Sapozhnikova Y., Bawardi O., Leifer I., Schlenk D. 2003. Biochemical effects of petroleum exposure in hornyhead turbot (*Pleuronichthys verticalis*) exposed to a gradient of sediments collected from a natural petroleum seep in CA, USA. *Aquatic Toxicology* 65:159-169.

Sloan C.A., Brown D.W., Ylitalo G.M., Buzitis J., Herman D.P., Burrows D.G., Yanagida G.K., Pearce R.W., Bolton J.L., Boyer R.H., Krahn M.M. 2006. Quality assurance plan for analyses of environmental samples for polycyclic aromatic compounds, persistent organic pollutants, fatty acids, stable isotope ratios, lipid classes, and metabolites of polycyclic aromatic compounds. U.S. Department of Commerce, NOAA Technical Memorandum NMFS-NWFSC-77, 30 pp.

Snyder S.M., Pulster E.L., Wetzel D.L., Murawski S.A. 2015. PAH exposure in Gulf of Mexico demersal fishes, post-Deepwater Horizon. *Environmental Science and Technology* 49:8786-8795.

Sol S.Y., Johnson L.L., Horness B.H., Collier T.K. 2000. Relationship between oil exposure and reproductive parameters in fish collected following the Exxon Valdez oil spill. *Marine Pollution Bulletin* 40:1139-1147.

Spies R.B., Stegeman J.J., Hinton D.E., Woodin B., Smolowitz R., Okihiro M., Shea D. 1996. Biomarkers of hydrocarbon exposure and sublethal effects in embiotocid fishes from a natural petroleum seep in the Santa Barbara Channel. *Aquatic Toxicology* 34:195-219.

Stout, S.A. 2016. Refugio Beach Oil Spill NRDA Investigation: Trustees Forensic Oil Source Analysis. Draft Report. NewFields Environmental Forensics Practice, Rockland, Massachusetts.

Varanasi U., Stein J.E. and Nishimoto M. 1989. Biotransformation and disposition of polycyclic aromatic hydrocarbons (PAH) in fish. In: *Metabolism of Polycyclic Aromatic Hydrocarbons in the Aquatic Environment* (U Varanasi, Ed). CRC Press, Boca Raton, FL, pp. 94-149.

## **Appendices:**

**Appendix 1: Analysis of Bile of Fish for Metabolites of Polycyclic Aromatic Compounds (PACs): Results from Samples Collected following the Refugio Beach Oil Spill, May 2015**

**Appendix 2: Analysis of Bile of Fish for Metabolites of Polycyclic Aromatic Compounds (PACs): Results from Samples Collected One Year after the 2015 Refugio Beach Oil Spill**

## **Analysis of Bile of Fish for Metabolites of Polycyclic Aromatic Compounds (PACs): Results from Samples Collected Following the Refugio Beach Oil Spill, May 2015**

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### **Summary**

Analyses that screen for metabolites of polycyclic aromatic compounds (PACs) were conducted for bile samples of fish collected from three sites in Santa Barbara, CA in the area of the Refugio Beach oil spill. Bile samples were collected from barred surfperch (*Amphistichus argenteus*) from the three sites that included the spill site, a natural oil seep site that was not reported as being oiled the day prior to the 23 May 2015 sampling and a reference site. Additionally, walleye surfperch (*Hyperprosopon argenteum*) were collected at the natural seep site. Concentrations of bile PAC metabolites showed site differences in the 3 types of metabolites measured, naphthalene (NPH), phenanthrene (PHN) and benzo[*a*]pyrene (BaP), based on wet weight or protein content. The levels in fish collected at both the spill and natural seep sites were higher than the reference site, and the concentrations in fish from the spill site were higher overall than the natural seep site. Barred surfperch from the Refugio Beach oil spill site had the highest levels of PAC metabolites, with concentrations being an order of magnitude higher than barred surfperch from Gaviota, the reference site, and approximately two times higher than barred surfperch from Campus Point, the natural seep site. Concentrations of PAC metabolites measured in bile of barred surfperch and walleye surfperch collected from Campus Point, the natural seep site, were 3 to 8 times higher than those in barred surfperch from Gaviota, the reference site. Bile PAC metabolites levels in barred surfperch from Refugio Beach were 2 times higher than those measured in barred surfperch from Campus Point and were 2 to 5 times higher compared to the walleye surfperch collected at this seep site.

### **Introduction**

PACs are chemical contaminants that include polycyclic aromatic hydrocarbons (PAHs) such as naphthalene, phenanthrene, pyrene, benzo[*a*]pyrene), as well as heterocyclic aromatic compounds (e.g., dibenzothiophene) that are primarily derived from petroleum or their combustion products. Concerns have been raised over the effects of exposure to PACs, alone or in combination with other toxic contaminants, on terrestrial and marine organisms because of the worldwide use of fossil fuels (Geraci and St. Aubin 1990; Peterson et al., 2003) and the occurrence of oil spills in regions that support populations of fish, birds, turtles and amphibians. In other vertebrates, such as fish, biological effects associated with exposure to PACs include developmental abnormalities, immunosuppression, hepatic lesions and altered growth (Myers et al. 1994; Heintz et al. 2000; Arkoosh et al. 2001; Meador et al. 2006; Reynaud and Deschaux

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2006; Incardona et al. 2004; Incardona et al. 2012). Routes of PAC exposure in fish include consumption of contaminated food, inhalation, and dermal absorption.

Vertebrates (e.g., fish, marine mammals) rapidly take up PACs present in their food and the environment and quickly metabolize these compounds to more polar compounds. The more polar PAC metabolites are then secreted into fluids such as bile and urine for elimination via the gastroenteric tract or kidneys (Roubal et al., 1977; Varanasi et al., 1989; Krahn et al., 1984). Therefore, assessment of bile for PACs provides information on recent input and exposure to these compounds.

### Methods

Bile samples were collected on May 23, 2015, 4 days after the spill occurred. Bile of barred surfperch was collected from Refugio Beach (n= 20), Campus Point (n = 5), and Gaviota (n = 5). Bile from walleye surfperch was collected from Campus Point (n= 7). Bile samples were collected from the gall bladder immediately after fish were sacrificed, placed into 4mL amber vials and kept on ice. Then, the samples were frozen and transported to the Northwest Fisheries Science Center and stored at -20°C until analyses.

Bile samples were analyzed using a high-performance liquid chromatography/fluorescence (HPLC-F) method described in Krahn et al., 1984. Briefly, bile was injected directly onto a Waters high-performance liquid chromatography/fluorescence system equipped with a C-18 reverse-phase column (Phenomenex Synergi Hydro). The fluorescent PAC metabolites were eluted with a linear gradient from 100% water (containing a trace amount of acetic acid) to 100% methanol at a flow of 1.0 mL/min. Chromatograms were recorded at the following wavelength pairs: 1) 292/335 nm where many 2-3 benzene ring aromatic compounds (e.g., naphthalene) fluoresce, 2) 260/380 nm where several 3-4 ring compounds (e.g., phenanthrene) fluoresce and 3) 380/430 nm where 4-5 ring compounds (e.g., benzo[a]pyrene) fluoresce. Peaks eluting after 9 minutes were integrated and the areas of these peaks were summed. The concentrations of fluorescent PACs in the bile samples of the fish were determined using naphthalene (NPH), phenanthrene (PHN) or benzo[a]pyrene (BaP) as external standards and converting the fluorescence response of bile to phenanthrene (ng PHN equivalents/g bile), naphthalene (ng NPH equivalents/g bile) or benzo[a]pyrene (ng BaP equivalents/g bile) equivalents. In addition, protein analysis as described in da Silva et al. (2006) was completed for all bile samples as previous laboratory contaminant exposure studies on fish have shown that normalization of biliary PAC metabolite concentrations to protein values can help account for variation in metabolite levels based on feeding status (Collier and Varanasi 1991).

To ensure that the HPLC/fluorescence system was in proper operating condition, a NPH/PHN/BaP calibration standard was analyzed numerous times (n ≥ 5) until a relative standard deviation < 15% was obtained for each PAC. As part of our laboratory quality assurance (QA) plan (Sloan et al. 2006), a method blank and a fish bile control sample (bile of Atlantic salmon exposed to 25 µg/mL of Monterey crude oil for 48 hours) were analyzed with each batch of fish bile samples. In addition, an aliquot of a harbor seal bile sample (Bile\_Ref\_Mat) was also analyzed during the sample sequence as part of the QA plan.

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### Results

Concentrations of fluorescent PAC metabolites (based on wet weight or biliary protein) and levels of protein measured in the bile samples reported in Table 1. Two bile samples (RFB011BI and RFB001BI) were not analyzed for protein content due to inadequate bile volume (< 10  $\mu$ L). For the barred surfperch from Refugio Beach, NPH equivalent concentrations ranged from 780,000 to 4,400,000 ng/g bile, wet weight and 67,000 to 270,000 ng/mg protein, PHN equivalents ranged from 200,000 to 1,200,000 ng/g bile, wet weight and 20,000 to 74,000 ng/mg protein and BaP equivalent values ranged from 1,100 to 6,700 ng/g bile, wet weight and 91 to 410 ng/mg protein. For the barred surfperch from Campus Point, NPH equivalent concentrations ranged from 420,000 to 1,800,000 ng/g bile, wet weight and 31,000 to 130,000 ng/mg protein, PHN equivalents ranged from 150,000 to 500,000 ng/g bile, wet weight and 10,000 to 35,000 ng/mg protein and BaP equivalent values ranged from 880 to 2,300 ng/g bile, wet weight and 55 to 160 ng/mg protein. For the walleye surfperch from Campus Point, NPH equivalent concentrations ranged from 150,000 to 610,000 ng/g bile, wet weight and 11,000 to 52,000 ng/mg protein, PHN equivalents ranged from 49,000 to 200,000 ng/g bile, wet weight and 3,700 to 19,000 ng/mg protein and BaP equivalent values ranged from 550 to 3,500 ng/g bile, wet weight and 48 to 340 ng/mg protein. For the barred surfperch from Gaviota, NPH equivalent concentrations ranged from 86,000 to 200,000 ng/g bile, wet weight and 14,000 to 26,000 ng/mg protein, PHN equivalents ranged from 22,000 to 43,000 ng/g bile, wet weight and 3,400 to 5,500 ng/mg protein and BaP equivalent values ranged from 240 to 510 ng/g bile, wet weight and 32 to 65 ng/mg protein. In addition, biliary protein concentrations ranged from 6.2 to 21.0 mg/mL in the barred surfperch bile samples and 4.2 to 15.2 mg/mL in the walleye surfperch.

The mean ( $\pm$  SD) biliary NPH, PHN and BaP equivalent concentrations (ng/g bile, wet weight) measured in barred surfperch collected at Refugio Beach, Campus Point and Gaviota are shown in Figure 1A–C. Significant differences (ANOVA  $p < 0.05$ ; Tukey-Kramer HSD test) in mean NPH, PHN and BaP equivalent concentrations (based on wet weight or biliary protein) were found among collection sites. For each PAC metabolite, barred surfperch from the oiled site had a significantly higher mean level than those determined in fish from the seep site or from the reference site. Mean PAC metabolite concentrations measured in bile of barred surfperch from the seep site were significantly higher than the same metabolites measured in barred surfperch from the reference site. At the Campus Point site, mean levels of biliary NPH and PHN equivalents (based on wet weight only) were significantly higher (ANOVA  $p < 0.05$ ; t-test) in barred surfperch than those measured in walleye surfperch; no other significant differences were found for mean PAC equivalent concentrations between species.

### References

Arkoosh,MR, Clemons E, Huffman P, Kagley AN, Casillas E, Adams N, Sanborn HR, Collier TK, Stein JE. 2001. Increased susceptibility of juvenile Chinook salmon to Vibriosis after exposure to chlorinated and aromatic compounds found in contaminated urban estuaries. *Journal of Aquatic Animal Health* 13(3):257-268.

## Appendix 1

Collier, T.K. and U. Varanasi. 1991. Hepatic activities of xenobiotic metabolizing enzymes and biliary levels of xenobiotics in English sole (*Parophrys vetulus*) exposed to environmental contaminants. *Arch. Environ. Contam. Toxicol.* 20:462-473.

da Silva DAM, Buzitis J, Krahn MM, Bicego MC, Pires-Vanin AMS. 2006. Metabolites in bile of fish from São Sebastião Channel, São Paulo, Brazil as biomarkers of exposure to petrogenic polycyclic aromatic compounds. *Marine Pollution Bulletin* 52:175-183.

Geraci JR and St. Aubin DJ (eds.). 1990. *Sea Mammals and Oil: Confronting the Risks*, Academic Press, San Diego, CA, 282 pp.

Heintz RA, Rice SD, Werthheimer AC, Bradshaw RF, Thrower FP, Joyce JE, Short JW. 2000. Delayed effects on growth and marine survival of pink salmon *Onchorynchus gorboscha* after exposure to crude oil during embryonic development. 2000. *Marine Ecology Progress Series* 208:205-216.

Incardona JP, Collier TK, Scholz NL. 2004. Defects in cardiac function precede morphological abnormalities in fish embryos exposed to polycyclic aromatic hydrocarbons. *Toxicology and Applied Pharmacology* 196:191-205.

Incardona JP, Vines CA, Anulacion BF, Baldwin DH, Day HL, French BL, Labenia JS, Linbo TL, Myers MS, Olson OP, Sloan CA, Sol S, Griffin FJ, Menard K, Morgan SG, West JE, Collier TK, Ylitalo GM, Cherr GN, Scholz NL. 2012. Unexpectedly high mortality in Pacific herring embryos exposed to the 2007 Cosco Busan oil spill in San Francisco Bay. *Proceedings of the National Academy of Sciences* 109(2):E51-58.

Krahn MM, Myers MS, Burrows DG, Malins DC. 1984. Determination of xenobiotics in bile of fish from polluted waterways. *Xenobiotica* 14: 633-646.

Meador JP, Sommers FC, Ylitalo GM, Sloan CA. 2006. Altered growth and related physiological responses in juvenile Chinook salmon (*Oncorhynchus tshawytscha*) from dietary exposure to polycyclic aromatic hydrocarbons (PAHs). *Canadian Journal of Fisheries and Aquatic Sciences* 63: 2364-2376.

Myers MS, Stehr CM, Olson OP, Johnson LL, McCain BB, Chan S-L, Varanasi U. 1994. Relationships between toxicopathic hepatic lesions and exposure to chemical contaminants in English sole (*Pleuronectes vetulus*), starry flounder (*Platichthys stellatus*), and white croaker (*Genyonemus lineatus*) from selected marine sites on the Pacific Coast, USA. *Environmental Health Perspectives* 102(2): 200-215.

Peterson CH, Rice SD, Short JW, Esler D, Bodkin JL, Ballachey BE, Irons DB. 2003. Long-term ecosystem response to the Exxon Valdez oil spill. *Science* 302: 2082-2086.

Reynaud S., Deschaux P. 2006. The effects of polycyclic aromatic hydrocarbons on the immune system of fish: A review. *Aquatic Toxicology* 77: 229-238.

## Appendix 1

Roubal WT, Collier TK, Malins DC. 1977. Accumulation and metabolism of carbon-14 labeled benzene, naphthalene, and anthracene by young coho salmon (*Oncorhynchus kisutch*). *Archives of Environmental Contamination and Toxicology* 5: 513-529.

Sloan CA Brown, DW, Ylitalo GM, Buzitis J, Herman DP, Burrows DG, Yanagida GK, Pearce RW, Bolton JL, Boyer RH, Krahn MM. 2006. Quality assurance plan for analyses of environmental samples for polycyclic aromatic compounds, persistent organic pollutants, fatty acids, stable isotope ratios, lipid classes, and metabolites of polycyclic aromatic compounds. U.S. Department of Commerce, NOAA Technical Memorandum NMFS-NWFSC-77, 30 pp.

Varanasi U, Stein JE and Nishimoto M. 1989. Biotransformation and disposition of polycyclic aromatic hydrocarbons (PAH) in fish. In: *Metabolism of Polycyclic Aromatic Hydrocarbons in the Aquatic Environment* (U Varanasi, Ed). CRC Press, Boca Raton, FL, pp. 94-149.



Appendix 1

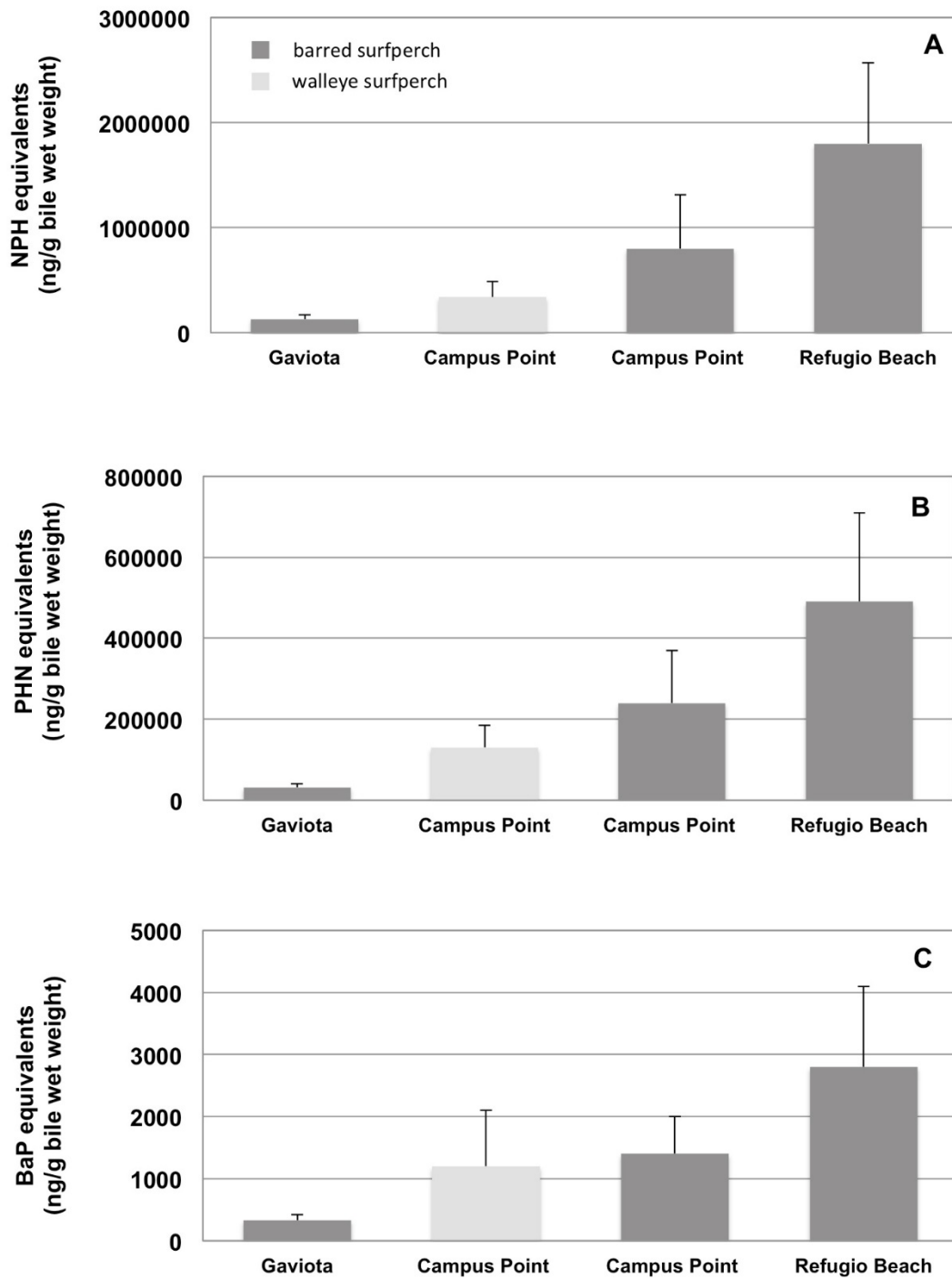


Figure 1. Mean ( $\pm$ SD) concentrations of bile equivalents of (A) naphthalene, NPH, (B) phenanthrene, PHN and (C) benzo[a]pyrene, BaP (ng/g bile wet weight) measured in two fish species collected from three sites in the area of the Refugio Beach oil spill.

Appendix 1

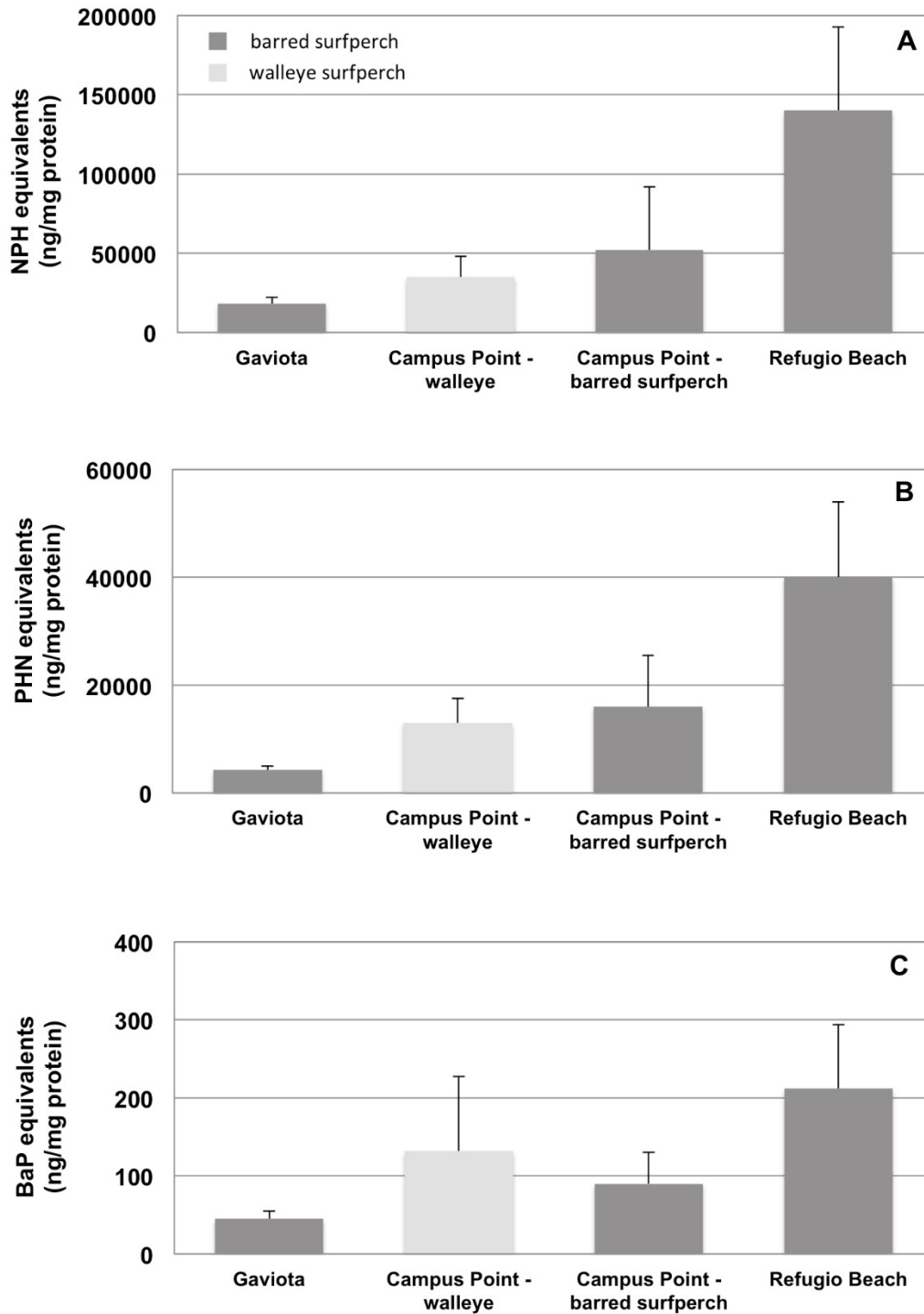


Figure 2. Mean ( $\pm$ SD) concentrations of bile equivalents (A) naphthalene, NPH, (B) phenanthrene, PHN and (C) benzo[a]pyrene, BaP (ng/mg protein) measured in two fish species collected from three sites in the area of the Refugio Beach oil spill.

Appendix 1

**Table 1. Concentrations of metabolites in polycyclic aromatic compounds measured in bile of fish collected on May 23, 2015 in the area of the Refugio Beach oil spill, Santa Barbara, CA**

| Site          | FIELD NUMBER | SPECIES          | Protein<br>mg/mL | Equivalents of fluorescent aromatic compounds<br>(ng/g bile, wet weight) |                                 |                                 | Equivalents of fluorescent aromatic compounds<br>(ng/mg protein) |                                 |                                 |
|---------------|--------------|------------------|------------------|--|---------------------------------|---------------------------------|--|---------------------------------|---------------------------------|
|               |              |                  |                  | NPH<br>Equivalents <sup>1</sup>  | PHN<br>Equivalents <sup>2</sup> | BaP<br>Equivalents <sup>3</sup> | NPH<br>Equivalents <sup>1</sup>                                  | PHN<br>Equivalents <sup>2</sup> | BaP<br>Equivalents <sup>3</sup> |
|               |              |                  |                  | Refugio Beach  | RFB009BI                        | barred surfperch                | 12.6   | 1,000,000                       | 280,000                         |
| Refugio Beach | RFB007BI     | barred surfperch | 15.1             | 3,100,000  | 940,000                         | 5,100                           | 210,000  | 62,000                          | 340                             |
| Refugio Beach | RFB015BI     | barred surfperch | 9.0              | 1,500,000  | 400,000                         | 2,300                           | 170,000  | 44,000                          | 260                             |
| Refugio Beach | RFB019BI     | barred surfperch | 9.9              | 2,100,000  | 560,000                         | 3,100                           | 210,000  | 57,000                          | 310                             |
| Refugio Beach | RFB012BI     | barred surfperch | 8.7              | 1,100,000  | 320,000                         | 2,000                           | 130,000  | 37,000                          | 230                             |
| Refugio Beach | RFB018BI     | barred surfperch | 16.3             | 4,400,000  | 1,200,000                       | 6,700                           | 270,000  | 74,000                          | 410                             |
| Refugio Beach | RFB002BI     | barred surfperch | 17.6             | 1,400,000  | 370,000                         | 1,600                           | 80,000   | 21,000                          | 91                              |
| Refugio Beach | RFB020BI     | barred surfperch | 9.3              | 1,700,000  | 440,000                         | 2,500                           | 180,000  | 47,000                          | 270                             |
| Refugio Beach | RFB017BI     | barred surfperch | 15.5             | 1,500,000  | 560,000                         | 2,800                           | 97,000   | 36,000                          | 180                             |
| Refugio Beach | RFB014BI     | barred surfperch | 19.7             | 1,900,000  | 530,000                         | 3,000                           | 96,000   | 27,000                          | 150                             |
| Refugio Beach | RFB004BI     | barred surfperch | 8.8              | 780,000  | 200,000                         | 1,100                           | 89,000   | 23,000                          | 130                             |
| Refugio Beach | RFB011BI     | barred surfperch | <i>ND</i>        | 2,200,000  | 560,000                         | 5,300                           |  |                                 |                                 |
| Refugio Beach | RFB008BI     | barred surfperch | 7.2              | 1,100,000  | 280,000                         | 1,300                           | 150,000  | 39,000                          | 180                             |

## Appendix 1

|                      |              |                   |           |           |         |       |         |        |     |
|----------------------|--------------|-------------------|-----------|-----------|---------|-------|---------|--------|-----|
| <b>Refugio Beach</b> | RFB006BI     | barred surfperch  | 16.1      | 2,200,000 | 620,000 | 3,300 | 140,000 | 39,000 | 200 |
| <b>Refugio Beach</b> | RFB010BI     | barred surfperch  | 8.6       | 1,500,000 | 430,000 | 2,400 | 170,000 | 50,000 | 280 |
| <b>Refugio Beach</b> | RFB001BI     | barred surfperch  | <i>ND</i> | 1,900,000 | 560,000 | 3,500 |         |        |     |
| <b>Refugio Beach</b> | RFB016BI     | barred surfperch  | 13.0      | 1,400,000 | 380,000 | 2,200 | 110,000 | 29,000 | 170 |
| <b>Refugio Beach</b> | RFB003BI     | barred surfperch  | 12.0      | 1,400,000 | 380,000 | 2,100 | 120,000 | 32,000 | 180 |
| <b>Refugio Beach</b> | RFB005BI     | barred surfperch  | 10.6      | 1,700,000 | 470,000 | 2,700 | 160,000 | 44,000 | 250 |
| <b>Refugio Beach</b> | RFB013BI     | barred surfperch  | 16.5      | 1,100,000 | 330,000 | 1,800 | 67,000  | 20,000 | 110 |
| <b>Campus Point</b>  | CMP002,003BI | barred surfperch  | 12.6      | 420,000   | 150,000 | 920   | 33,000  | 12,000 | 73  |
| <b>Campus Point</b>  | CMP004,005BI | barred surfperch  | 14.9      | 480,000   | 180,000 | 1,000 | 32,000  | 12,000 | 67  |
| <b>Campus Point</b>  | CMP006,007BI | barred surfperch  | 16.1      | 570,000   | 180,000 | 880   | 35,000  | 11,000 | 55  |
| <b>Campus Point</b>  | CMP008,009BI | barred surfperch  | 14.4      | 1,800,000 | 500,000 | 2,300 | 130,000 | 35,000 | 160 |
| <b>Campus Point</b>  | CMP010,011BI | barred surfperch  | 21.0      | 660,000   | 210,000 | 1,700 | 31,000  | 10,000 | 81  |
| <b>Campus Point</b>  | CMP017BI     | walleye surfperch | 4.2       | 220,000   | 79,000  | 550   | 52,000  | 19,000 | 130 |
| <b>Campus Point</b>  | CMP012,013BI | walleye surfperch | 7.3       | 380,000   | 140,000 | 1,900 | 52,000  | 19,000 | 260 |
| <b>Campus Point</b>  | CMP025,026BI | walleye surfperch | 11.7      | 610,000   | 190,000 | 1,200 | 52,000  | 16,000 | 100 |
| <b>Campus Point</b>  | CMP018,019BI | walleye surfperch | 10.3      | 240,000   | 97,000  | 3,500 | 23,000  | 9,400  | 340 |
| <b>Campus Point</b>  | CMP020,021BI | walleye surfperch | 13.1      | 150,000   | 49,000  | 630   | 11,000  | 3,700  | 48  |

## Appendix 1

|                     |              |                   |      |         |         |     |        |        |    |
|---------------------|--------------|-------------------|------|---------|---------|-----|--------|--------|----|
| <b>Campus Point</b> | CMP014,015BI | walleye surfperch | 15.2 | 500,000 | 200,000 | 990 | 33,000 | 13,000 | 65 |
| <b>Campus Point</b> | CMP023,024BI | walleye surfperch | 8.0  | 250,000 | 92,000  | 740 | 31,000 | 12,000 | 93 |
| <b>Gaviota</b>      | GAV002BI     | barred surfperch  | 6.2  | 96,000  | 26,000  | 270 | 15,000 | 4,200  | 44 |
| <b>Gaviota</b>      | GAV004,005BI | barred surfperch  | 7.9  | 110,000 | 27,000  | 250 | 14,000 | 3,400  | 32 |
| <b>Gaviota</b>      | GAV006,007BI | barred surfperch  | 7.4  | 140,000 | 35,000  | 340 | 19,000 | 4,700  | 46 |
| <b>Gaviota</b>      | GAV008BI     | barred surfperch  | 7.8  | 200,000 | 43,000  | 510 | 26,000 | 5,500  | 65 |
| <b>Gaviota</b>      | GAV009BI     | barred surfperch  | 8.7  | 170,000 | 40,000  | 390 | 20,000 | 4,600  | 45 |

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<sup>1</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of naphthalene standard at 292/335 nm wavelengths.

<sup>2</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of phenanthrene standard at 260/380 nm wavelengths.

<sup>3</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of benzo[a]pyrene standard at 380/430 nm wavelengths.

*ND* – no data due to inadequate amount of sample to conduct protein analyses.

**Appendix 2: Analysis of Bile of Fish for Metabolites of Polycyclic  
Aromatic Compounds (PACs):  
Results from Samples Collected One Year after the 2015 Refugio  
Beach Oil Spill**

### **Analysis of Bile of Fish for Metabolites of Polycyclic Aromatic Compounds (PACs): Results from Samples Collected One Year after the 2015 Refugio Beach Oil Spill**

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#### **Summary**

Analyses that screen for metabolites of polycyclic aromatic compounds (PACs) have been completed for bile samples of fish collected from three sites in Santa Barbara, CA one year after the 2015 Refugio Beach oil spill. Bile samples were collected from barred surfperch (*Amphistichus argenteus*) and walleye surfperch (*Hyperprosopon argenteum*) from the three sites that included the spill site, a natural oil seep site that was not reported as being oiled the day prior to the 23 May 2015 sampling but may have had some oil exposure subsequently and a reference site. Mean biliary concentrations of PAC metabolites did not show significant site differences in the three types of metabolites measured, naphthalene (NPH), phenanthrene (PHN) and benzo[*a*]pyrene (BaP), based on wet weight or protein content except the mean PHN equivalent concentrations in barred surfperch from the oiled Refugio Beach site and the reference site (Gaviota). At the Refugio Beach oiled site, mean concentrations of all PAC metabolites in the 2016 collected fish were significantly higher in walleye surfperch compared to barred surfperch except protein-corrected NPH. Barred surfperch collected at Refugio Beach in 2015 had significantly higher mean NPH, PHN and BaP concentrations (wet weight and protein-corrected) than those determined in the 2016 barred surfperch collected one year later. In contrast, Gaviota barred surfperch collected in 2016 had significantly higher mean concentrations of NPH equivalents (wet weight and protein-corrected) and PHN equivalents (wet weight only) than the mean values of the 2015 Gaviota barred surfperch. Similarly, walleye surfperch collected from Campus Point in 2016 had significantly higher mean concentrations of protein-corrected NPH, PHN and BaP equivalents, as well as NPH equivalents (wet weight only) than walleye collected from this site in 2015.

#### **Introduction**

Polycyclic aromatic hydrocarbons (PACs) are chemical contaminants that include polycyclic aromatic hydrocarbons (PAHs) such as naphthalene, phenanthrene, pyrene, benzo[*a*]pyrene), as well as heterocyclic aromatic compounds (e.g., dibenzothiophene) that are primarily derived from petroleum or their combustion products. Concerns have been raised over the effects of exposure to PACs, alone or in combination with other toxic contaminants, on terrestrial and marine organisms because of the worldwide use of fossil fuels (Geraci and St. Aubin 1990; Peterson et al., 2003) and the occurrence of oil spills in regions that support populations of fish, birds, turtles and amphibians. In other vertebrates, such as fish, biological effects associated with exposure to PACs include developmental abnormalities, immunosuppression, hepatic

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lesions and altered growth (Myers et al. 1994; Heintz et al. 2000; Arkoosh et al. 2001; Meador et al. 2006; Reynaud and Deschaux 2006; Incardona et al. 2004; Incardona et al. 2012). Routes of PAC exposure in fish include consumption of contaminated food or sediment, respiration and dermal absorption.

Vertebrates (e.g., fish, marine mammals) rapidly take up PACs present in their food and the environment and quickly metabolize these compounds to more polar compounds. The more polar PAC metabolites are then secreted into fluids such as bile and urine for elimination via the gastroenteric tract or kidneys (Roubal et al., 1977; Varanasi et al., 1989; Krahn et al., 1984). Therefore, assessment of bile for PACs provides information on recent input and exposure to these compounds.

### Methods

Bile samples were collected on May 18, 2016, one year after the Refugio Beach oil spill occurred. Bile samples of barred surfperch were collected from Refugio Beach (n = 7), Campus Point (n = 1), and Gaviota (n = 5). Bile from walleye surfperch were collected from Refugio Beach (n = 13), Campus Point (n = 19), and Gaviota (n = 1). Bile samples were collected from the gall bladder immediately after fish were sacrificed, placed into 4mL amber vials and kept on ice. Then, the samples were frozen and transported to the Northwest Fisheries Science Center and stored at -20°C until analyses.

Bile samples were analyzed using a high-performance liquid chromatography/fluorescence (HPLC-F) method described in Krahn et al., 1984. Briefly, bile was injected directly onto a Waters high-performance liquid chromatography/fluorescence system equipped with a C-18 reverse-phase column (Phenomenex Synergi Hydro). The fluorescent PAC metabolites were eluted with a linear gradient from 100% water (containing a trace amount of acetic acid) to 100% methanol at a flow of 1.0 mL/min. Chromatograms were recorded at the following wavelength pairs: 1) 292/335 nm where many 2-3 benzene ring aromatic compounds (e.g., naphthalene) fluoresce, 2) 260/380 nm where several 3-4 ring compounds (e.g., phenanthrene) fluoresce and 3) 380/430 nm where 4-5 ring compounds (e.g., benzo[a]pyrene) fluoresce. Peaks eluting after 9 minutes were integrated and the areas of these peaks were summed. The concentrations of fluorescent PACs in the bile samples of the fish were determined using naphthalene (NPH), phenanthrene (PHN) or benzo[a]pyrene (BaP) as external standards and converting the fluorescence response of bile to phenanthrene (ng PHN equivalents/g bile), naphthalene (ng NPH equivalents/g bile) or benzo[a]pyrene (ng BaP equivalents/g bile) equivalents. In addition, protein analysis as described in da Silva et al. (2006) was completed for all bile samples as previous laboratory contaminant exposure studies on fish have shown that normalization of biliary PAC metabolite concentrations to protein values can help account for variation in metabolite levels based on feeding status (Collier and Varanasi 1991).

To ensure that the HPLC/fluorescence system was in proper operating condition, a NPH/PHN/BaP calibration standard was analyzed numerous times (n ≥ 5) until a relative standard deviation < 15% was obtained for each PAC. As part of our laboratory quality assurance (QA) plan (Sloan et al. 2006), a method blank and a fish bile control sample (bile of



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Atlantic salmon exposed to 25 µg/mL of Monterey crude oil for 48 hours) were analyzed with each batch of fish bile samples.

### Results

Concentrations of fluorescent PAC metabolites (based on wet weight or biliary protein) and protein content measured in the bile samples collected in 2016 are reported in Table 1. One bile sample from a walleye surfperch from Refugio Beach (RSBF11051816BI7) was not analyzed for protein content due to inadequate bile volume (< 10 µL). All sample batches met our laboratory quality assurance criteria (Tables 2 and 3).

A wide range of fluorescent PAC metabolite concentrations were measured in the bile of the fish collected in 2016 (Table 1). For the barred surfperch from Refugio Beach, the NPH equivalent concentrations ranged from 230,000 to 590,000 ng/g bile, wet weight and 23,000 to 78,000 ng/mg protein, PHN equivalents ranged from 51,000 to 140,000 ng/g bile, wet weight and 4,900 to 22,000 ng/mg protein and BaP equivalent values ranged from 430 to 1,200 ng/g bile, wet weight and 33 to 180 ng/mg protein. For the barred surfperch from Campus Point, NPH equivalent concentrations were 550,000 ng/g bile, wet weight and 63,000 ng/mg protein, PHN equivalents were 140,000 ng/g bile, wet weight and 16,000 ng/mg protein and BaP equivalent values were 1,300 ng/g bile, wet weight and 150 ng/mg protein. For the barred surfperch from Gaviota, the NPH equivalent concentrations ranged from 190,000 to 530,000 ng/g bile, wet weight and 22,000 to 44,000 ng/mg protein, PHN equivalents ranged from 44,000 to 100,000 ng/g bile, wet weight and 4,400 to 9,800 ng/mg protein and BaP equivalent values ranged from 540 to 7,200 ng/g bile, wet weight and 29 to 650 ng/mg protein. For the walleye surfperch from Refugio Beach, the NPH equivalent concentrations ranged from 410,000 to 6,700,000 ng/g bile, wet weight and 29,000 to 420,000 ng/mg protein, PHN equivalents ranged from 93,000 to 2,000,000 ng/g bile, wet weight and 7,300 to 130,000 ng/mg protein and BaP equivalent values ranged from 550 to 19,000 ng/g bile, wet weight and 74 to 1,200 ng/mg protein. For the walleye surfperch from Campus Point, the NPH equivalent concentrations ranged from 150,000 to 1,100,000 ng/g bile, wet weight and 19,000 to 94,000 ng/mg protein, PHN equivalents ranged from 41,000 to 290,000 ng/g bile, wet weight and 5,100 to 27,000 ng/mg protein and BaP equivalent values ranged from 550 to 5,900 ng/g bile, wet weight and 68 to 840 ng/mg protein. For the walleye surfperch from Gaviota, the NPH equivalent concentrations were 810,000 ng/g bile, wet weight and 91,000 ng/mg protein, PHN equivalents were 230,000 ng/g bile, wet weight and 26,000 ng/mg protein and BaP equivalent values were 2,200 ng/g bile, wet weight and 250 ng/mg protein. Biliary protein concentrations in barred surfperch and walleye surfperch ranged from 4.5 to 22.5 ng/mg and 4.3 to 19.9 ng/mg, respectively.

Mean NPH, PHN and BaP equivalent concentrations based on wet weight (Figure 1) or biliary protein (data not shown) ( $\log_{10}$  transformed data) for each species collected at the oiled (Refugio Beach), natural seep (Campus Point) and reference (Gaviota) sites were not significantly different (ANOVA  $p > 0.05$ ) in fish collected one year after the spill except PHN

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equivalent levels (wet weight only) in barred surfperch from the oiled Refugio Beach site and Gaviota (reference site) ( $p = 0.0487$ ). This is in contrast to our findings for barred surfperch collected at the same three sampling sites approximately one week (May 2015) after the spill occurred (Anulacion and Ylitalo 2015), in which the mean PAC metabolite levels in fish collected at both the spill and natural seep sites were significantly higher than those at the reference site, and the concentrations in barred surfperch from the spill site were higher overall than the natural seep site.

At the Refugio Beach site, mean concentrations of all PAC metabolites based on wet weight (Figure 1) ( $\log_{10}$  transformed data) in the 2016 collected fish were significantly higher in walleye surfperch compared to barred surfperch (ANOVA  $p < 0.05$ ; t-test). Similarly, mean levels of protein-corrected PAC metabolites were significantly higher (ANOVA  $p < 0.05$ ; t-test) in walleye compared to barred surfperch except protein-corrected NPH ( $p = 0.0505$ ). Comparisons between species at the other two collection sites were not conducted due to inadequate numbers of bile samples ( $n = 1$  for barred surfperch from Campus Point and  $n = 1$  for surfperch from Gaviota).

We examined differences in mean concentrations of PAC metabolites based on sampling year for each species collected at the same site (Figure 2). Barred surfperch collected at Refugio Beach in 2015 had significantly higher (ANOVA  $p < 0.0001$ ; t-test) mean NPH, PHN and BaP concentrations (wet weight and protein-corrected) ( $\log_{10}$  transformed data) than those determined in the 2016 collected barred surfperch. In contrast, Gaviota barred surfperch collected in 2016 had significantly higher mean concentrations ( $\log_{10}$  transformed data) (ANOVA  $p < 0.05$ ; t-test) of NPH equivalents (wet weight and protein-corrected) and PHN equivalents (wet weight only) than the mean values of the 2015 fish; no other significant differences were found for barred surfperch from this site. Walleye surfperch collected from Campus Point in 2016 had significantly higher mean concentrations of protein-corrected NPH, PHN and BaP equivalents, as well as NPH equivalents (wet weight only). No other significant differences (ANOVA  $p > 0.05$ ) in mean concentrations of PAC metabolites were found for walleye surfperch from this site.

## References

Anulacion BF, Ylitalo GM. 2015. Final Technical Report entitled "Analysis of bile of fish for metabolites of polycyclic aromatic compounds (PACs) Results from samples collected following the Refugio Beach oil spill, May 2015." Sent to NRDA Trustees on July 29, 2015

Arkoosh MR, Clemons E, Huffman P, Kagle AN, Casillas E, Adams N, Sanborn HR, Collier TK, Stein JE. 2001. Increased susceptibility of juvenile Chinook salmon to Vibriosis after exposure to chlorinated and aromatic compounds found in contaminated urban estuaries. *Journal of Aquatic Animal Health* 13(3):257-268.

Collier TK, Varanasi U. 1991. Hepatic activities of xenobiotic metabolizing enzymes and biliary levels of xenobiotics in English sole (*Parophrys vetulus*) exposed to environmental contaminants. *Arch. Environ. Contam. Toxicol.* 20:462-473.

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da Silva DAM, Buzitis J, Krahn MM, Bicego MC, Pires-Vanin AMS. 2006. Metabolites in bile of fish from São Sebastião Channel, São Paulo, Brazil as biomarkers of exposure to petrogenic polycyclic aromatic compounds. *Marine Pollution Bulletin* 52:175-183.

Geraci JR and St. Aubin DJ (eds.). 1990. *Sea Mammals and Oil: Confronting the Risks*, Academic Press, San Diego, CA, 282 pp.

Heintz RA, Rice SD, Werthheimer AC, Bradshaw RF, Thrower FP, Joyce JE, Short JW. 2000. Delayed effects on growth and marine survival of pink salmon *Onchorynchus gorbuscha* after exposure to crude oil during embryonic development. 2000. *Marine Ecology Progress Series* 208:205-216.

Incardona JP, Collier TK, Scholz NL. 2004. Defects in cardiac function precede morphological abnormalities in fish embryos exposed to polycyclic aromatic hydrocarbons. *Toxicology and Applied Pharmacology* 196:191-205.

Incardona JP, Vines CA, Anulacion BF, Baldwin DH, Day HL, French BL, Labenia JS, Linbo TL, Myers MS, Olson OP, Sloan CA, Sol S, Griffin FJ, Menard K, Morgan SG, West JE, Collier TK, Ylitalo GM, Cherr GN, Scholz NL. 2012. Unexpectedly high mortality in Pacific herring embryos exposed to the 2007 Cosco Busan oil spill in San Francisco Bay. *Proceedings of the National Academy of Sciences* 109(2):E51-58.

Krahn MM, Myers MS, Burrows DG, Malins DC. 1984. Determination of xenobiotics in bile of fish from polluted waterways. *Xenobiotica* 14:633-646.

Meador JP, Sommers FC, Ylitalo GM, Sloan CA. 2006. Altered growth and related physiological responses in juvenile Chinook salmon (*Oncorhynchus tshawytscha*) from dietary exposure to polycyclic aromatic hydrocarbons (PAHs). *Canadian Journal of Fisheries and Aquatic Sciences* 63:2364-2376.

Myers MS, Stehr CM, Olson OP, Johnson LL, McCain BB, Chan S-L, Varanasi U. 1994. Relationships between toxicopathic hepatic lesions and exposure to chemical contaminants in English sole (*Pleuronectes vetulus*), starry flounder (*Platichthys stellatus*), and white croaker (*Genyonemus lineatus*) from selected marine sites on the Pacific Coast, USA. *Environmental Health Perspectives* 102(2):200-215.

Peterson CH, Rice SD, Short JW, Esler D, Bodkin JL, Ballachey BE, Irons DB. 2003. Long-term ecosystem response to the Exxon Valdez oil spill. *Science* 302: 2082-2086.

Reynaud S., Deschaux P. 2006. The effects of polycyclic aromatic hydrocarbons on the immune system of fish: A review. *Aquatic Toxicology* 77:229-238.

Roubal WT, Collier TK, Malins DC. 1977. Accumulation and metabolism of carbon-14 labeled benzene, naphthalene, and anthracene by young coho salmon (*Oncorhynchus kisutch*). *Archives of Environmental Contamination and Toxicology* 5:513-529.

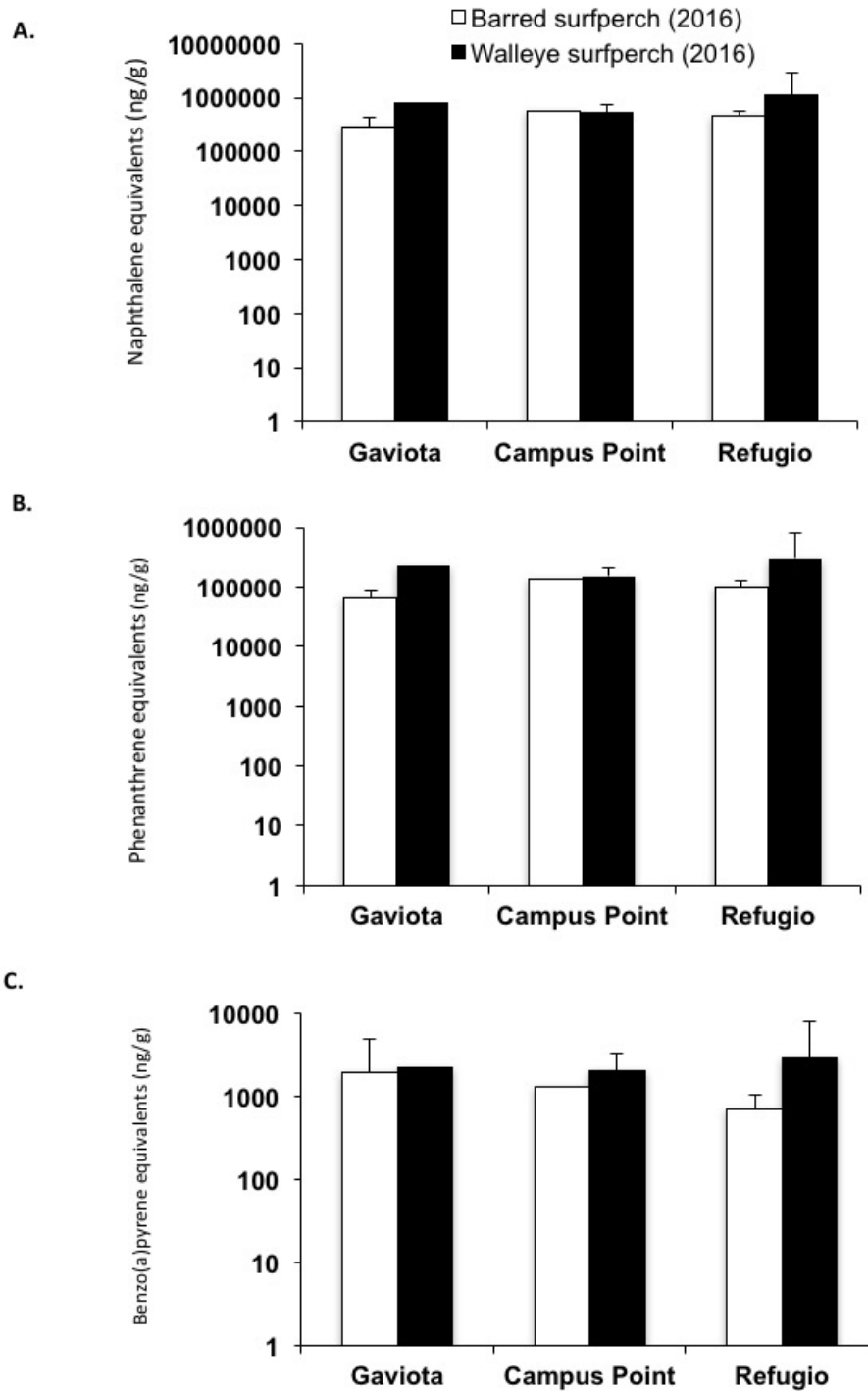
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Sloan CA Brown, DW, Ylitalo GM, Buzitis J, Herman DP, Burrows DG, Yanagida GK, Pearce RW, Bolton JL, Boyer RH, Krahn MM. 2006. Quality assurance plan for analyses of environmental samples for polycyclic aromatic compounds, persistent organic pollutants, fatty acids, stable isotope ratios, lipid classes, and metabolites of polycyclic aromatic compounds. U.S. Department of Commerce, NOAA Technical Memorandum NMFS-NWFSC-77, 30 pp.

Varanasi U, Stein JE, Nishimoto M. 1989. Biotransformation and disposition of polycyclic aromatic hydrocarbons (PAH) in fish. In: *Metabolism of Polycyclic Aromatic Hydrocarbons in the Aquatic Environment* (U Varanasi, Ed). CRC Press, Boca Raton, FL, pp. 94-149.

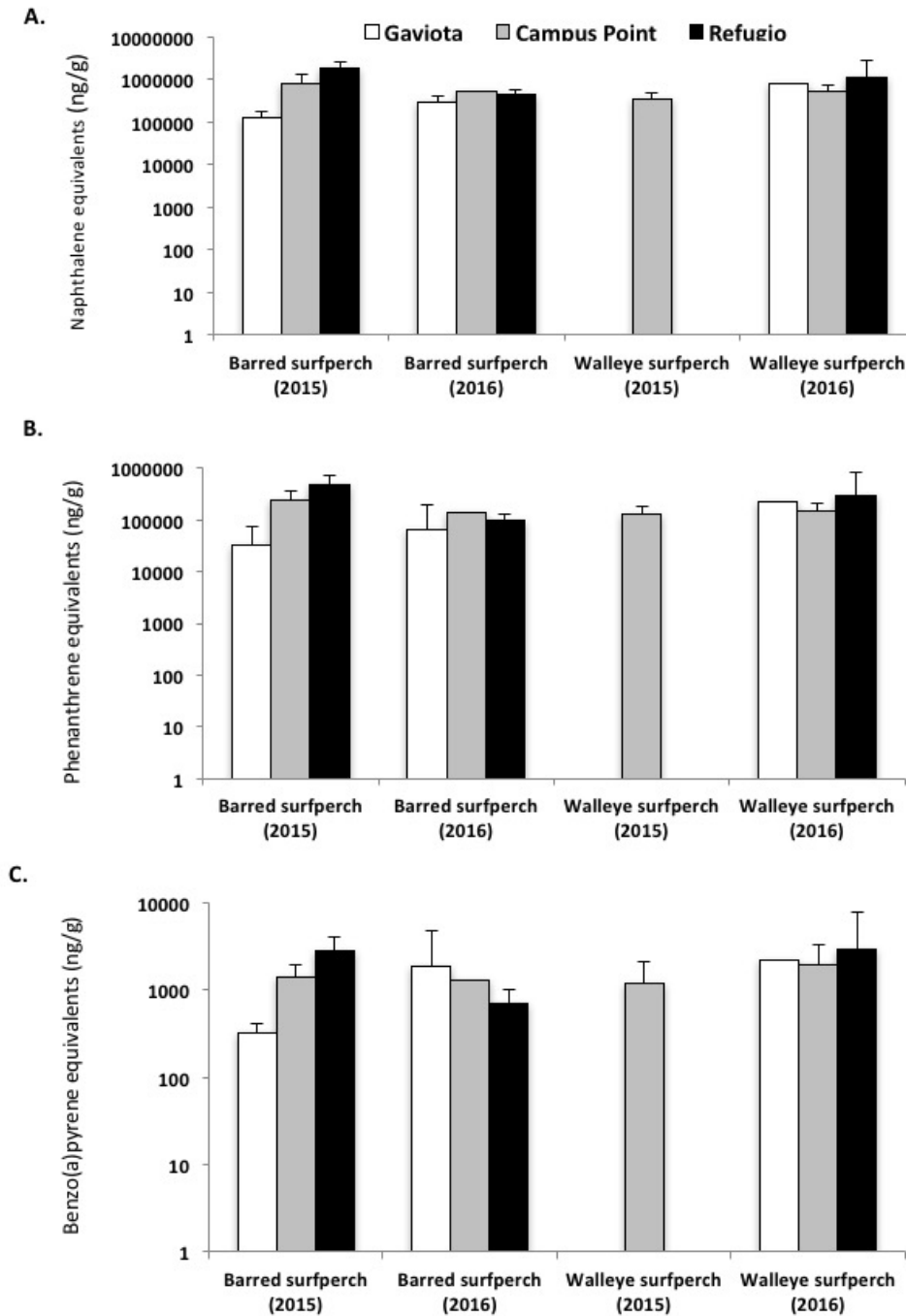
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Figure 1. Mean ( $\pm$ SD) concentrations of bile equivalents of (A) naphthalene, NPH, (B) phenanthrene, PHN and (C) benzo[a]pyrene, BaP (ng/g bile wet weight) measured in two fish species collected one year following the 2015 Refugio Beach oil spill.



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Figure 2. Mean ( $\pm$ SD) concentrations of bile equivalents (A) naphthalene, NPH, (B) phenanthrene, PHN and (C) benzo[a]pyrene, BaP (ng/g bile wet weight) measured in two fish species collected in 2015 and 2016 showing the differences in mean concentrations of PAC metabolites based on sampling year for each species collected at the same site



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**Table 1. Concentrations of metabolites of polycyclic aromatic compounds measured in bile of barred surfperch and walleye surfperch collected in 2016, one year after the 2015 Refugio Beach Oil Spill.**

| Species                 | Laboratory ID number     | Collection site | Field ID number  | Protein, mg/mL   | Equivalents of fluorescent aromatic compounds (ng/g bile, wet weight) |                  |                  | Protein-corrected equivalents of fluorescent aromatic compounds (ng/mg biliary protein) |                  |                  |     |
|-------------------------|--------------------------|-----------------|------------------|------------------|---|------------------|------------------|---|------------------|------------------|-----|
|                         |                          |                 |                  |                  | NPH <sup>1</sup>  | PHN <sup>2</sup> | BaP <sup>3</sup> | NPH <sup>1</sup>  | PHN <sup>2</sup> | BaP <sup>3</sup> |     |
| <i>Barred surfperch</i> | 136-0436                 | Campus Point    | CMPFI1051816BI20 | 8.8              | 550,000   | 140,000          | 1,300            | 63,000  | 16,000           | 150              |     |
|                         | 136-0444                 | Gaviota         | GAVFI1051816BI5  | 4.5              | 200,000   | 44,000           | 540              | 44,000  | 9,800            | 120              |     |
|                         | 136-0393                 | Gaviota         | GAVFI1051816BI4  | 7.7              | 220,000   | 47,000           | 560              | 29,000  | 6,100            | 73               |     |
|                         | 136-0422                 | Gaviota         | GAVFI1051816BI2  | 11.1             | 240,000   | 63,000           | 7,200            | 22,000  | 5,700            | 650              |     |
|                         | 136-0434                 | Gaviota         | GAVFI1051816BI3  | 10.4             | 280,000   | 64,000           | 630              | 27,000  | 6,200            | 61               |     |
|                         | 136-0404                 | Gaviota         | GAVFI1051816BI1  | 22.5             | 530,000   | 100,000          | 650              | 24,000  | 4,400            | 29               |     |
|                         | 136-0406                 | Refugio Beach   | RSBFI1051816BI4  | 7.1              | 240,000   | 52,000           | 460              | 34,000  | 7,300            | 65               |     |
|                         | 136-0400                 | Refugio Beach   | RSBFI1051816BI5  | 15.4             | 360,000   | 75,000           | 510              | 23,000  | 4,900            | 33               |     |
|                         | 136-0414                 | Refugio Beach   | RSBFI1051816BI6  | 14.6             | 360,000   | 80,000           | 500              | 25,000  | 5,500            | 34               |     |
|                         | 136-0412                 | Refugio Beach   | RSBFI1051816BI1  | 14.4             | 500,000   | 110,000          | 950              | 35,000  | 7,600            | 66               |     |
|                         | 136-0402                 | Refugio Beach   | RSBFI1051816BI17 | 6.5              | 510,000   | 140,000          | 1,200            | 78,000  | 22,000           | 180              |     |
|                         | 136-0431                 | Refugio Beach   | RSBFI1051816BI2  | 20.2             | 560,000   | 120,000          | 910              | 28,000  | 5,900            | 45               |     |
|                         | 136-0432                 | Refugio Beach   | RSBFI1051816BI3  | 10.3             | 590,000   | 120,000          | 430              | 57,000  | 12,000           | 42               |     |
|                         | <i>Walleye surfperch</i> | 136-0413        | Campus Point     | CMPFI1051816BI10 | 8.1   | 150,000          | 41,000           | 550   | 19,000           | 5,100            | 68  |
|                         |                          | 136-0405        | Campus Point     | CMPFI1051816BI17 | 5.5   | 290,000          | 81,000           | 4,600   | 53,000           | 15,000           | 840 |
|                         |                          | 136-0430        | Campus Point     | CMPFI1051816BI8  | 4.3   | 290,000          | 81,000           | 1,000   | 67,000           | 19,000           | 230 |
| 136-0440                |                          | Campus Point    | CMPFI1051816BI11 | 4.5              | 330,000   | 98,000           | 1,100            | 73,000  | 22,000           | 240              |     |
| 136-0398                |                          | Campus Point    | CMPFI1051816BI1  | 4.5              | 360,000   | 110,000          | 1,100            | 80,000  | 24,000           | 240              |     |
| 136-0443                |                          | Campus Point    | CMPFI1051816BI3  | 5.7              | 370,000   | 110,000          | 1,100            | 65,000  | 19,000           | 190              |     |
| 136-0401                |                          | Campus Point    | CMPFI1051816BI19 | 8.8              | 390,000   | 98,000           | 5,900            | 44,000  | 11,000           | 670              |     |
| 136-0437                |                          | Campus Point    | CMPFI1051816BI2  | 10.9             | 410,000   | 130,000          | 1,000            | 38,000  | 12,000           | 92               |     |
| 136-0397                |                          | Campus Point    | CMPFI1051816BI18 | 4.9              | 440,000   | 120,000          | 2,900            | 90,000  | 24,000           | 590              |     |
| 136-0396                |                          | Campus Point    | CMPFI1051816BI4  | 6.1              | 480,000   | 120,000          | 1,700            | 79,000  | 20,000           | 280              |     |
| 136-0421                |                          | Campus Point    | CMPFI1051816BI14 | 7.4              | 510,000   | 160,000          | 1,500            | 69,000  | 22,000           | 200              |     |
| 136-0435                |                          | Campus Point    | CMPFI1051816BI16 | 6.8              | 540,000   | 170,000          | 1,700            | 79,000  | 25,000           | 250              |     |
| 136-0419                |                          | Campus Point    | CMPFI1051816BI12 | 9.1              | 550,000   | 150,000          | 1,800            | 60,000  | 16,000           | 200              |     |
| 136-0442                |                          | Campus Point    | CMPFI1051816BI13 | 8.4              | 670,000   | 200,000          | 1,900            | 80,000  | 24,000           | 230              |     |
| 136-0418                |                          | Campus Point    | CMPFI1051816BI15 | 8.6              | 730,000   | 230,000          | 2,200            | 85,000  | 27,000           | 260              |     |
| 136-0420                |                          | Campus Point    | CMPFI1051816BI6  | 7.8              | 730,000   | 210,000          | 1,800            | 94,000  | 27,000           | 230              |     |
| 136-0415                |                          | Campus Point    | CMPFI1051816BI5  | 15.1             | 740,000   | 170,000          | 2,000            | 49,000  | 11,000           | 130              |     |
| 136-0394                |                          | Campus Point    | CMPFI1051816BI9  | 9.1              | 760,000   | 200,000          | 2,700            | 84,000  | 22,000           | 300              |     |
| 136-0425                |                          | Campus Point    | CMPFI1051816BI7  | 15.3             | 1,100,000   | 290,000          | 2,400            | 72,000  | 19,000           | 160              |     |
| 136-0408                |                          | Gaviota         | GAVFI1051816BI6  | 8.9              | 810,000   | 230,000          | 2,200            | 91,000  | 26,000           | 250              |     |
| 136-0399                |                          | Refugio Beach   | RSBFI1051816BI15 | 9.4              | 410,000   | 93,000           | 1,300            | 44,000  | 9,900            | 140              |     |
| 136-0426                |                          | Refugio Beach   | RSBFI1051816BI14 | 6.8              | 420,000   | 120,000          | 1,000            | 62,000  | 18,000           | 150              |     |
| 136-0407                |                          | Refugio Beach   | RSBFI1051816BI7  | IS <sup>4</sup>  | 470,000   | 96,000           | 550              | ND  | ND               | ND               |     |
| 136-0433                |                          | Refugio Beach   | RSBFI1051816BI16 | 13.3             | 520,000   | 120,000          | 1,300            | 39,000  | 9,000            | 98               |     |
| 136-0441                |                          | Refugio Beach   | RSBFI1051816BI18 | 17.8             | 520,000   | 130,000          | 1,400            | 29,000  | 7,300            | 79               |     |
| 136-0438                |                          | Refugio Beach   | RSBFI1051816BI12 | 14.8             | 540,000   | 110,000          | 1,100            | 36,000  | 7,400            | 74               |     |
| 136-0439                |                          | Refugio Beach   | RSBFI1051816BI20 | 8.1              | 550,000   | 160,000          | 1,600            | 68,000  | 20,000           | 200              |     |
| 136-0416                |                          | Refugio Beach   | RSBFI1051816BI13 | 11.1             | 600,000   | 170,000          | 1,900            | 54,000  | 15,000           | 170              |     |
| 136-0417                |                          | Refugio Beach   | RSBFI1051816BI10 | 5.6              | 740,000   | 200,000          | 1,600            | 130,000   | 36,000           | 290              |     |
| 136-0424                |                          | Refugio Beach   | RSBFI1051816BI9  | 14.5             | 780,000   | 210,000          | 2,100            | 54,000  | 14,000           | 140              |     |
| 136-0395                |                          | Refugio Beach   | RSBFI1051816BI11 | 12.3             | 870,000   | 240,000          | 2,300            | 71,000  | 20,000           | 190              |     |
| 136-0423                |                          | Refugio Beach   | RSBFI1051816BI8  | 19.9             | 1,200,000   | 320,000          | 2,800            | 60,000  | 16,000           | 140              |     |
| 136-0403                |                          | Refugio Beach   | RSBFI1051816BI19 | 15.9             | 6,700,000   | 2,000,000        | 19,000           | 420,000   | 130,000          | 1,200            |     |

<sup>1</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of naphthalene standard at 292/335 nm wavelengths.

<sup>2</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of phenanthrene standard at 260/380 nm wavelengths.

<sup>3</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of benzo[a]pyrene standard at 380/430 nm wavelengths.

<sup>4</sup>IS - insufficient sample available for protein analysis

ND = protein-corrected PAC metabolite concentrations not determined due to insufficient sample available for protein analysis

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**Table 2. Concentrations of metabolites of polycyclic aromatic compounds measured in bile reference material and method blanks analyzed one year after the 2015 Refugio Beach Oil Spill.**

| Quality assurance sample type    | Quality assurance sample information                            | Analysis date | Equivalents of fluorescent aromatic compounds (ng/g bile, wet weight) |                  |                  |
|----------------------------------|---|---------------|---|------------------|------------------|
|                                  |   |               | NPH <sup>1</sup>  | PHN <sup>2</sup> | BaP <sup>3</sup> |
| <i>ASMBC2</i> <sup>4</sup>       | Atlantic salmon exposed to Monterey Bay crude oil for 48 hours. | 06/22/16      | 160,000   | 53,000           | 1,500            |
|                                  | Atlantic salmon exposed to Monterey Bay crude oil for 48 hours. | 06/22/16      | 180,000   | 58,000           | 1,700            |
|                                  | Atlantic salmon exposed to Monterey Bay crude oil for 48 hours. | 06/22/16      | 180,000   | 56,000           | 1,700            |
|                                  | Atlantic salmon exposed to Monterey Bay crude oil for 48 hours. | 06/24/16      | 160,000   | 52,000           | 1,500            |
|                                  | Atlantic salmon exposed to Monterey Bay crude oil for 48 hours. | 06/24/16      | 160,000   | 52,000           | 1,500            |
|                                  | Atlantic salmon exposed to Monterey Bay crude oil for 48 hours. | 06/24/16      | 160,000   | 52,000           | 1,600            |
|                                  | Atlantic salmon exposed to Monterey Bay crude oil for 48 hours. | 06/25/16      | 160,000   | 52,000           | 1,600            |
|                                  | Atlantic salmon exposed to Monterey Bay crude oil for 48 hours. | 06/25/16      | 160,000   | 50,000           | 1,400            |
|                                  | Atlantic salmon exposed to Monterey Bay crude oil for 48 hours. | 06/25/16      | 160,000   | 51,000           | 1,500            |
| <i>Method blank</i> <sup>5</sup> | Methanol blank A  | 06/22/16      | 610   | 83               | 21               |
|                                  | Methanol blank C  | 06/22/16      | 400   | 54               | 26               |
|                                  | Methanol blank F  | 06/22/16      | 640   | 94               | 26               |
|                                  | Methanol blank A  | 06/24/16      | 600   | 100              | 29               |
|                                  | Methanol blank C  | 06/24/16      | 890   | 100              | 27               |
|                                  | Methanol blank F  | 06/24/16      | 810   | 90               | 26               |
|                                  | Methanol blank A  | 06/25/16      | 640   | 130              | 27               |
|                                  | Methanol blank C  | 06/25/16      | 780   | 91               | 27               |
|                                  | Methanol blank F  | 06/25/16      | 590   | 99               | 23               |

| Bile Reference Material ASMBC2 | Equivalents of fluorescent aromatic compounds (ng/g bile, wet weight) |                  |                  |
|--------------------------------|---|------------------|------------------|
|                                | NPH <sup>1</sup>  | PHN <sup>2</sup> | BaP <sup>3</sup> |
| Mean                           | 150,000   | 50,000           | 1,200            |
| SD                             | 14,000  | 4,900            | 230              |
| Upper Control Limit            | 180,000   | 60,000           | 1,700            |
| Lower Control Limit            | 120,000   | 40,000           | 740              |

<sup>1</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of naphthalene standard at 292/335 nm wavelengths.

<sup>2</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of phenanthrene standard at 260/380 nm wavelengths.

<sup>3</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of benzo[a]pyrene standard at 380/430 nm wavelengths.

<sup>4</sup>NWFSC Quality Assurance Criterion (from Sloan et al. 2006): Reference material (3/set): analyte concentrations will be  $\leq 2$  SD of historic values

<sup>5</sup>NWFSC Quality Assurance Criterion (from Sloan et al. 2006): Method blank (3/set): analyte concentrations in samples will be  $\geq 10$  times the maximum blank value.



## Appendix 2

**Table 3. Results of duplicate analyses<sup>1</sup> for metabolites of polycyclic aromatic compounds of selected bile samples of field captured fish analyzed one year after the 2015 Refugio Beach Oil Spill.**

| Species           | Laboratory ID number | Collection site | Field ID number | Protein mg/mL | Equivalents of fluorescent aromatic compounds (ng/g bile, wet weight) |                  |                  | Protein-corrected equivalents of fluorescent aromatic compounds (ng/mg biliary protein) |                  |                  |
|-------------------|----------------------|-----------------|-----------------|---------------|---|------------------|------------------|---|------------------|------------------|
|                   |                      |                 |                 |               | NPH <sup>2</sup>  | PHN <sup>3</sup> | BaP <sup>4</sup> | NPH <sup>2</sup>  | PHN <sup>3</sup> | BaP <sup>4</sup> |
|                   |                      |                 |                 |               | Barred surfperch  | 136-0444         | Gaviota          | GAVFI1051816BI5   | 4.5              | 200,000          |
|                   | 136-0444R            | Gaviota         | GAVFI1051816BI5 | 4.5           | 190,000   | 42,000           | 560              | 42,000  | 9,300            | 120              |
| Barred surfperch  | 136-0406             | Refugio Beach   | RSBFI1051816BI4 | 7.1           | 240,000   | 52,000           | 460              | 73,000  | 34,000           | 65               |
|                   | 136-0406R            | Refugio Beach   | RSBFI1051816BI4 | 7.1           | 230,000   | 51,000           | 480              | 72,000  | 32,000           | 68               |
| Walleye surfperch | 136-0425             | Campus Point    | CMPFI1051816BI7 | 15.3          | 1,100,000   | 290,000          | 2,400            | 72,000  | 19,000           | 160              |
|                   | 136-0425R            | Campus Point    | CMPFI1051816BI7 | 15.3          | 1,100,000   | 280,000          | 2,400            | 72,000  | 18,000           | 160              |

<sup>1</sup>NWFSC Quality Assurance Criterion (from Sloan et al. 2006): Sample duplicates (at least 1 for every 20 field samples analyzed): relative percent difference for each analyte  $\leq$  60% for duplicates.

<sup>2</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of naphthalene standard at 292/335 nm wavelengths.

<sup>3</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of phenanthrene standard at 260/380 nm wavelengths.

<sup>4</sup>Concentrations in part per billion (ng/g) based on total area compared to the fluorescence of benzo[a]pyrene standard at 380/430 nm wavelengths.