

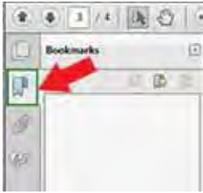
California Fish and Game Commission
Marine Resources Committee
Meeting Binder



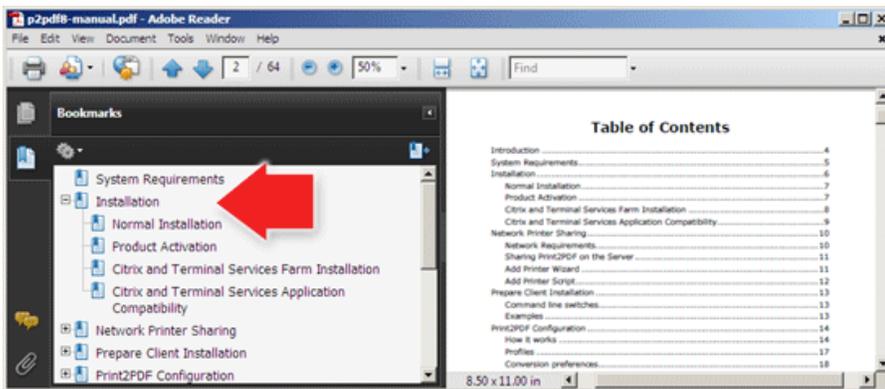
November 10, 2020
Webinar/Teleconference

EASY GUIDE TO USING THE BINDER

1. Download and open the binder document using your Adobe Acrobat program/app.
2. If a bookmark panel does not automatically appear on either the top or left side of the screen, click/tap on the “bookmark symbol” located near the top left-hand corner.



3. To make adjustments to the view, use the Page Display option in the View tab. You should see something like:



4. We suggest leaving open the bookmark panel to help you move efficiently among the staff summaries and numerous supporting documents in the binder. It's helpful to think of these bookmarks as a table of contents that allows you to go to specific points in the binder without having to scroll through hundreds of pages.
5. You can resize the two panels by placing your cursor in the dark, vertical line  located between the panels and using a long click /tap to move in either direction.
6. You may also adjust the sizing of the documents by adjusting the sizing preferences located on the Page Display icons found in the top toolbar or in the View tab.
7. Upon locating a staff summary for an agenda item, notice that you can obtain more information by clicking/tapping on any item underlined in blue.
8. Return to the staff summary by simply clicking/tapping on the item in the bookmark panel.
9. Do not hesitate to contact staff if you have any questions or would like assistance.

OVERVIEW OF FISH AND GAME COMMISSION COMMITTEE MEETING

- Welcome to this meeting of the Marine Resources Committee. The Committee is comprised of up to two Commissioners who co-chair each meeting; members are assigned annually by the Commission.
- Our goal today is informed discussion to guide future decision making, and, we need your cooperation to ensure a lively and comprehensive dialogue.
- We are operating under Bagley-Keene Open Meeting Act, but it is important to note that the Committee chairs cannot take action independent of the full Commission; instead, the chairs make recommendations to the full Commission at regularly scheduled meetings.
- These proceedings may be recorded and posted to our website for reference and archival purposes.
- Items may be heard in any order pursuant to the determination of the Committee co-chairs.
- In the unlikely event of an emergency, please locate the nearest emergency exits.
- As a general rule, requests for regulatory change need to be redirected to the full Commission and submitted on the required petition form, FGC 1, titled “Petition to the California Fish and Game Commission for Regulation Change” (Section 662, Title 14, CCR). However, at the Committee’s discretion, the Committee may request that staff follow up on items of potential interest to the Committee and possible recommendation to the Commission.
- Committee meetings operate informally and provide opportunity for everyone to provide comment on agenda items. If you wish to speak on an agenda item, please follow these guidelines:
 1. Raise your hand and wait to be recognized by the Committee.
 2. Provide your name, affiliation (if any), and the number of people you represent.
 3. Time is limited; please keep your comments precise to give others time to speak.
 4. If several speakers have the same concerns, please appoint a group spokesperson.
 5. If speaking during general public comment, the subject matter you present should not be related to any item on the current agenda (public comment on agenda items will be taken at the time the Committee discusses that item).

INTRODUCTIONS FOR FISH AND GAME COMMISSION MARINE RESOURCES COMMITTEE

FISH AND GAME COMMISSIONERS

Peter Silva Co-Chair (Jamul)
Samantha Murray Co-Chair (Del Mar)

COMMISSION STAFF

Melissa Miller-Henson Executive Director
Rachel Ballanti Deputy Executive Director
Susan Ashcraft Marine Advisor
Sherrie Fonbuena Program Analyst
Rose Dodgen Sea Grant State Fellow

DEPARTMENT OF FISH AND WILDLIFE

Mike Stefanik Assistant Chief, Law Enforcement Division
Randy Lovell Statewide Aquaculture Coordinator
Craig Shuman Regional Manager, Marine Region
Sonke Mastrup Program Manager, Invertebrate Fisheries, Marine Region
Becky Ota Program Manager, Habitat Conservation, Marine Region
Kirsten Ramey Program Manager, State Fisheries, Marine Region
Tom Mason Senior Environmental Scientist Supervisor, Marine Region
Steve Wertz Senior Environmental Scientist Supervisor, Marine Region
Debbie Aseltine-Neilson Senior Environmental Scientist Specialist, Marine Region
Armand Barilotti Environmental Scientist, Marine Region
Amanda Van Diggelen Environmental Scientist, Marine Region

INVITED SPEAKERS

Jenn Eckerle Deputy Director, California Ocean Protection Council

I would also like to acknowledge special guests who are present:
(i.e., key DFW staff, elected officials, tribal chairpersons, other special guests)

Commissioners
Eric Sklar, President
Saint Helena

Samantha Murray, Vice President
Del Mar

Jacque Hostler-Carmesin, Member
McKinleyville

Peter S. Silva, Member
Jamul

Vacant, Member

STATE OF CALIFORNIA
Gavin Newsom, Governor

Fish and Game Commission



*Celebrating 150 Years of
Wildlife Heritage and Conservation!*

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fgc@fgc.ca.gov
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MARINE RESOURCES COMMITTEE

Committee co-chairs: Commissioner Silva and Commissioner Murray

Meeting Agenda

November 10, 2020; 9:00 a.m.

Webinar / Teleconference

The California Fish and Game Commission is conducting this committee meeting by webinar and teleconference to avoid a public gathering and protect public health during the COVID-19 pandemic, consistent with Executive Order N-33-20.

Pursuant to Executive Order N-29-20, members may participate in meetings remotely. The public may provide public comment during the public comment periods, and otherwise observe remotely consistent with the Bagley-Keene Open Meeting Act.

To participate in the meeting, please join via Zoom or by telephone.

Please click here or go to <http://nrm.dfg.ca.gov/FileHandler.ashx?DocumentID=183745&inline> for instructions on how to join the meeting.

Note: See important meeting deadlines and procedures, including written public comment deadlines, starting on page 5. Unless otherwise indicated, the California Department of Fish and Wildlife is identified as Department. All agenda items are informational and/or discussion only; the Committee develops recommendations to the Commission but does not have authority to make policy or regulatory decisions on behalf of the Commission.

Call to order

1. Approve agenda and order of items

2. General public comment for items not on agenda

The Committee may not discuss or take action on any matter raised during this item, except to consider whether to recommend that the matter be added to the agenda of a future meeting [Sections 11125, 11125.7(a), Government Code].

3. Recreational California grunion

Discuss and consider potential committee recommendation on proposed regulations for the California grunion recreational fishery.

4. Existing structures in marine protected areas

Receive Department update and consider potential committee recommendation on potential regulation to allow for the operation, maintenance, and repair of existing artificial structures in marine protected areas.

5. New marine aquaculture leases in California

Discuss and consider potential committee recommendation to continue a temporary hiatus on receipt of new applications for state water bottom leases for the purpose of aquaculture (excepting previously received applications currently under consideration).

6. Non-native invasive marine kelp and algae species

Discuss concerns related to the spread and control of marine invasive species, including *Sargassum horneri*, and possible ecological, management, and policy considerations associated with intervention.

7. Marine Life Management Act master plan implementation

Receive Department update on implementation efforts for the 2018 master plan for fisheries.

- (A) Review of California halibut fishery management
- (B) Invertebrate fisheries prioritization
- (C) Potential commercial pink shrimp trawl fishery management plan

8. Staff and agency updates requested by the Committee

Receive updates from staff and other agencies related to topics for which the Committee has requested an update.

Note: To enhance meeting efficiency in the webinar/teleconference format, the Committee intends to receive updates primarily in writing. The public will be given an opportunity to provide comments, although the level of in-meeting discussion will be at the discretion of the Committee.

- (A) California Ocean Protection Council
 - I. Aquaculture principles and action plan development update
- (B) Department
 - I. Update on recreational red abalone fishery management plan development
 - II. Update on developing proposed regulations governing commercial kelp and algae harvest, including outreach efforts with affected industry members and interested parties.
 - III. Update on the Pacific herring quota under the new fishery management plan
- (C) Commission staff: Update on Coastal Fishing Communities Project

9. Future agenda items

- (A) Review work plan agenda topics, priorities, and timeline
- (B) Potential new agenda topics for Commission consideration

Adjourn

California Fish and Game Commission Meeting Schedule

Note: As meeting dates and locations can change, please visit www.fgc.ca.gov for the most current list of meeting dates and locations.

Meeting Date	Commission Meeting	Committee Meeting
November 16, 2020		Tribal Webinar/teleconference
December 9-10, 2020	Webinar/teleconference	
January 12, 2021	Webinar/teleconference	
January 12, 2021		Wildlife Resources Webinar/teleconference
February 10-11, 2021	Webinar/teleconference	
March 16, 2021		Marine Resources Webinar/teleconference
April 13, 2021		Tribal Webinar/teleconference
April 14-15, 2021	Webinar/teleconference	
May 11, 2021		Wildlife Resources Webinar/teleconference
May 11, 2021	Webinar/teleconference	
June 16-17, 2021	Webinar/teleconference	
July 20, 2021		Marine Resources Sacramento
August 17, 2021		Tribal Sacramento
August 18-19, 2021	Sacramento	
September 16, 2021		Wildlife Resources Sacramento
October 13-14, 2021	Sacramento	
November 9, 2021		Marine Resources Sacramento
December 14, 2021		Tribal Sacramento
December 15-16, 2021	Sacramento	

OTHER MEETINGS OF INTEREST

Association of Fish and Wildlife Agencies

- September 12-15, 2021, Providence, RI

Pacific Fishery Management Council

- November 13-20, 2020, Garden Grove, CA
- March 3-10, 2021, Seattle, WA
- April 6-13, 2021, San Jose, CA
- June 22-29, 2021, Vancouver, WA
- September 8-15, 2021, Spokane, WA
- November 15-22, 2021, Costa Mesa, CA

Pacific Flyway Council

- March 9, 2021, Grand Rapids, MI
- August or September 2021, TBD

Western Association of Fish and Wildlife Agencies

- January 7-10, 2021, Santa Ana Pueblo, NM
- July 18-23, 2021 Santa Fe, NM

Wildlife Conservation Board

- November 18, 2020, Sacramento, CA
- 2021 TBD

Important Committee Meeting Procedures Information

Welcome to a meeting of the California Fish and Game Commission's Marine Resources Committee. The Committee is chaired by up to two Commissioners; these assignments are made by the Commission.

The goal of the Committee is to allow greater time to investigate issues before the Commission than would otherwise be possible. Committee meetings are less formal in nature and provide for additional access to the Commission. The Committee follows the noticing requirements of the Bagley-Keene Open Meeting Act. It is important to note that the Committee chairs cannot take action independent of the full Commission; instead, the chairs make recommendations to the full Commission at regularly scheduled meetings.

The Commission's goal is the preservation of our heritage and conservation of our natural resources through informed decision-making; Committee meetings are vital in developing recommendations to help the Commission achieve that goal. In that spirit, we provide the following information to be as effective and efficient toward that end. Welcome, and please let us know if you have any questions.

PERSONS WITH DISABILITIES

Persons with disabilities needing reasonable accommodation to participate in public meetings or other Commission activities are invited to contact the Reasonable Accommodation Coordinator at (916) 651-1214. Requests for facility and/or meeting accessibility should be received at least 10 working days prior to the meeting to ensure the request can be accommodated.

SUBMITTING WRITTEN MATERIALS

The public is encouraged to attend Committee meetings and engage in the discussion about items on the agenda; the public is also welcome to comment on agenda items in writing. You may submit your written comments by one of the following methods (only one is necessary): **Email** to fgc@fgc.ca.gov; **mail** to California Fish and Game Commission, P.O. Box 944209, Sacramento, CA 94244-2090; or **deliver** to California Fish and Game Commission, 1416 Ninth Street, Suite 1320, Sacramento, CA 95814.

COMMENT DEADLINES

The **Written Comment Deadline** for this meeting is **5:00 p.m. on October 28, 2020**. Written comments received at the Commission office by this deadline will be made available to Commissioners prior to the meeting.

The **Supplemental Comment Deadline** for this meeting is **noon on November 5, 2020**. Written comments received by this deadline will be made available to Commissioners at the meeting.

The Committee will not consider comments regarding proposed changes to regulations that have been noticed by the Commission. If you wish to provide comment on a noticed item, please provide your comments during Commission business meetings, via email, or deliver to the Commission office.

Note: Materials provided to the Committee may be made available to the general public.

PETITIONS FOR REGULATION CHANGE

As a general rule, requests for regulatory change need to be redirected to the full Commission and submitted on the required petition form, FGC 1, *Petition to the California Fish and Game Commission for Regulation Change* (Section 662, Title 14, California Code of Regulations). However, at the Committee's discretion, the Committee may request that staff follow up on items of potential interest to the Committee and possible recommendation to the Commission.

SPEAKING AT THE MEETING

1. Committee meetings operate informally and provide opportunity for everyone to comment on agenda items. If you wish to speak on an agenda item, please follow these guidelines:
2. Raise your hand and wait to be recognized by the Committee co-chair(s). You will raise your hand via the "hand raise" button on Zoom or by pressing "#2" if you are on the phone.
3. Once recognized, please begin by giving your name and affiliation (if any) and the number of people you represent.
4. Time is limited; please keep your comments concise so that everyone has an opportunity to speak.
5. If there are several speakers with the same concerns, please try to appoint a spokesperson and avoid repetitive comments.
6. If speaking during general public comment, the subject matter you present should not be related to any item on the current agenda (public comment on agenda items will be taken at the time the Committee members discuss that item). As a general rule, general public comment is an opportunity to bring matters to the attention of the Committee, but you may also do so via email or standard mail. At the discretion of the Committee, staff may be requested to follow up on the subject you raise.

VISUAL PRESENTATIONS/MATERIALS

All electronic presentations must be submitted by the **Supplemental Comment Deadline** and approved by the Commission executive director before the meeting.

1. Electronic presentations must be provided by email to fgc@fgc.ca.gov or delivered to the Commission on a USB flash drive by the deadline.
2. All electronic formats must be Windows PC compatible.

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

2. GENERAL PUBLIC COMMENT**Today's Item****Information** **Action**

Receive public comment for items not on the agenda.

Summary of Previous/Future Actions (N/A)**Background**

MRC receives two types of correspondence or comment under general public comment: requests for MRC to consider new topics and informational items. As a general rule, requests for regulatory change must be submitted to FGC on petition form FGC 1, *Petition to the California Fish and Game Commission for Regulation Change* (Section 662). However, MRC may, at its discretion, request that staff follow up on items of potential interest for possible recommendation to FGC.

Significant Public Comments (N/A)**Recommendation**

Staff recommends any new agenda items based on issues raised and within FGC's authority be held for discussion under Agenda Item 9, Future Agenda Items.

Exhibits (N/A)**Committee Direction/Recommendation (N/A)**

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

3. RECREATIONAL CALIFORNIA GRUNION**Today's Item**Information Action

Discuss and consider potential committee recommendation on proposed regulations for the California grunion recreational fishery.

Summary of Previous/Future Actions

- FGC granted regulation change petition #2019-014 Feb 21, 2020; Sacramento
- MRC discussed potential management measures Jul 29, 2020; MRC, Webinar/Teleconference
- **Today's discussion** **Nov 10, 2020; MRC, Webinar/Teleconference**

Background

California grunion is known primarily for its unique spawning behavior, referred to as “grunion runs”, along southern California beaches on predictable nights of the year. Grunion may be harvested recreationally from Jun 1 through Mar 31 under current regulations.

In Feb 2020, FGC granted a petition to amend recreational take regulations for California grunion to be more conservative, and requested that DFW develop specific proposed changes upon completing an enhanced status report (ESR) for the species. At the Jul 2020 MRC meeting, DFW provided a written update (Exhibit 1), reporting that it completed the grunion ESR in May (available in the California Marine Species Portal at <https://marinespecies.wildlife.ca.gov/california-grunion/>) and, consistent with its findings, was developing potential regulation changes as requested by FGC, commencing with an online public survey and tribal outreach. Today, DFW will present specific potential regulation changes for MRC consideration and potential recommendation (Exhibit 2).

Significant Public Comments

The petitioner has offered to continue to support this rulemaking effort in any way possible.

Recommendation

FGC staff: Support the proposed management measures in a rulemaking as recommended by DFW under a timeline to be determined contingent upon regulatory staff capacity.

DFW: Advance a rulemaking to amend recreational take regulations for California grunion, to include: add a bag and possession limit of between 10 and 20 fish; and reduce the fishing season by one month, leading to a revised open season of Jul 1–Mar 31.

Exhibits

1. [DFW written update on California grunion](#), received Jul 13, 2020
2. [DFW presentation](#)

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

Committee Direction/Recommendation

The Marine Resources Committee recommends that the Commission advance a rulemaking with the proposed management measures for the California grunion recreational fishery as recommended by the California Department of Fish and Wildlife on a timeline to be determined.

OR

The Marine Resources Committee recommends that the Commission advance a rulemaking with the proposed management measures for the California grunion recreational fishery as recommended by the California Department of Fish and Wildlife, except _____, on a timeline to be determined.

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

4. EXISTING STRUCTURES IN MARINE PROTECTED AREAS**Today's Item**Information Action

Receive DFW update and consider a potential committee recommendation on a rulemaking to allow for operation, maintenance, and repair of existing artificial structures in marine protected areas (MPAs).

Summary of Previous/Future Action

- | | |
|--|--|
| • FGC referred topic to MRC | Jun 24-25, 2020; Webinar/Teleconference |
| • DFW presented overview to MRC | Jul 21, 2020; MRC, Webinar/Teleconference |
| • Today's discussion and potential recommendation | Nov 10, 2020; MRC, Webinar/Teleconference |

Background

At its Jun 2020 meeting, FGC referred to MRC an emerging management issue related to MPAs: the operation, maintenance and repair of artificial structures that were installed under permits issued by federal, state, or local agencies prior to MPA designation. Additionally, some structures must be replaced or removed. The required operation and maintenance, or repair, replacement or removal of artificial structures may result in incidental injury, damage, take or possession of living, geological or cultural resources that are otherwise protected.

Operation, maintenance and repair of existing artificial structures was identified as an issue during the south coast regional MPA planning process, and regulations for specific south coast MPAs were written to allow for these activities. However, artificial structures within other MPAs throughout the state also require operation, maintenance and repair activities not explicitly authorized in MPA regulations, resulting in an unintended constraint for agencies with jurisdiction over these structures. The MPA Statewide Leadership Team (leadership team) has discussed the need for resolution.

At the Jul 2020 MRC meeting, DFW presented an overview and a general approach to addressing the issue. Additionally, DFW has discussed with the leadership team a potential regulatory pathway to allow ongoing operation and maintenance, or repair, replacement or removal of artificial structures installed prior to MPA designation. Permitting agencies on the leadership team support revising regulations to address this issue in light of immediate maintenance and repair needs currently under agency review. As operation and maintenance, and repair, replacement or removal activities necessarily involve the area immediately adjacent to artificial structures, the permitting agencies are providing input on what distance would accommodate the activities, to be defined as a "buffer zone."

Today, DFW will present an overview of the proposed regulatory pathway for MRC review and potential recommendation to FGC.

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

Significant Public Comments (N/A)**Recommendation**

FGC staff: Support (1) the proposed regulatory pathway as recommended by DFW for consideration in a rulemaking under a timeline to be determined contingent upon regulatory staff capacity, (2) request that DFW develop options for a buffer zone size in consultation with the leadership team in advance of the rulemaking, and (3) present a recommendation to FGC for consideration.

DFW: Advance a rulemaking to: (1) Add a definition for pre-existing artificial structure where “pre-existing” means prior to specified implementation dates for MPAs, and (2) authorize operation, maintenance, repair, replacement, and removal of pre-existing artificial structures within a defined buffer zone around the structure (with zone size to be determined). Suggested timeline is to authorize notice in Feb 2021.

Exhibits

1. [DFW presentation](#)

Committee Direction/Recommendation

The Marine Resources Committee recommends that the Commission support the proposed regulatory pathway to allow ongoing operation, maintenance and repair of artificial structures that were installed prior to marine protected areas being designated, as recommended by the California Department of Fish and Wildlife, and request that the Department develop options for a buffer zone size in consultation with the Marine Protected Area Statewide Leadership Team.

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

5. NEW MARINE AQUACULTURE LEASES**Today's Item**Information Action

Discuss and consider potential MRC recommendation regarding a temporary hiatus on receipt of new applications for state water bottom leases for the purpose of aquaculture (excepting previously-received applications currently under consideration).

Summary of Previous/Future Actions

- | | |
|--|--|
| • FGC referred discussion of potential temporary hiatus on new lease applications to MRC | Feb 21, 2020; Sacramento |
| • MRC discussion and recommendation for six-month hiatus on new lease applications | Apr 29, 2020; MRC (part 2), webinar/teleconference |
| • FGC approved MRC recommendation for six-month hiatus on new lease applications | Jun 24-25, 2020; webinar/teleconference |
| • MRC review of hiatus and potential recommendation | Nov 10, 2020; MRC, webinar/teleconference |

Background

In Feb 2020, FGC referred to MRC discussion about a potential temporary hiatus in considering new state water bottom lease applications, excluding the applications already received (two proposed offshore sites in Southern California, and one proposed site in Tomales Bay). With the exception of Santa Barbara Mariculture, where reconfiguration of its existing lease was administered as a new lease application for purposes of the California Environmental Quality Act (CEQA), the three applications are the first for new lease areas that FGC has received in over 25 years; much has changed in the subsequent years and the methods and processes for reviewing leases have had to be created anew.

At the Apr 29 MRC meeting, FGC staff highlighted the need to establish an administrative process and standards to guide FGC review and consideration of new lease applications, further develop coordination protocols between FGC and DFW staff, more clearly articulate staff roles and responsibilities, and refine practices for communicating expectations with lease applicants. Available FGC and DFW staff resources were identified as a particular concern; staff is responsible for managing 17 existing leases that must necessarily take priority, in addition to processing the three lease applications already under consideration, before it can consider undertaking additional new lease reviews.

MRC recommended, and FGC approved at its Jun 24-25, 2020 meeting, a six-month hiatus on accepting any new state water bottom lease applications for aquaculture purposes; the approved hiatus is slated to expire Dec 24, 2020.

Update

Marine aquaculture is an adapting and growing industry, with increased interest in supporting locally-grown seafood. Optimally, decisions regarding prospective new lease applications would be made within the context of a broader policy and vision, in addition to the enhanced

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

administrative process being developed. FGC staff is participating in an effort led by the California Ocean Protection Council (OPC) to develop statewide aquaculture principles and a statewide aquaculture action plan, recognizing the need to have a common vision among the multiple state agencies of jurisdiction and to more efficiently and effectively coordinate the resources currently allocated to permitting and managing aquaculture in California. OPC's effort is likely to identify the need for additional state support if the state's goal is to increase sustainable aquaculture.

Specific to the approved hiatus, FGC and DFW staff has made progress in administrative coordination of application review, clarifying respective roles, advancing environmental review under CEQA for one application, and improving coordination with other agencies of jurisdiction (there are at minimum seven, and usually more depending on the project). Additional progress is still needed to support a consistent review process for new lease applications, especially with regard to meeting CEQA requirements. Meeting the review and coordination requirements in a time frame preferred by applicants will continue to be a challenge.

Concurrent to the existing lease application review processes, staff is also focused on responding to requests from several existing lessees for lease amendments, transfers, or other remedies related to authorized culture species, culture methods, lease boundaries and/or operations. Some requests are discretionary; however, for the majority of the current requests, the principle driver is the need to comply with new conditions established through other agency permitting processes that are raising questions and concerns not previously identified or addressed. The current requests from existing lessees have not been simple and have required research, interagency consultation, and environmental review.

Staff recognizes that continuing the hiatus on any new lease applications will not serve to remedy the challenges facing FGC and DFW staff in the receipt and review of lease applications; therefore, staff is not requesting a continuation of the hiatus. However, staff anticipates that OPC's effort to develop statewide aquaculture principles will contribute to articulating a vision and framework that will support how FGC reviews and considers aquaculture lease applications while a statewide aquaculture action plan is being developed. Based upon initial conversations, staff believes the principles will be consistent with concepts and values that FGC has previously expressed regarding aquaculture in California.

Unless directed otherwise, staff will prioritize existing lessee requests first, followed by the three lease applications already under consideration before initiating a review process for any new applications that may be received in the future.

Significant Public Comments

An aquaculture leaseholder operating offshore from Santa Barbara urges that FGC not approve any new state water bottom leases until a clear vision is defined and comprehensive management program for implementing new leases developed, including the applications already received by FGC. Offers specific recommendations related to leveraging the capacity of other organizations, supporting training and internship opportunities, setting more stringent experience and qualification requirements, and authorizing complementary rather than competing culture operations where available sites are constrained (Exhibit 1).

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

Recommendation

FGC staff: Allow the current hiatus on receipt of new lease applications to lapse, recognizing the limitations in staff and resources; direct staff to continue developing and refining review processes with DFW and other agencies of jurisdiction; schedule an update related to aquaculture principles and action plan details for the Mar 2020 MRC meeting, and schedule an update on aquaculture leases for a future MRC meeting.

Exhibits

1. [Background document: Staff summary for Mar 17, 2020 MRC meeting](#), Agenda Item 7
2. [Email from Bernard Friedman, Santa Barbara Mariculture Company](#), received Oct 27, 2020

Committee Direction/Recommendation (N/A)

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

6. NON-NATIVE INVASIVE MARINE KELP AND ALGAE

Today's Item	Information <input checked="" type="checkbox"/>	Action <input type="checkbox"/>
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Discuss concerns related to the spread and control of marine invasive species, including *Sargassum horneri*, and possible ecological, management, and policy considerations associated with intervention.

Summary of Previous/Future Actions

- | | | |
|-----------------------------|--|--|
| • FGC referred topic to MRC | Oct 14, 2020; Webinar/Teleconference | |
| • Today's discussion | Nov 10, 2020; MRC, Webinar/Teleconference | |

Background

Invasive marine species have been an environmental issue along the California coast for decades, and a variety of strategies for their removal have been explored with various goals for what removal will achieve. Interest in removing specific species to support restoration of marine ecosystems, such as kelp forests, has increasingly been brought into the FGC arena.

At the Jul 2020 MRC meeting, several public commenters raised concerns about *Sargassum horneri*, a large, invasive, non-native, algal species that can compete with native kelp. The species has colonized in multiple southern California locations, although it has recently been observed in locations northward along California's coast. Commenters proposed solutions, and requested that FGC and DFW implement a broad strategy to address the spread of *Sargassum horneri* and remove regulatory or policy barriers to public engagement in the strategy, including within marine protected areas (MPAs).

In Oct 2020, FGC received three requests related to *Sargassum horneri*:

1. A petition for regulation change (Petition #2020-014 AM 1; Exhibit 1) to authorize recreational and research removal, including inside specified MPAs);
2. a petition to amend California Fish and Game Code to add *Sargassum horneri* and an invasive bryozoan to the list of aquatic invasive species authorized to be removed for scientific research (since withdrawn as the request is outside of FGC authority); and
3. an application to commercially harvest *Sargassum horneri* under current commercial kelp regulations as a financial incentive to remove the invasive species (Exhibit 2).

After extensive public comment at the Oct 2020 meeting urging FGC and DFW to approve the actions proposed to eliminate *Sargassum horneri* generally, and within MPAs specifically, FGC referred the topic of invasive, non-native kelp and algae to MRC for discussion. Since the meeting, staff from FGC, DFW and the California Ocean Protection Council (OPC) met to discuss policy considerations associated with intervention, restoration, MPAs, and mitigation, and agreed to further explore these issues. Some of the factors that should be considered when exploring potential intervention, in no particular order, include:

- Ecosystem context,
- ecological risk and uncertainty;

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

- policy and management context;
- best available science and knowledge gaps, including efficacy of removal methods;
- legal constraints;
- potential unintended consequences; and
- state priorities.

Today's meeting presents an opportunity to explore the full breadth of concerns associated with the spread of invasive, non-native kelp and algae species, and discuss ecological, management, and policy considerations associated with intervention and restoration. DFW and OPC staff will provide background and context associated with *Sargassum horneri* specifically, including recent science (see exhibits 3 and 4 as examples), and policy considerations associated with non-native species removal more generally.

Significant Public Comments

Ten comments in support of Petition #2020-014 were received after the comment deadline but will be provided with supplemental handouts. Commenters urged FGC and DFW to authorize removing *Sargassum horneri* from MPAs so that MPAs do not protect species that were not intended to be protected.

Recommendation

FGC Staff: Solicit input from stakeholders and DFW, and identify any desired follow-up.

Exhibits

1. [Petition #2020-014 AM 1: Recreational and research take of *Sargassum horneri*](#), received Sep 19, 2020.
2. [Email from Jeff Maassen and application for commercial harvest of *Sargassum horneri*](#), received Oct 1, 2020
3. [Research article: Assessment of control methods for the invasive seaweed *Sargassum horneri* in California, USA](#), by Marks *et al.*, special issue of *Management of Biological Invasions*, Jun 2017
4. [Research article: Impacts of the non-native alga *Sargassum horneri* on benthic community production in a California kelp forest](#), by Sullaway and Edwards, *Marine Ecology Progress Series*, Mar 5, 2020 (posted with author permission)

Committee Direction/Recommendation (N/A)

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

7. Marine Life Management Act (MLMA) Master Plan Implementation**Today's Item****Information** ☒**Action** ☒

Receive DFW update on next implementation efforts for the 2018 master plan for fisheries.

Summary of Previous/Future Actions

- | | |
|---|--|
| • FGC approved MRC recommendation to develop a rulemaking for pink shrimp trawl fishery regulations | Dec 6-7, 2017, San Diego |
| • FGC adopted 2018 MLMA master plan for fisheries | Jun 20-21, 2018; Sacramento |
| • DFW updates on MLMA master plan implementation | 2019-2020, MRC, various |
| • Today's update and discussion | Nov 10, 2020; MRC, webinar/teleconference |

Background

This is a standing agenda item for MRC to track progress on actions related to the *2018 Master Plan for Fisheries: A Guide for Implementation of the Marine Life Management Act* (MLMA Master Plan), which was adopted by FGC and is being implemented by DFW as a framework for fisheries management.

Consistent with California Fish and Game Code subdivision 7073(b)(2), the MLMA Master Plan calls for creating a priority list for developing fisheries management plans (FMPs) and other scaled management efforts based on the prioritization framework established through the plan. Fisheries that DFW determines have the greatest need for changes in conservation and management measures to comply with MLMA policies and principles are given highest priority.

DFW prepared an interim prioritization list in 2018 for 45 state-managed fisheries; in Nov 2019, DFW presented MRC with an updated prioritization list for key finfish fisheries and highlighted progress on integrating invertebrate species. In Feb 2020, DFW presented FGC with an up-to-date implementation work plan, including an updated priority list that now includes several invertebrate species. Today, DFW's update will focus on three topics:

1. *Review of California halibut fishery management*: The most recent DFW update was provided to MRC on Apr 29.
2. *Invertebrate fisheries prioritization*: A DFW presentation is provided in Exhibit 1.
3. *Potential pink shrimp FMP* (Exhibit 2): This is the first MRC update since a Nov 2017 discussion about management, permit capacity, and research needs for the commercial pink shrimp trawl fishery, after which FGC approved an MRC recommendation to address management concerns in the existing fishery through regulation changes, on a timeline to be determined. DFW staff has continued to develop management options for pink shrimp, shaped by the MLMA Master Plan scaled management framework, and today will present its proposal to develop an FMP

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

and implementing regulations as the preferred way to address pink shrimp management needs.

Significant Public Comments

1. The Crescent City Harbor District Board of Harbor Commissioners requests that FGC expedite review and approval of a pink shrimp FMP, as it is the final step for the fishery to obtain Marine Stewardship Council recognition as a certified sustainable fishery. The board describes the financial and time burdens currently faced by commercial fishermen who offload a large percentage of California-caught pink shrimp out of state due to California's lack of council certification (Exhibit 3).

Recommendation

FGC staff: Provide guidance regarding the next MLMA Master Plan implementation steps identified by DFW; and recommend that FGC support development of an FMP for pink shrimp as proposed and discussed today.

DFW: Support development of a pink shrimp FMP through the process presented at this meeting, including MRC vetting, outreach to tribes, and environmental review under the California Environmental Quality Act.

Exhibits

1. [DFW presentation on invertebrate fisheries prioritization](#)
2. [DFW presentation on potential pink shrimp FMP](#)
3. [Letter from Crescent City Harbor District Board of Harbor Commissioners](#), received Oct 13, 2020

Committee Direction/Recommendation

The Marine Resources Committee recommends that the Commission support development of a fishery management plan for California pink shrimp as recommended by the California Department of Fish and Wildlife, and add California pink shrimp FMP to the committee work plan.

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

8. STAFF AND AGENCY UPDATES**Today's Item**Information Action

Receive updates from staff and other agencies related to topics for which the Committee has requested an update.

Summary of Previous/Future Actions (N/A)**Background**

This is a standing item for DFW and other government agencies to provide an update on marine-related activities of interest.

(A) California Ocean Protection Council (OPC)*Aquaculture principles and action plan development*

In Jul 2020, OPC provided a written update (Exhibit 1) highlighting that it would convene state agency leadership to develop shared marine aquaculture principles to inform development of a more comprehensive statewide aquaculture action plan. Today, OPC staff will provide a verbal update on efforts to develop aquaculture principles and possible next steps.

(B) DFW**I. Update on recreational red abalone fishery management plan (FMP) development**

At its Aug 2020 meeting, FGC approved MRC's recommendation to support DFW developing for further MRC and public review a draft FMP to include all FMP elements recommended in the [harvest control rule integration] administrative team report, including both spawning potential ratio and density metrics, two fishing management zones with a framework for a third zone, a de minimis fishery option with biological fishery provision, and potential tribal allocation. Today, DFW will provide a verbal update on FMP drafting efforts.

II. Update on developing proposed regulations for commercial kelp and algae harvest

In Mar 2020, DFW provided MRC an overview of its initial draft proposed regulation changes for commercial kelp and algae harvest (see Exhibit 2 for background). Based on feedback from commercial kelp harvesters, MRC recommended, and in Apr 2020 FGC approved, (1) requesting that DFW conduct additional outreach with affected industry members, tribes and other interested parties, and (2) continuing the item to a future MRC meeting. In response, DFW hosted two outreach meetings via webinar on May 20 and Jun 2, 2020. Today, DFW will provide a verbal update on the outcomes of the outreach and next steps in developing a regulatory proposal. Note that at the Jun 2020 FGC meeting, the proposed schedule for this rulemaking was shifted to "to be determined" and the MRC update was continued to today's meeting.

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

III. *Update on the Pacific herring quota under the new fishery management plan*

The 2020-21 commercial fishing season is the first to operate under the California Pacific Herring FMP and implementing regulations adopted by FGC in Dec 2019. DFW's director has set the season quotas for all areas and sectors according to the FMP management strategy and in consideration of recommendations from the DFW Director's Herring Advisory Committee. A letter from Director Bonham (Exhibit 3), sent to notify permit holders of the season quotas, includes a table listing the management tier assignment and resulting quota for each management area and fishing sector as prescribed in the FMP. Of note, the San Francisco Bay herring biomass was below the cut-off level in the FMP, resulting in a zero-ton quota for the 2020-21 season. Today, DFW will present a fishery update, including challenges with conducting the biomass survey during the up-coming spawning season.

(C) **FGC staff***Coastal Fishing Communities Project update*

The MRC Coastal Fishing Communities project has been underway since 2015. In Dec 2019, FGC adopted the final *Staff Synthesis Report on California Coastal Fishing Communities Meetings, 2016-2018* (available on the project website at <https://fgc.ca.gov/Committees/Marine/Coastal-Fishing-Communities-Project>). The report proposed ten staff recommendations (SRs) as "initial concepts for potential development". MRC directed staff to further develop the SRs to help evaluate and prioritize the recommendations FGC may choose to act upon.

At the Jul 2020 MRC meeting, staff provided an overview of efforts to develop update reports for each of the ten SRs, and presented a sample for the first SR. Staff also proposed a draft analytical approach for a more in-depth analysis of each SR. MRC approved moving forward with developing the two products further.

Staff has completed update reports for each of the SRs (compiled as Exhibit 4). The update reports record efforts made by staff, collaborators, and/or external entities that are relevant to each SR as a starting point for the analytical evaluation. In addition, drawing from the framework presented in Jul 2020, staff conducted an analysis of SR 1 (Exhibit 5); SR 1 suggests developing and adopting an FGC policy and definition for coastal fishing communities (Exhibit 5).

Today, staff will present outcomes of the SR 1 analysis and suggest possible next steps for MRC discussion and direction to staff. In particular, the SR 1 analysis indicates that an FGC policy on coastal fishing communities has the potential to affect approaches to the other SRs (Exhibit 6). Therefore, at MRC's direction, staff could work with stakeholders to explore SR 1 while the other SRs are being further analyzed and return to the next MRC meeting in Mar 2021 with a recommendation for SR 1.

Significant Public Comments

Two retired DFW abalone project scientists provide substantial input and data regarding lessons learned from the abalone fishery collapse in southern California, and guidance for

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

developing a red abalone FMP for northern California under a collapsed fishery condition (exhibits 7-9).

Recommendation

FGC staff: For the Coastal Fishing Communities Project, direct staff to continue developing analyses for SRs 2–10 and, consistent with the SR 1 analysis, direct staff to explore a potential policy in consultation with stakeholders and return to the Mar 2021 MRC meeting with a recommendation.

Exhibits

1. [OPC agency update, received Jul 16, 2020](#) (for background purposes)
2. [Committee Staff Summary for Agenda Item 6 regarding kelp and algae, Mar 2020](#) (for background purposes)
3. [Sample letter from DFW Director Charlton Bonham to all herring commercial permit holders](#) informing them of permit quotas for 2020-21, dated Oct 30, 2020
4. [FGC staff report with updates for coastal fishing community staff recommendations](#), dated Nov 4, 2020
5. [FGC staff analysis of SR 1 from the 2019 staff synthesis report](#), dated Nov 4, 2020
6. FGC staff presentation for coastal fishing communities project (*will be provided at meeting*)
7. [Email from Konstantin Karpov](#), received Oct 9, 2020
8. [Email from Peter Haaker](#), received Oct 13, 2020
9. [Email and attached letter from Konstantin Karpov](#), received Oct 22, 2020

Committee Recommendation (N/A)

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

9. FUTURE AGENDA ITEMS**Today's Item**Information Action

Review upcoming agenda items scheduled for the next and future MRC meetings, discuss priorities and timeline, and consider requests for new agenda items.

Summary of Previous/Future Actions

- | | |
|-----------------------------|--|
| • FGC approved MRC agenda | Oct 14, 2020; Webinar/Teleconference |
| • Today's discussion | Nov 10, 2020; MRC, Webinar/Teleconference |
| • Next MRC meeting | Mar 16, 2021; MRC, Webinar/Teleconference |

Background

Committee topics are referred by FGC and scheduled as appropriate. FGC-referred topics and their current schedule are shown in the MRC work plan (Exhibit 1), and currently include several complex and time-intensive topics under development. MRC has placed emphasis on issues of imminent regulatory or management importance; thus, scheduling current topics and considering new topics for MRC review will require planning relative to existing workload and timing considerations.

MRC Work Plan and Timeline

In addition to standing agency and staff updates, four draft agenda topics are identified for the Mar 2021 MRC meeting and are grouped by the type of anticipated action to help inform workload and prioritization, if needed.

Updates

1. MLMA master plan for fisheries implementation
2. Red abalone FMP development
3. Kelp restoration and recovery tracking

Discussion and Potential Recommendations

4. Coastal Fishing Communities Project (*pending MRC direction under Agenda Item 8(C), this meeting*)

There are several referred topics on the work plan that could be added to the list of scheduled topics for the Mar 2021 MRC meeting, depending on today's discussions, topic readiness, and prioritization. Capacity to add topics is in part dependent on whether topics will be added as updates or for discussion and/or recommendation, with the latter typically requiring significantly more staff preparation. In the case of item 4, the next iteration of coastal fishing communities products may be a significant workload for the Mar 2021 meeting, pending direction from MRC under Agenda Item 8 (today's meeting).

COMMITTEE STAFF SUMMARY FOR NOVEMBER 10, 2020 MRC

Discuss and Recommend New MRC Topics

Today is an opportunity to identify any potential new agenda topics to recommend to FGC for referral to MRC. No new topics are recommended by staff.

Significant Public Comments (N/A)

Recommendation

FGC staff: Discuss priorities, review list of topics to clarify those to schedule as updates versus discussion or recommendation items, and determine if any additional referred topics on the work plan should be scheduled for the Mar 2021 MRC meeting.

Exhibits

1. [MRC work plan](#), dated Nov 2, 2020
2. [FGC perpetual timetable for regulatory actions](#), dated Oct 20, 2020

Committee Direction/Recommendation

The Marine Resources Committee recommends that the Committee work plan be updated with the following changes: _____.

Update on California Grunion
MRC Meeting July 29, 2020

- A petition was submitted by Dr. Karen Martin to the California Fish and Game Commission (Commission) in June 2019 proposing to establish a bag limit and increase the seasonal closure regulations for California Grunion (Tracking number: 2019-14).
- The California Department of Fish and Wildlife (Department) reviewed the petition and agreed that a change in regulations may be needed due to observed declines of Grunion spawning on beaches over the past decade. In February 2020, the Commission granted the petition in concept for consideration in a future rulemaking and the Department was asked to develop a regulatory package for the Commission's consideration.
- At the February 2020 Commission meeting, the Department proposed to prepare an Enhanced Status Report (ESR) on California Grunion as outlined in the 2018 Master Plan for Fisheries scaled-management framework to help address the petition.
- The Department completed an ESR for Grunion in May 2020. The ESR presents available information on the species, fishery, current management, and monitoring efforts. The Department agrees the existing management measures for the Grunion fishery may need to be adjusted to address declines in the population and habitat loss concerns. The Department will provide specific possible regulatory amendments to the Marine Resources Committee in November 2020.
- In June 2020, the Department received the most recently collected data on Grunion abundance from Dr. Karen Martin, who is the executive director of the Grunion Greeters, a citizen-science organization. Numbers of Grunion spawning on beaches remains low based on the qualitative data collected.
- In June 2020, letters were sent to 95 representatives of California Native American Tribes notifying them about potential regulation changes for California Grunion harvest and requested their input.
- Further outreach is being conducted by the Department. A questionnaire regarding the California Grunion fishery was developed in July 2020 and is available for the public on the Department website at (<https://wildlife.ca.gov/fishing/ocean/grunion>).



Proposed Regulations for the California Grunion Recreational Fishery



Photo Credit: CDFW

10 November 2020

Presented to:

**Marine Resources
Committee**

Presented by:

**Armand Barilotti
Environmental Scientist
Marine Region**



Life History of California Grunion

- Scientific name: *Leuresthes tenuis* (family *Atherinopsidae*)
- Range: Bahía Magdalena, México to Tomales Bay, CA
- Habitat: Nearshore coastal waters and bays
- Size: up to 8 inches (19 cm) in total length
- Life span: may live to be at least 4 years old
- Spawning season: February - September, peak April - June
- Reproduction: beach themselves at night during the 4 high tides following a full or new moon.

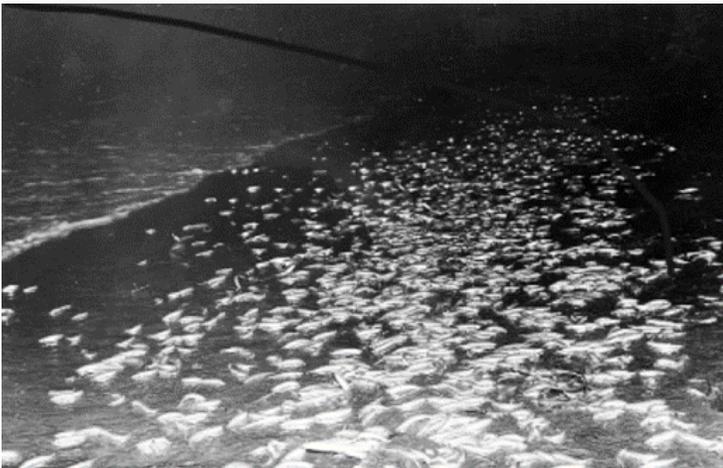


Photo Credit K. Walker, CDFW



Cultural Significance of Grunion

- Traditional food source for Native Americans
 - Found in middens dating back centuries
- Popular activity for Californians since the early 1900s
- First regulations implemented in 1927 due to declining population
 - Seasonal closure
 - Prohibited the use of equipment to catch Grunion

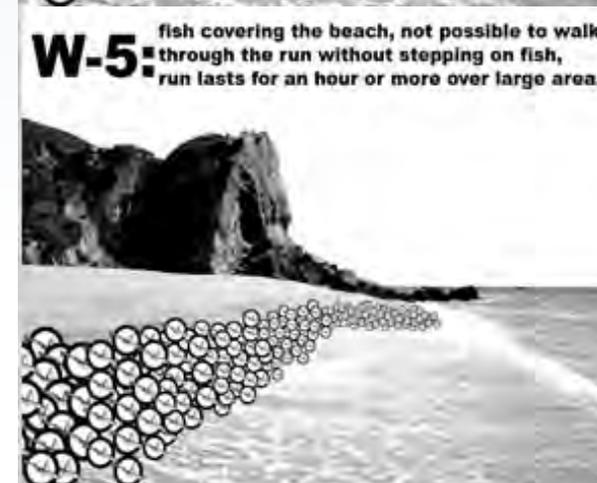
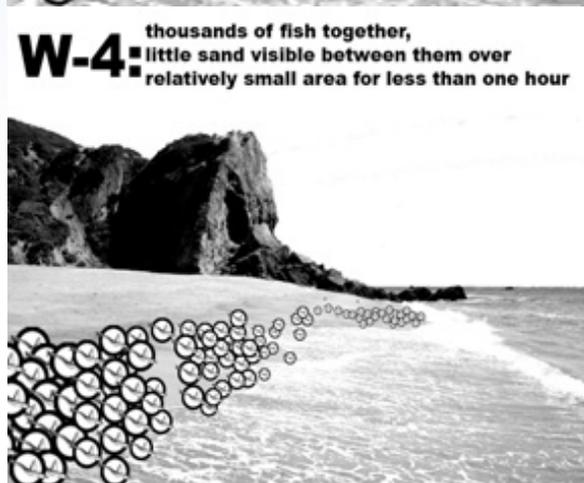
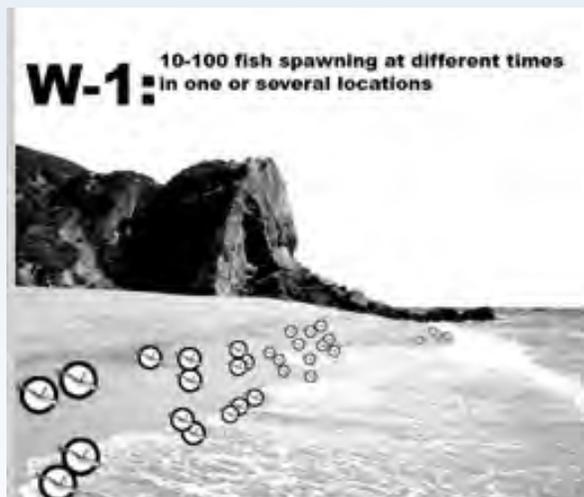
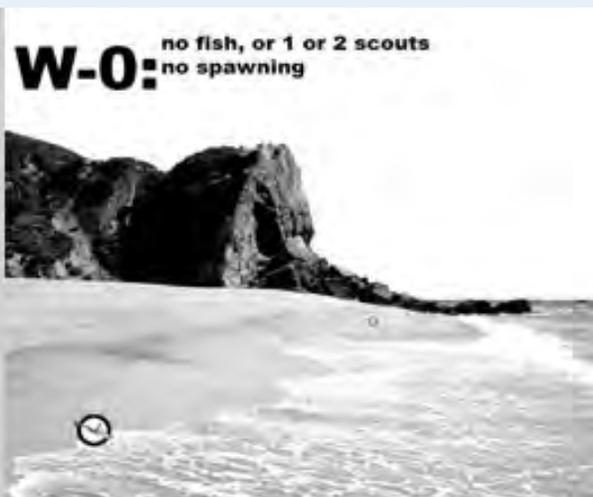


Source B. Walker, 1952



Monitoring Grunion Abundance

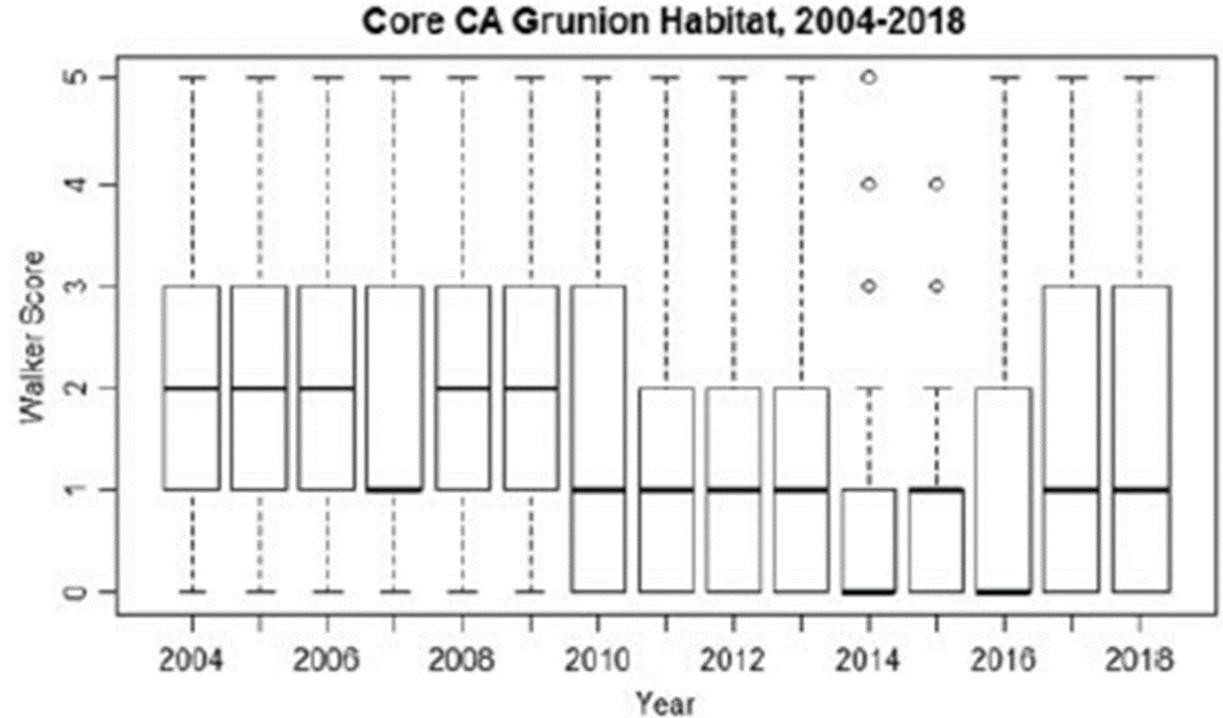
- Grunion Greeters (citizen scientists) monitor Grunion population in California
- Walker Scale: qualitative metric used to estimate Grunion abundance





Declining Grunion Abundance 2004 - 2018

- Walker Scale Graph (Martin et al. 2019)
 - Median Grunion run scores drop from W2 in the 2000s to W1 or less by 2010s
 - 75% of Walker scores are W3 or less for all years
 - Range includes Grunion runs of W0 to W5 for most years
 - Only available Grunion abundance data

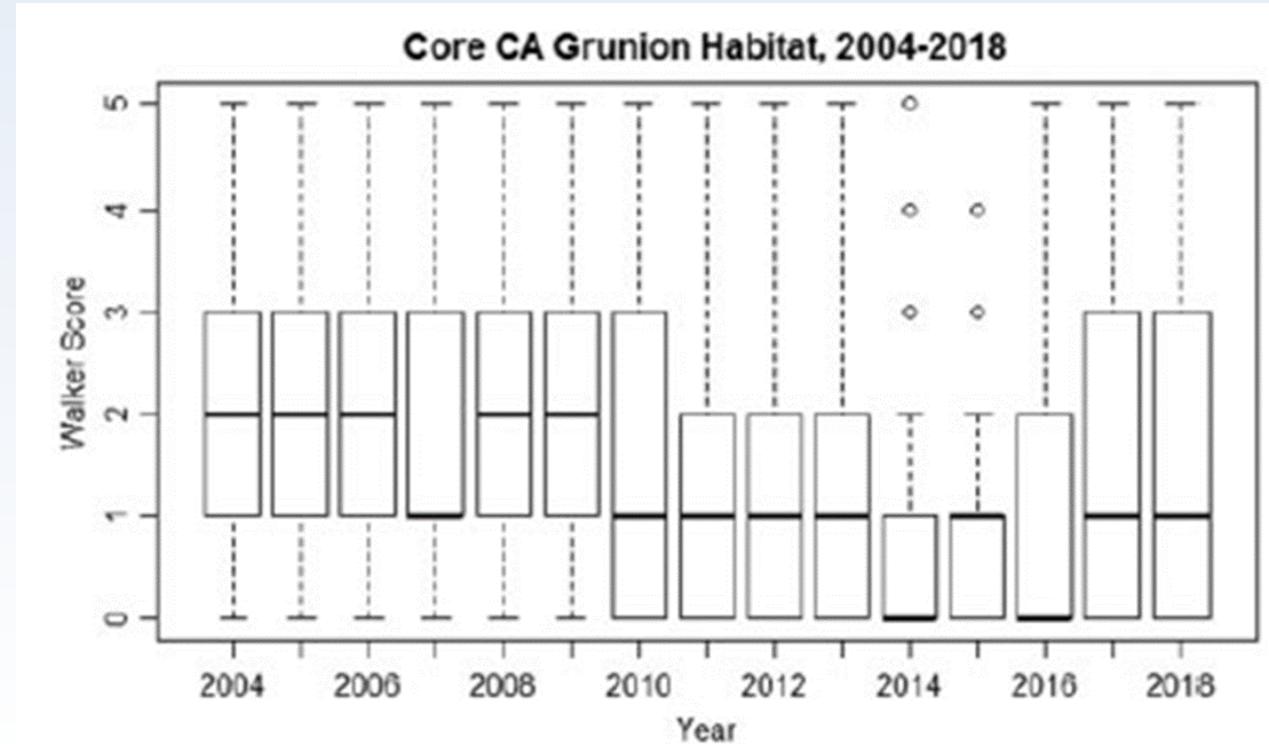


Source Martin et al. 2019



Declining Grunion Abundance 2004 - 2018

- Potential causes for the observed decline in the Grunion population
 - Habitat loss due to sea level rise
 - Beach grooming which results in the significant loss of laid eggs
 - Overharvesting of the resource
 - Sand nourishment projects
 - Pollution
 - Other effects from climate change



Source Martin et al. 2019



Current Regulations for Recreational Fishery

- Grunion do not have a bag or possession limit
- Grunion may only be taken by hand, no gear is permitted
- Season closed from April 1 – May 31



Photo Credit K. Walker, CDFW





Petition to the Commission

Dr. Karen Martin

- 10 fish possession limit
- Season closed from April 1 – June 30 south of Point Conception
- Season closed from April 1 – August 30 north of Point Conception



Photo Credit K. Walker, CDFW





Public Outreach

- Notification of the Californian Native American Tribes
 - 95 letters mailed to tribal leaders
 - Received responses from 6 Tribes:
 - 3 Tribes support more regulation
 - 3 Tribes had no comment
- Grunion fishery questionnaire
 - Locations, effort, and take
 - Potential new regulations
 - Online survey announced via Facebook post and Marine Region Blog
 - Posted July 2020



Photo Credit CDFW



Public Outreach

- Survey results as of October 20, 2020:
 - 6 participants ranging from San Diego to San Francisco
 - Most take 20 or less Grunion per night
 - June and July are the most targeted months
 - Primarily used for food or bait
 - 5 out of 6 do not want additional seasonal closure
 - 10 to 20 Grunion was deemed an appropriate possession limit



Proposed Regulations

- Creation of a possession limit for Grunion from 10 - 20 fish
- Proposed open season July 1 – March 31, add June to fishing closure for whole state.



Photo Credit K. Walker, CDFW



Thank You

Armand Barilotti

Environmental Scientist

Southern California Fisheries Research and Management Project

Department of Fish and Wildlife, Marine Region

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Los Alamitos, CA 90720

Phone: (562) 342-7164

Email: Armand.Barilotti@Wildlife.ca.gov

<https://marinespecies.wildlife.ca.gov/california-grunion/true/>



Existing Structures in Marine Protected Areas

10 November 2020

Presented to:

**Marine Resources Committee
Virtual Webinar**

Presented by:

**Amanda Van Diggelen
Environmental Scientist
Marine Region**



Need for Proposed Amendments



Piers



Seawater intake
pipes



Seawalls



Need for Proposed Amendments

- Pre-existing artificial structures were not accounted for in central and north central coast MPA planning processes
- Proposal has been developed in consultation with the MPA Statewide Leadership Team
- No change in fishing regulations within MPAs



Summary of Proposed Amendments

- Establish a definition of a pre-existing artificial structure
- Allow for maintenance, repair, replacement, and removal of pre-existing artificial structures
- Establish a definition of a buffer zone
 - Immediate area around pre-existing artificial structures
- Establish regional implementation dates



Proposed Timeline

- Notice to CA Tribes and tribal communities: October
- Marine Resources Committee Update: November 10
- Notice Hearing: February 2021
- Discussion Hearing: April 2021
- Adoption Hearing: June 2021
- Potential Implementation in fall of 2021

Thank You

Amanda Van Diggelen, Environmental Scientist

Amanda.VanDiggelen@wildlife.ca.gov



COMMITTEE STAFF SUMMARY FOR MARCH 17, 2020 MRC*For Background Purposes***7. MARINE AQUACULTURE IN CALIFORNIA****Today's Item**Information Action

Receive update on marine aquaculture and discuss near-term priorities and potential committee recommendations related to:

- (A) DFW aquaculture informational report, status of programmatic environmental impact report (PEIR), and proposed next steps; and
- (B) Potential temporary hiatus in considering new state water bottom lease applications.

Summary of Previous/Future Actions

- Discussed best management practices in shellfish aquaculture 2016-2017; FGC and MRC, various
- FGC referred topic of future lease planning to MRC Jun 21-22, 2017; Smith River
- MRC initial discussion on future lease planning Jul 20, 2017; MRC, Santa Rosa
- MRC received overview of current aquaculture leases and update on future lease planning Mar 6, 2018; MRC, Santa Rosa
- FGC referred PEIR topic to MRC Apr 18-19, 2018; Ventura
- MRC received general overview of PEIR Nov 14, 2018; MRC, Sacramento
- MRC received PEIR update Mar 20, 2019; MRC, Sacramento
- FGC referred discussion of potential temporary hiatus on new lease applications to MRC Feb 21, 2020; Sacramento
- **Today's program update and discussion Mar 17, 2020; MRC, Santa Rosa**

Background

FGC has the authority to lease state water bottoms to any person for the purpose of conducting aquaculture in marine waters of the State, under terms agreed upon between FGC and the lessee (sections 15400 and 15405, California Fish and Game Code). FGC is prohibited from issuing leases for commercial offshore marine finfish aquaculture in California until a programmatic environmental impact report (PEIR) evaluates a management framework for potential future offshore marine aquaculture.

There are currently 17 active, FGC-issued, state water bottom leases held by 10 growers across the state for cultivating shellfish (16 leases) or seaweed culture (1 lease). In addition, FGC has received 3 applications for new state water bottom leases that are currently undergoing DFW and/or environmental reviews necessary before FGC schedules them for consideration.

Topics related to current lease management, desired enhancement of the state aquaculture program, and possible pathways to achieving an enhanced program have been discussed at various FGC and MRC meetings since 2016.

COMMITTEE STAFF SUMMARY FOR MARCH 17, 2020 MRC***For Background Purposes***

In Mar 2018, MRC received an overview of existing leases and current management efforts from DFW, and discussed how management efforts by DFW and other agencies may contribute to future aquaculture planning and enhanced management of the state aquaculture program (Exhibit 1). However, the discussion highlighted a disparity between proposed program development areas and staff capacity to pursue them. In light of the competing interests and needs, MRC made a recommendation for how to prioritize the various planning efforts.

In Apr 2018, FGC accepted the MRC recommendation and, based on FGC direction, MRC received an overview and update on PEIR development at the Nov 2018 and Mar 2019 meetings (see Exhibit 2 for background).

For today's meeting there are two areas of focus for discussion: aquaculture in California generally and new state water bottom leases.

- (A) DFW will provide an update on its recommendations regarding the aquaculture PEIR, including discussions and public engagement it believes are necessary to clarify a long-range vision for California's marine aquaculture development. DFW is developing an aquaculture information report and anticipates the report will be available at today's meeting. DFW suggests that the report could serve as a foundation to engage interested parties in discussions about current and future marine aquaculture management and development in California.
- (B) FGC referred to MRC a discussion about a potential temporary hiatus in considering new state water bottom lease applications, excluding the applications already received (two proposed offshore sites in southern California, and one proposed site in Tomales Bay). The three applications are the first new lease applications FGC has received in over 25 years; currently there is not an established process to guide FGC review and consideration of lease applications, coordination protocols between FGC and DFW staff need to be further developed, staff roles and responsibilities need to be more clearly articulated, and practices for communicating expectations with lease applicants need to be refined. Available staff resources are a concern; staff needs to focus on managing the 17 existing leases and processing the three applications already under consideration before undertaking additional new leases. It may be helpful for decisions regarding prospective new lease applications to be made within the context of a broader statewide policy and vision.

Significant Public Comments

1. A mariculturist supports placing a hiatus on considering new state water bottom leases, requests that future lessees be subject to more stringent experience and qualification requirements, and recommends provisions for a program that would train new lessees in mariculture, such as providing small trial plots to new lessees and internships in mariculture. Requests clarification on where future leases will be placed (Exhibit 3).
2. A non-governmental organization expresses support for placing a hiatus on considering new state water bottom leases until a review of aquaculture activities by FGC and other agencies is complete, and asks that FGC exercise caution when considering new leases, especially in Tomales Bay, due to potential impacts of shellfish farms on bay food webs and shorebird populations (Exhibit 4).

COMMITTEE STAFF SUMMARY FOR MARCH 17, 2020 MRC

For Background Purposes

3. A non-governmental organization expresses a desire for a more workable permitting process for restorative aquaculture, requests that the State remove barriers to entry into restorative aquaculture, and asks that a completed PEIR and a more streamlined permitting process be established by the end of 2020 (Exhibit 5).

Recommendation

- (A) Consider requests received from DFW during the meeting, and
- (B) Consider supporting a temporary hiatus on considering new state water bottom lease applications not already received by FGC and schedule a follow-up discussion for a future MRC meeting.

Exhibits

1. Background document: Staff summary for Mar 6, 2018 MRC meeting, Agenda Item 8
2. Background document: Staff summary for Mar 20, 2019 MRC meeting, Agenda Item 8
3. Email from Bernard Friedman, Santa Barbara Mariculture Company, received Mar 2, 2020
4. Email from Nils Warnock, Audubon Canyon Ranch, received Mar 4, 2020
5. Email from Katherine O'Dea, Save Our Shores, received Mar 5, 2020

Committee Direction/Recommendation (N/A)

From: bernard@sbmariculture.com <bernard@sbmariculture.com>

Sent: Tuesday, October 27, 2020 8:56 PM

To: FGC <FGC@fgc.ca.gov>

Subject: FGC MRC Meeting Nov. 10, 2020 Agenda Item 5. New Marine Aquaculture Leases in California

FGC MRC Meeting Nov. 10, 2020

Agenda Item 5. New Marine Aquaculture Leases in California

I am a major proponent for offshore aquaculture. I have made it my life's work to develop and promote it. Please, for the future of aquaculture, **do not create any new leases** until you have created a vision and a comprehensive management program for implementation of new state water bottom leases.

Please direct all new applicants (including the ones that are already in the queue) to port districts that are willing to do the heavy lifting for training and vetting of new offshore leases. I am lobbying the Ventura Shellfish Enterprise to head back into State waters. They will be able to have more success as offshore mussel farmers in more protected waters. I will be applying for one of those leases when their operations plan is more comprehensive and realistic.

Your permit counter does not work. How can you accept any applicants when it does not work for your current State water bottom lessees.

Right now, applicants will still be going through the permit process sequentially for each agency. **I will be voicing my opposition of your first applicant throughout this process.** My reasons have been detailed in a letter I sent to the department and the commission dated July 2, 2020.

The last offshore mussel farm to copy my farm was a complete disaster. Catalina Sea Ranch was non-compliant to the permit conditions of their lease and ended up in bankruptcy with major liabilities. It still hasn't been made clear to why a non-compliant and bankrupt company can be bought at a private auction for \$1.75 million so the next unqualified lease holder can get another chance. That lease should be going back into the public trust. What message does this send to the people of California?

This state is not ready to manage individual leases at the State level. The last 20 years should be proof of that. The Fish and Wildlife department has accepted an application for 100 acres right next to my farm by an unqualified candidate. **This minimizes my life's work and is completely disrespectful to all my success which I have worked so hard for.**

There is only enough room for about 3 shellfish farms off the coast of Santa Barbara in State Waters. I would love an experienced and capable neighbor to contribute to what I have already built to make a stronger group of shellfish farmers. Seaweed and scallop farming would be very good compliments. This applicant that has applied to be next to me will only compete with what I have already created. The applicant adds nothing to the future of aquaculture. It is such a slap in the face for the Department to allow for someone to completely rip me off.

By giving a lease to an unqualified farmer, you are sacrificing future opportunities for more qualified applicants. Applicants that will contribute significant contributions to offshore aquaculture. **Please do not sacrifice what little that is available to such a weak applicant.**

Bernard Friedman
Santa Barbara Mariculture Company

From: [REDACTED]
Sent: Friday, September 18, 2020 12:14 AM
To: FGC <FGC@fgc.ca.gov>
Cc: Ashcraft, Susan@FGC <Susan.Ashcraft@fgc.ca.gov>
Subject: FG 1 petition for regulation change

Dear Commissioners,

I am submitting an FG 1 Petition to the California Fish and Game Commission for Regulation Change. Please find the attached FG 1 petition "FG Comm Petition_Nancy Caruso" and all the referenced supporting documents for the September 22, 2020 Fish and Game Commission Meeting. They were sent before the Supplemental Comment Deadline of September 18 at noon.

Commissioners Please Note: Supporting Reference materials for *MBC 2019* report has not yet been distributed to the public but is available online here https://1drv.ms/u/s!AklZpj2SiR6xpG_MWqq-Hs8Rqsuo?e=ouAaVG

A Powerpoint overview of that report is attached

Sincerely,

Nancy L. Caruso
Marine Biologist/Founder
Get Inspired

[REDACTED]
www.GetInspiredinc.org



You can Support our Green Abalone Project here www.gofundme.com/abalone



Tracking Number: (2020-014)

To request a change to regulations under the authority of the California Fish and Game Commission (Commission), you are required to submit this completed form to: California Fish and Game Commission, (physical address) 1416 Ninth Street, Suite 1320, Sacramento, CA 95814, (mailing address) P.O. Box 944209, Sacramento, CA 94244-2090 or via email to FGC@fgc.ca.gov. Note: This form is not intended for listing petitions for threatened or endangered species (see Section 670.1 of Title 14).

Incomplete forms will not be accepted. A petition is incomplete if it is not submitted on this form or fails to contain necessary information in each of the required categories listed on this form (Section I). A petition will be rejected if it does not pertain to issues under the Commission’s authority. A petition may be denied if any petition requesting a functionally equivalent regulation change was considered within the previous 12 months and no information or data is being submitted beyond what was previously submitted. If you need help with this form, please contact Commission staff at (916) 653-4899 or FGC@fgc.ca.gov.

SECTION I: Required Information.

Please be succinct. Responses for Section I should not exceed five pages

1. Person or organization requesting the change (Required)

Name of primary contact person: Nancy Caruso, Marine Biologist, Executive Director of Get Inspired

Address: [Redacted]

Telephone number: [Redacted]

Email address: [Redacted]

2. Rulemaking Authority (Required) - Reference to the statutory or constitutional authority of the Commission to take the action requested:

Section 200 and 205, Fish and Game Code §632. Marine Protected Areas (MPAs), Marine Managed Areas (MMAs), and Special Closures. “The commission may authorize research, education, and recreational activities, and certain commercial and recreational harvest of marine resources, provided that these uses do not compromise protection of the species of interest, natural community, habitat, or geological features.” “The designating entity or managing agency may permit research, education, and recreational activities, and certain commercial and recreational harvest of marine resources PRC §36710(c).”

Added per Nancy Caruso email - 10/4/20: For Section 30 of T14CCR: Section 6750, Fish and Game Code. Section 632 of T14CCR: Sections 200, 205(c), 265, 399, 1590, 1591, 2860, 2861 and 6750, Fish and Game Code; and Sections 36725(a) and 36725(e), Public Resources Code

3. Overview (Required) - Summarize the proposed changes to regulations:

1. Request to modify Section 30.00, Title 14, CCR30.00.

KELP GENERAL. (a) Except as provided in this section and in Section 30.10 there is no closed season, closed hours or minimum size limit for any species of marine aquatic plant. The daily bag limit on all marine aquatic plants for which the take is authorized, except as provided in Section 28.60, is 10 pounds wet weight in the aggregate. (b) Marine aquatic plants may not be cut or harvested in state marine reserves. Regulations within state marine conservation areas and state marine parks may prohibit cutting or harvesting of marine aquatic plants per sub-section 632(b) [marine protected area regulations].

-Change the recreational take of *Sargassum horneri* from 10 pounds wet weight to “no limit” April through October (during non-reproductive season).



2. Request to modify 14 CCR § 632 Crystal Cove SMCA: Area restrictions defined in subsection 632(a)(1)(C) apply, with the following specified exceptions:

-Allow for unlimited recreational take of *Sargassum horneri* in the Crystal Cove SMCA April through October (during non-reproductive season).

3. “The commission may authorize research, education, and recreational activities, and certain commercial and recreational harvest of marine resources...”

-Allow for localized, controlled, year-round removal of *Sargassum* for 3 years as a research project in Crystal Cove SMCA under direction of Nancy Caruso of Get Inspired to determine if *Sargassum* is prohibiting kelp recruitment, recovery, and experiment with techniques for eradication.

4. Rationale (Required) - Describe the problem and the reason for the proposed change:

The problem is that *Sargassum horneri* has invaded our coast and is spreading rapidly. It is having a negative impact on our kelp forest ecosystem. DFW has not acted in accordance with the Aquatic Invasive Species Management Plan. Below, we lay out the reasons for the proposed changes to make strides to eradicate it.

1. DFW failed to respond and stop the spread of the invasive species *Sargassum horneri*

Sargassum horneri is native to Eastern Asia. It has spread aggressively throughout southern California, USA, and Baja California, México since it was discovered in Long Beach in 2003 and poses a major threat to the sustainability of native marine ecosystems in this region (Marks et al. 2015). Now it is ubiquitous in the region and had been found at three of the five Channel Islands (Anacapa, Santa Cruz and Santa Barbara) (Marks et al. 2015). Earlier this year, it was documented by divers in Monterey, CA (pers comm, 2020). Kaplanis et al. 2016 reported that the rapid and uncontrolled spread of *Sargassum* has serious implications for its expansion along the west coast of North America.

“California does not have an official rapid response plan for AIS, does not have a designated funding source for providing a rapid response, and no agency is designated with overall responsibility for AIS management. For this reason, it is unknown whether the necessary elements to conduct a rapid response operation will come together when the need arises. If the commitment, expertise, and funding fail to coalesce, the state could be faced with substantial environmental and economic consequences caused by AIS infestations.” (CA AIS Mgmt plan Appendices 2008).

The invasion of *Caulerpa taxifolia* in Southern California, in 2000, was met with swift action and eradication. This species could have easily spread and caused widespread issues in our bays and wetland areas. The Southern California Caulerpa Action Team (SCCAT) was established to quickly and effectively respond to the discovery of this algae in Southern California. *Caulerpa* was quickly contained and even treated with chlorine, killing the plant and its roots. There was no such effort for *Sargassum*. Now let us, the divers who love our reefs, remove this invasive species. Hopefully we can make an impact on eradication of this species which is of no benefit to our California coast. I hope it is not too late to stop this invasion, so I ask that you allow the community to help eradicate it in the areas that are important to them: where they dive, spearfish, or swim. By allowing unlimited take of *Sargassum*, we can make an impact and help our kelp to thrive.

2. *Sargassum horneri* is not a marine resource

“MPAs protect the diversity and abundance of marine life, the habitats they depend on, and the integrity of marine ecosystems.”. <https://wildlife.ca.gov/conservation/marine/MPAS> *Sargassum* threatens the integrity of our marine ecosystem. Currently, *Sargassum* is being protected in our MPAs as a “marine resource” and the giant kelp is suffering. In the Crystal Cove SMCA in Orange County, you can take finfish, urchins, and lobsters but you can’t take an invasive species. This is illogical and must be



changed. The proposed “season” for recreational take from April-October was meant to disentangle from the argument that it can be spread when reproductive. Sargassum is an annual. In general, it recruits in early Summer, becomes reproductive in November, and dies off in April. By creating this “season” of take, that argument cannot be used, as it has for the last 17 years of Sargassum’s spread. You have nothing to lose.

In 2015, Cruz-Trejo et. al. studied Sargassum in Baja, Mexico and found the most significant impact to be severe reduction of the canopy forming species on their study sites. In 1982 Ambrose and Nelson found that *Sargassum muticum* appeared to prevent giant kelp recruitment and removal of the invasive species resulted in a significant increase in giant kelp recruitment. They also found higher densities of giant kelp in removal areas. Shading at a critical time in the giant kelp life cycle is suggested as a possible mechanism for the inhibition of giant kelp recruitment (Ambrose and Nelson 1982).

I have been observing and monitoring the reefs of Orange County for 18 years. The warm water events from 2014-2016 gave us our first look at *Macrocystis* recovery, after a disruption event, WITH *Sargassum horneri* in its ecosystem. Sargassum is an annual and recruits in early summer, BEFORE giant kelp recruits later in winter months. Sargassum has taken advantage of the *Macrocystis* winter recruitment cycle. When the warm water and high surf decreased kelp and other native algal densities during the warm water “blob” of 2014-2015 followed by an El Nino in 2016, the Sargassum took advantage of the space on the reef prohibiting kelp from recruiting and recovering from these “disruption” events. This is evident in the MBC Aquatic Sciences *Status of the Kelp 2019* report. This report is released annually on the status of the Southern California kelp beds. It contains aerial surveys of our kelp and even tracks local available nitrate (*nutrient quotient*) for kelp growth. Kelp surveys, from this report, confirm that even though the 2018-2019 years had adequate nutrients and temperatures conducive to kelp recovery and growth, *Macrocystis* densities did not rebound after the 3 years of warm water. Why? There is no room to recruit on the reefs.

Most herbivores do not prefer Sargassum as a food choice and this has helped lead to its success (Marks et. al 2020). *Sargassum horneri* forms monospecific dense forests that fish cannot even swim through, it also limits light penetration to the reef further inhibiting competitors.

Marks et al 2017 findings suggested that controlling *S. horneri* via removal will be most effective if done over large areas during cool-water years that favor native algae. She goes on to suggest that such efforts should be targeted in places such as novel introduction sites or recently invaded areas of special biological or cultural significance. I think the Crystal Cove SMCA fits this description and this year is the year to do it because a *La Nina* is projected. On the Crystal Cove SMCA reefs, in particular, there has been a shift, since our kelp restoration activities in Orange County in 2002-2010, from a *Macrocystis* forest with healthy understory of other alga and encrusting organisms to a desolate Sargassum covered reef.

3. Reasons we want to do research in the Crystal Cove SMCA

- It is one of the least restrictive MPAs in the system: Take of lobster, finfish, and urchins is already permitted
- We have an 18-year history working in the kelp forests of Crystal Cove, Newport Beach and Laguna Beach
- We have a team of over 300 volunteer divers to help with the effort
- The annual kelp surveys and nutrient data collected by MBC Aquatic Sciences includes this MPA
- Good beach diving access, good boat diving access (Newport Harbor)
- Sargassum densities currently as high as 13.85 plants/m²
- Kelp has decreased 98% in 2019



- The “nutrient quotient”, calculated by MBC Aquatic Sciences, is calculated from data taken at the Newport Pier just 2 miles away. The next closest location is Oceanside (35 miles away). This will give us valuable insight.
- We have the historical knowledge of where giant kelp used to grow in this SMCA
- All the rocky reefs in Orange County with Sargassum growing on them are located in MPAs.
- DFW’s recommended test site (Decision Tree) in San Clemente does not contain Sargassum

We have already asked for an SCP for this research project but because of the “Decision Tree”, it was denied with the rationale that “It can be done somewhere else” but the next closest rocky reef outside of our Orange County MPA network has no Sargassum (San Clemente, CA). We argue that the requested project location (Crystal Cove SMCA) is unique in several ways and we lay that argument out below. For project details see attached “Timing on *Sargassum horneri* removal as a technique for eradication”

It is clear that Sargassum is a threat to our current native kelp ([Cruz-Trejo et al 2015](#)). We have found it at densities as high as 13.8 plants per meter square in Crystal Cove SMCA which is 100% cover in that same meter square at maturity (per observation). It is also clear that despite favorable ocean conditions for the last 2 years kelp densities have decreased (MBC, 2019). Almost all of Orange Counties rocky reefs are in MPAs. In 2019 Crystal Cove SMCA has lost 98% of its kelp, The Laguna Beach SMR lost 89% in North Laguna and 95% in South Laguna, the South Laguna/ Dana Point SMCA kelp beds totally disappeared (MBC, 2019). I believe this is because of Sargassum. In a time when we are relying on these protected areas to preserve our ecosystems, it is vital that we eliminate this threat and study how we can stop its spread. If we do not act, we are countering the very reasoning and rationale for establishing the MPAs. There is no downside to taking this action. We have hundreds of volunteers ready to help. The knowledge gained by this study can be used to eradicate Sargassum in other areas.

The precedence has already been set for this type of action on the North Coast where divers have been given permission to cull purple urchins in the Pacific Grove Gardens SMCA in an effort to restore our precious kelp beds struggling to survive. In an all-out effort, divers are coming together to figure out how they can help preserve the kelp and save our abalone populations as well as the other species that rely on kelp. We only hope that it is not too late. It is clear we cannot afford to wait any longer with regards to Sargassum. It is in the spirit of the MLPA that these areas be protected from invasive threats to allow our native wildlife to thrive, that was the intention. Please use your authority for adaptive management to allow the public to help with this problem. We can help to “save” our reefs from the takeover of Sargassum.

Nancy Caruso, marine biologist, has led a team of more than 300 volunteer divers working on the reefs of Newport Beach and Laguna Beach for 18 years. Restoring giant kelp, monitoring kelp forest recovery, fishes, algae and invertebrates. We have also outplanted abalone (Caruso, 2018) and we are monitoring abalone density, size, recruitment, and mapping abalone populations. We have partnered with DFW as well. This is our community and our reefs that we spoke out for at meetings to implement the MLPA. With the help of 500 students who grew kelp in their classrooms and 250 volunteer divers, we restored our kelp after being gone for 2 decades and want it protected and preserved. We see degradation of our reef communities by *Sargassum horneri* and we want to help fix it. We will conduct a localized removal experiment to test whether Sargassum is hindering kelp recruitment. We will work on the some of the same reefs where we conducted kelp restoration activities, abalone monitoring and restoration since 2002. This SMCA is a familiar large rocky reef system that offers plenty of expanse for a replicated, controlled studies. All of the rocky reefs in Orange County are in MPAs except for the Wheeler Reef system in San Clemente. Steve Schroeter of UCSB (who is managing the monitoring



program for the reef) stated that they found only two Sargassum plants in their 92 transects in 2019. The Crystal Cove SMCA used to be a lush garden of algae and a healthy kelp forest where Wheeler North conducted many kelp restoration experiments. We have data going back to the 1980's from Joe Valensic and we collected data on these reefs from 2002-2012. As concerned scientists, we see a problem, we think we may have the answer, and we want to test it. We can add to the available science through a controlled research approach and then share this information with you, for better management practices and to manage our kelp forests like the important resources that they are.

SECTION II: Optional Information

5. **Date of Petition:** 9/18/2020

6. **Category of Proposed Change**

- Sport Fishing
- Commercial Fishing
- Hunting
- Other, please specify: Click here to enter text.

7. **The proposal is to:** *(To determine section number(s), see current year regulation booklet or <https://govt.westlaw.com/calregs>)*

X Amend Title 14 Section(s): 30.00

(a) Except as provided in this section and in Section 30.10 there is no closed season, closed hours or minimum size limit for any species of marine aquatic plant. The daily bag limit on all marine aquatic plants, except as provided in 30.00 (c), for which the take is authorized, except as provided in Section 28.60, is 10 pounds wet weight in the aggregate.

(b) Marine aquatic plants, except as provided in 30.00 (c), may not be cut or harvested in state marine reserves. Regulations within state marine conservation areas and state marine parks may prohibit cutting or harvesting of marine aquatic plants, except as provided in 30.00 (c), per sub-section 632(b)
(c) Title 14 CCR § 632 Crystal Cove State Marine Conservation Area.

(B) Area restrictions defined in subsection 632(a)(1)(C) apply, with the following specified exceptions:

- 1. The recreational take of finfish [subsection 632(a)(2)] by hook and line or by spearfishing [Section 1.76], and spiny lobster and sea urchin is allowed.*
- 2. The commercial take of sea urchin; spiny lobster by trap; and costal pelagic species [Section 1.39] by round haul net [Section 8750, Fish and Game Code], brail gear [Section 53.01(a)], and light boat [Section 53.01(k)] is allowed. Not more than five percent by weight of any commercial coastal pelagic species catch landed or possessed shall be other incidentally taken species.*
- 3. Take pursuant to activities authorized under subsection 632(b)(133)(C) is allowed.*

(C) Beach nourishment and other sediment management activities, and operation and maintenance of artificial structures inside the conservation area is allowed pursuant to any required federal, state and local permits, or as otherwise authorized by the department.

(D) Take of all living marine resources from inside tidepools is prohibited. For purposes of this section, tidepools are defined as the area encompassing the rocky pools that are filled with seawater due to retracting tides between the mean higher high tide line and the mean lower low tide line.



X Add New Title 14 Section(s):

Subsection 30

(c) Invasive marine aquatic plant *Sargassum horneri* may be removed without a daily bag limit when the plants are not reproductive (April-October).

CCR § 632

(E) Unlimited recreational take of *Sargassum horneri* by hand during the months of April-October.

Repeal Title 14 Section(s): [Click here to enter text.](#)

8. **If the proposal is related to a previously submitted petition that was rejected, specify the tracking number of the previously submitted petition** [Click here to enter text.](#)

Or X Not applicable.

9. **Effective date:** If applicable, identify the desired effective date of the regulation. If the proposed change requires immediate implementation, explain the nature of the emergency: We request that this petition be approved immediately. We hope to take advantage of the forecasted La Nina conditions by starting the project now.

10. **Supporting documentation:** Identify and attach to the petition any information supporting the proposal including data, reports and other documents: (Attached)

(Identified)

-Research Proposal: *Timing of Sargassum horneri as a Removal Technique for Eradication*

-Cruz-Trejo et al 2015 Presence of *Sargassum horneri* at Todos Santos Bay, Baja California, Mexico: Its Effects on the Local Macroalgae Community *American Journal of Plant Sciences*, 2015, 6, 2693-2707
Published Online October 2015 in SciRes.

-Aquatic Invasive Species Management Plan and Appendices DFW 2008

-Caruso, Nancy L. (2017). Outplanting large adult green abalone (*Haliotis fulgens*) as a strategy for population restoration. *California Fish and Game* 103(4): 183-194

-Kaplanis NJ, et al (2016) Distribution patterns of the non-native seaweeds *Sargassum horneri* (Turner) C. Agardh and *Undaria pinnatifida* (Harvey) Suringar on the San Diego and Pacific coast of North America. *Aquatic Invasions* 11: 111–124,

-Marks, L.M. et al. 2015. Range expansion of a non-native, invasive macroalga *Sargassum horneri* (Turner) C. Agardh, 1820 in the eastern Pacific. *BioInvasions Records* 4(4)243-248.

-Marks LM, et al 2017. Assessment of control methods for the invasive seaweed *Sargassum horneri* in California, USA. *Manag Biol Invasion.*;8(2): 205–213. 10.3391/mbi.2017.8.2.08

-Marks, L.M. 2020. Niche Complementarity and Resistance to Grazing Promote the Invasion Success of *Sargassum horneri* in North America. *Diversity* 2020, 12, 54.

-MBC Aquatic Sciences 2020. *2019 Status of the Kelp Beds Orange and San Diego Counties*

Prepared for the Region Nine Kelp Survey Consortium. https://1drv.ms/u/s!AkLZpj2SiR6xpG_MWqq-Hs8Rqsuo?e=ouAaVG

11. **Economic or Fiscal Impacts:** Identify any known impacts of the proposed regulation change on revenues to the California Department of Fish and Wildlife, individuals, businesses, jobs, other state agencies, local agencies, schools, or housing: Increase in revenues to the California Department of Fish and Wildlife for fishing licenses that are required for removal activities. There are no additional costs predicted to the state of local agencies.



12. Forms: If applicable, list any forms to be created, amended or repealed:

[n/a]

SECTION 3: FGC Staff Only

Date received: [Click here to enter text.]

FGC staff action:

- Accept - complete
- Reject - incomplete
- Reject - outside scope of FGC authority

Tracking Number

Date petitioner was notified of receipt of petition and pending action: _____

Meeting date for FGC consideration: _____

FGC action:

- Denied by FGC
- Denied - same as petition _____
- Granted for consideration of regulation change

Tracking Number

**CALIFORNIA
AQUATIC INVASIVE SPECIES
MANAGEMENT PLAN**

January 2008

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APPENDIX A:
DRAFT
AUGUST 2007

RAPID RESPONSE PLAN

FOR AQUATIC INVASIVE SPECIES IN CALIFORNIA



Prepared by:

California Department of Fish & Game
Habitat Conservation Branch
The Invasive Species Program

**RAPID RESPONSE PLAN
FOR AQUATIC INVASIVE SPECIES IN CALIFORNIA
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VII. Interagency Agreements (attached following approval)

I. INTRODUCTION

Goal: The purpose of this plan is to provide a framework for an effective rapid response to the discovery of any aquatic invasive species (AIS) that is new to California, or of a population of established AIS that is outside of its known distribution in California.

In this document, "rapid response" means that soon after an aquatic species new to the State of California or a specific region of the state is discovered, 1) the state will make a determination of whether it is potentially detrimental and/or invasive and 2) if that is the case, the state will develop and implement a course of action. This also would apply to AIS that are discovered in an adjacent state in a waterway or lake that ultimately enters California.

Possible courses of action for newly discovered AIS may include an effort to eradicate the species, control its spread, prevent future introductions, minimize or mitigate the damage it causes, or study it further before any other action is taken. Rapid response is the second line of defense after prevention to minimize the negative impacts of AIS on the environment and economy of California. Once non-native invasive species become widespread, efforts to control them are typically more expensive and less successful than rapid response measures. The damage caused by an AIS that becomes widespread, and the actions that are taken to control it, may be more harmful to the environment than a successful rapid response.

California does not have an official rapid response plan for AIS, does not have a designated funding source for providing a rapid response, and no agency is designated with overall responsibility for AIS management. For this reason, it is unknown whether the necessary elements to conduct a rapid response operation will come together when the need arises. If the commitment, expertise and funding fail to coalesce, the state could be faced with substantial environmental and economic consequences caused by AIS infestations. Even if an ad hoc rapid response effort is made, the following consequences may result:

1. The effort may be compromised by less than adequate staff levels, authority and funding to carry out necessary actions.
2. Staff assigned on an ad hoc basis are less likely to have received training in advance that would help them function as effectively and efficiently as possible in this situation (e.g. Incident Command System training).
3. The effort may be compromised indirectly by staff in charge of the ad hoc effort spending their time trying to secure staff and funding for the response instead of leading the response itself.
4. The effort may not have the level of organization and accountability to be gained from following an official plan.
5. Some governmental and non-governmental entities may be less cooperative with an ad-hoc response than they would be if the response is a standard procedure that is based on official agency agreements.
6. Any resulting confusion could lead to a perception that public funds are mismanaged, that environmental regulations are not being followed, or that the interests of community leaders have been disregarded.

To address the threat posed to California habitats by new AIS introductions, and the lack of an organized plan and funding to address this threat, Chapter 6 (Task 4A1) of the California Aquatic Invasive Species Management Plan (CAISMP) calls for the development and implementation of a rapid response plan. The CAISMP was completed by the California Department of Fish and Game (DFG) in 2007. The CAISMP acknowledges that rapid response

to AIS in California may often require cooperation among a variety of local, state and federal agencies and organizations, and that formal agreement on a plan, in advance of need, increases the likelihood of responding in an effective manner.

This draft Rapid Response Plan will be available for review by agencies and organizations that are likely to have an interest in rapid response. DFG's Invasive Species Program will revise the plan based on the comments received. The goal is to arrive at a plan that can be the basis for agreements to cooperate on rapid response to AIS. In order to finalize, fund and implement the plan, it is hoped that cooperating agencies will assign staff to participate. DFG Invasive Species Program staff will provide coordination for the interagency activities called for in the agreement(s).

Please note that the procedure section of this plan (Section III) is followed by the planning section (Section IV). The order of these sections is deliberate and meant to emphasize that the objective is to have a working product. Both the procedure and planning sections of this document discuss the need to collect data to evaluate the feasibility and success of the plan. This rapid response plan is meant to fit into an adaptive management strategy where evaluation can lead to improved procedures.

It is not possible to plan proactively for every species that might become a nuisance in state waters, hence the need for this generic plan. It stands to reason, however, that a generic plan cannot be implemented as efficiently as a species- or location-specific plan. Therefore, rapid response plans for individual species or related groups of species at high risk of being introduced and becoming destructive should be formulated. This step is called for in Action 4A3 of the CAISMP.

To effectively protect state aquatic habitats from the impacts of AIS, California needs to develop and implement a comprehensive AIS early detection and reporting plan. This document does not attempt to address the issue of early detection, nor provide a detailed discussion of mechanisms for reporting AIS. It focuses on what happens after detection of a suspect AIS. Since some early detection and reporting of AIS already occurs, a rapid response procedure is considered the most immediate need.

II. LEGAL AUTHORITY FOR RAPID RESPONSE

Appendices B and C in the CAISMP provide general information on the federal and state government agencies and regulations involved in the management of AIS. Rapid response activities could potentially require state and/or federal permits, consultations or agreements related to the placement of fill or structures into state and/or federal waters, protection of state or federally listed species, or the protection of other special status plant or animal species. The normal timeline for obtaining permits issued under these laws may critically delay rapid response efforts. A streamlined regulatory permitting process for implementing the Rapid Response Plan will need to be developed and approved by participating agencies. Additionally, permission is necessary to work on private and public properties. Clear protocols need to be developed to avoid misunderstandings or illegal trespassing, while making the process of obtaining access as efficient as possible.

In addition to the laws relevant to AIS discussed in the CAISMP, there are laws that specifically address taking action during an emergency or under special circumstances. These laws can facilitate the implementation of a rapid response procedure. Examples include:

Creation of Emergency Regulations

Under California Government Code Section 11346.1, rulemaking state agencies, departments, commissions, offices and boards can adopt emergency regulations, which can remain in effect for up to 120 days. These are regulations that must take effect immediately for "preservation of the public peace, health and safety or general welfare" and must meet other requirements of that code section. The process for adoption of emergency regulations can be found at the Office of Administrative Law's web site (www.oal.ca.gov/emergency_reg.htm).

The California Department of Food and Agriculture (DFA) has specific statutory authority to establish quarantines to protect the state's agricultural industry from pests (Food and Agriculture Code Section 5301). If an AIS is discovered that has the potential to severely damage crops, water delivery, or flood control systems that support agriculture, DFA can invoke their authority to establish a quarantine area.

According to Section 660 of the Harbor and Navigation Code, any entity, local or state, authorized by law to adopt rules or regulations that govern matters relating to boats or vessels may adopt emergency measures within their jurisdiction as long as they are not in conflict with the general laws of the state relating to those matters. The emergency rules or regulations can be effective for up to 60 days and must be submitted to the Department of Boating and Waterways (DBW) on or before their adoption. DBW can authorize these emergency rules or regulations to be in effect for over 60 days if it is deemed necessary.

Use of a Pesticide Outside of its Registered Use

When dealing with species that are new to California, the technical experts participating in a rapid response incident may determine that the best solution is to use a pesticide outside of its registered use or to deploy a new end use product. Section 18 of the Federal Insecticide, Fungicide and Rodenticide Act (FIFRA) allows states to apply to use a pesticide for an unregistered use for a limited amount of time if the EPA determines that emergency conditions exist (<http://www.epa.gov/oppr001/section18>). Under Section 6206 of Title 3 of the California Code of Regulations (CCR), the DFA Director is permitted to apply for a Section 18 exemption when emergency conditions exist. Section 24 of FIFRA authorizes states to register an additional use of a federally registered pesticide or a new end use product to meet a special local need (www.epa.gov/oppr001/24c).

Experimental Unregistered Use of a Pesticide

Section 6260 of Title 3 of the CCR provides the conditions for obtaining a Research Authorization for the experimental use of a pesticide outside of its registered uses. Research Authorizations are administered by the California Department of Pesticide Regulation (DPR).

III. RAPID RESPONSE PROCEDURE

The initial steps in this procedure result in the determination of whether an active response is immediately necessary after a potential invasive species is reported. If immediate action is necessary, and requires more than simple, highly localized measures, resource management staff may decide to implement an incident command system (ICS) response. A set of criteria will be developed to help in this decision making process. Many of the steps listed below are likely to take place simultaneously or overlap to some degree. Examples of these include outreach, rapid assessment, and containment activities. A flow chart showing the general steps of this rapid response procedure is provided as Chart 1.

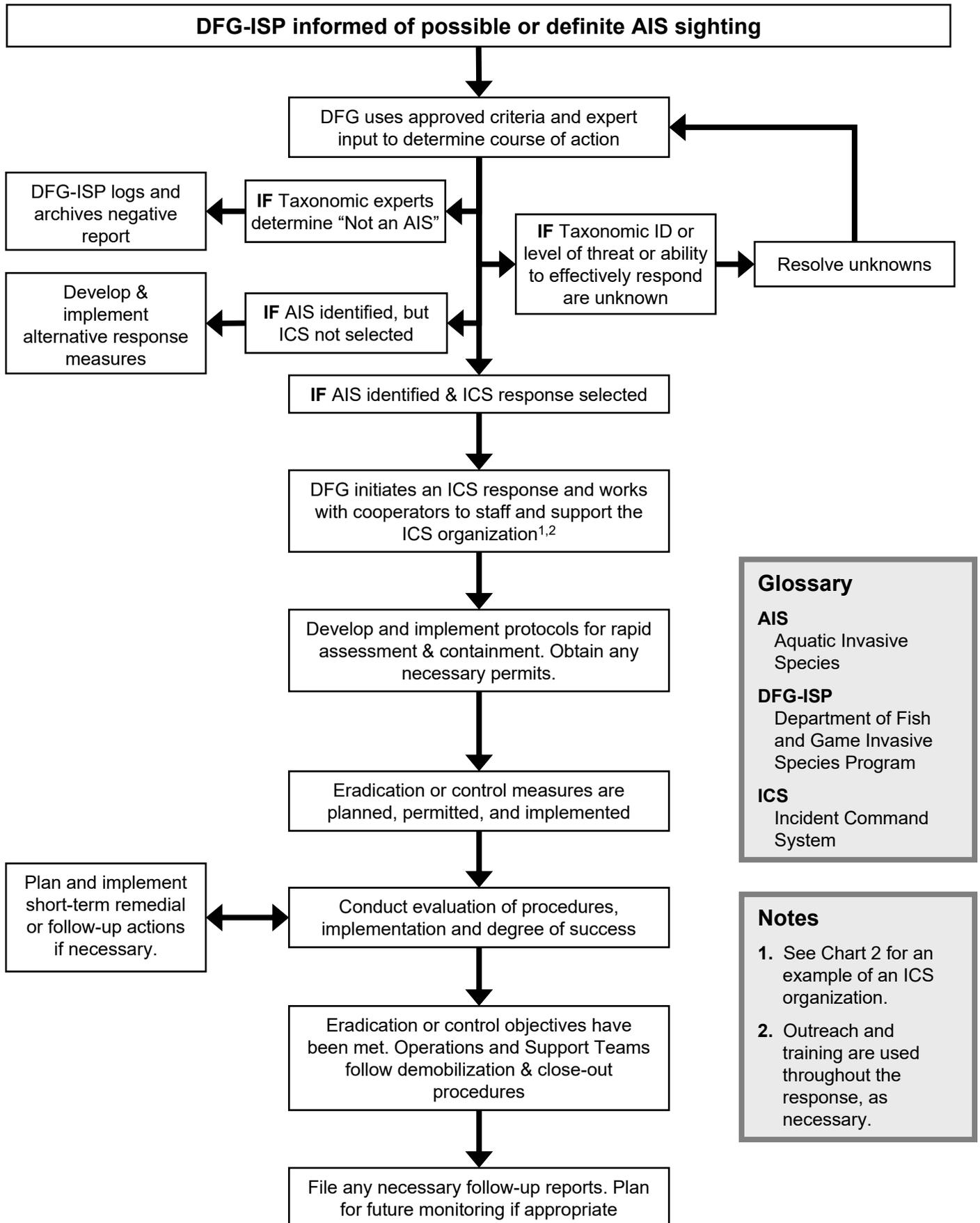
In an ICS response, participants are assigned specific roles in a well-defined hierarchical system that can be expanded or collapsed based on the size and complexity of the incident. The ICS was developed to allow staff from different government agencies and organizations to work

effectively and efficiently together to respond to a natural disaster. Participants essentially check their individual agency identities at the door and participate as members of the ICS organization, dedicated to responding to a particular incident. The system's success relies on participants understanding their role, a clear chain of command and communication, managers having an appropriate span of control, and a standardized process for identifying and communicating objectives, strategies, tasks and deadlines. Because of its proven effectiveness, the ICS has recently been integrated into the National Incident Management System (NIMS). For more information about the principles and features of the ICS go to Lessons 2 and 3 at <http://emilms.fema.gov/ICS100G/index.htm>. To learn more about the integration of ICS into NIMS, please visit www.fema.gov/emergency/nims. An example of how the ICS staff organization scheme has been applied to an AIS rapid response in California is provided in Chart 2.

Optimal use of this system requires that participants be trained in advance per Section IV (Planning) of this document. The Planning Section also discusses the need to develop the finer details of the procedure, the lists and directories that are referred to in the procedure, and the designation of alternates. This last item ensures that none of the positions described in the procedure are ever vacant.

The procedure that will be followed for a given incident may follow the generic plan provided below or be based on a species-specific rapid response plan approved by the participating agencies. As species-specific plans are developed and approved, staff that have been identified as potential responders will be notified of their approval and location on the Internet. Basic information about each species specific plan will be incorporated into AIS rapid response training.

Chart 1. DRAFT General Procedure for Rapid Response Following Detection of New Aquatic Invasive Species Infestation



Glossary

AIS
Aquatic Invasive Species

DFG-ISP
Department of Fish and Game Invasive Species Program

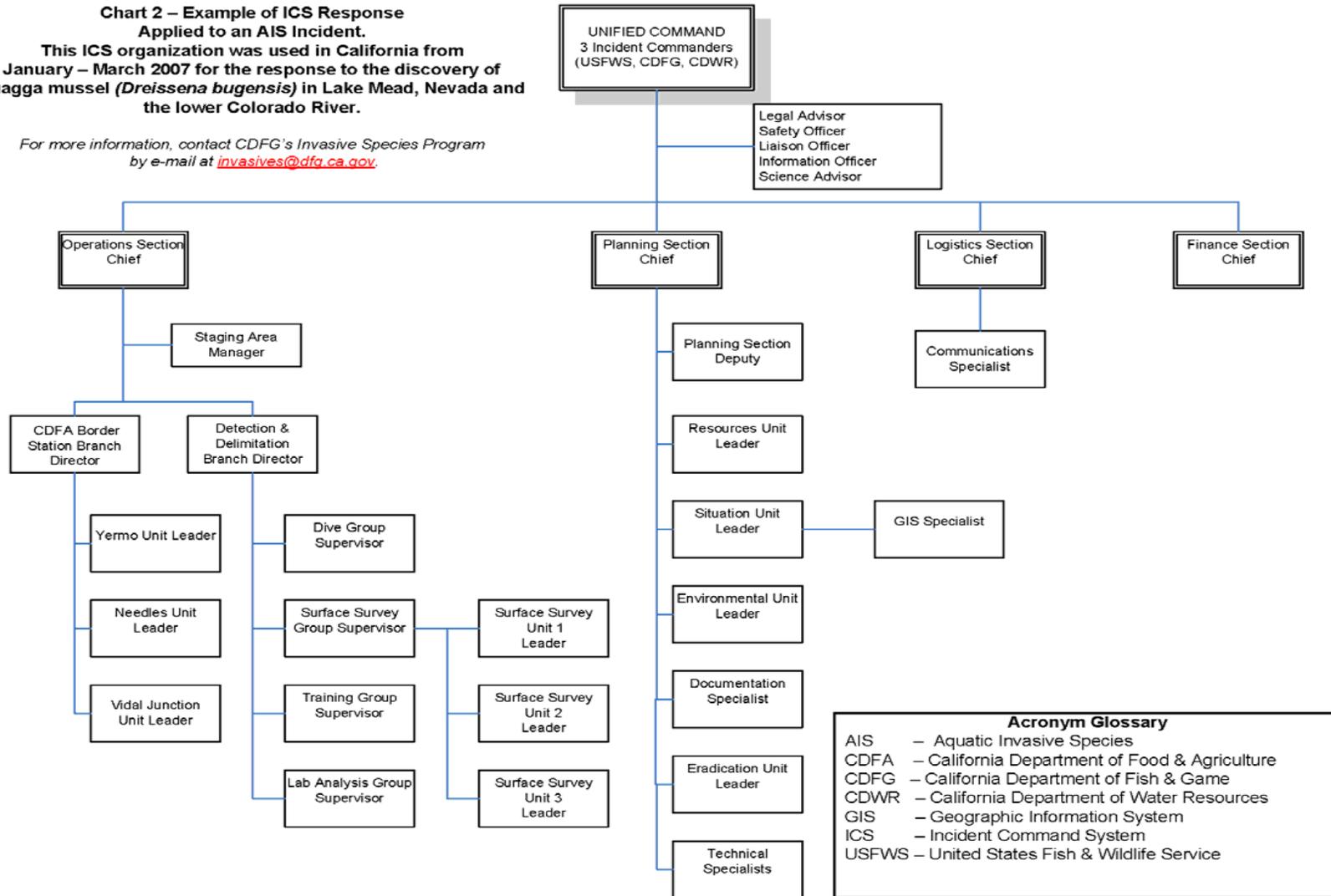
ICS
Incident Command System

Notes

1. See Chart 2 for an example of an ICS organization.
2. Outreach and training are used throughout the response, as necessary.

Chart 2 – Example of ICS Response Applied to an AIS Incident.
 This ICS organization was used in California from January – March 2007 for the response to the discovery of quagga mussel (*Dreissena bugensis*) in Lake Mead, Nevada and the lower Colorado River.

For more information, contact CDFG's Invasive Species Program by e-mail at invasives@dfa.ca.gov.



17-STEP RAPID RESPONSE PROCEDURE

Step 1. Identify species and notify authorities

- a. Sighting Report: There are three ways in which DFG is likely to receive a report of an AIS sighting.
 1. Either a sighting is reported to DFG via a hotline phone number or e-mail address (Invasives@dfg.ca.gov), and catalogued on *RR Form 1: Suspect AIS Sighting Report* (see Section V).
 2. Staff from another agency or cooperator discovers the AIS and submits the collected information directly to DFG's Invasive Species Program staff.
 3. The initial report is made to one of the federal invasive species reporting systems (e.g. "United States Geological Survey Nonindigenous Aquatic Species Alert System" or the "100th Meridian Initiative") which in turn will alert DFG.
- b. Sighting Transmittal: This initial information is transmitted to the DFG Invasive Species Coordinator (ISC). If there is uncertainty about the identification of the species, the Invasive Species Program staff will work with taxonomic experts to resolve the issue.
- c. For the purpose of documentation, and to assist making a determination of how to proceed following the initial report, the more detailed *RR Form 2: AIS Alert Report* (see Section V) should be completed.
- d. Negative ID: If the identification is negative for AIS no further action is necessary.
- e. Indefinite ID and/or level of threat: If uncertainty remains after initial fact-finding, the DFG Invasive Species staff should continue to work with experts from cooperating agencies and research institutions to determine the status of the species reported and the level of threat.
- f. Positive ID with a high level of threat: If the discovered organism is invasive and in the presence of vectors that could cause its spread to uninfested areas, DFG Invasive Species Coordinator will consult with DFG executive level staff to determine if an ICS response is appropriate.
 1. If the identification is positive, the DFG Invasive Species staff will ensure that a report is sent to the United States Geological Survey Nonindigenous Aquatic Species Alert System (<http://nas.er.usgs.gov/SightingReport.asp>). During the response, the alert system should receive updates on any additional locations of the AIS that are found.
 2. Fill out an Incident Brief Form (ICS Form 201).
 3. ICS forms are available at:
http://training.fema.gov/EMIWeb/IS/ICSResource/ICSResCntr_Forms.htm

Step 2. Activate command-level participants

- a. Incident Command Staff: The executive level DFG staff will work with the Invasive Species Coordinator and executive level staff of cooperating agencies to identify the Incident Command staff. They can utilize the Rapid Response Personnel Directory discussed in the Planning Section of this document.

1. The Incident Commander is the overall supervisor and coordinator for the incident. A detailed description of the responsibilities of an Incident Commander and the other Incident Command officers and General Staff positions, can be found in Lessons 3 and 4 at <http://emilms.fema.gov/ICS100G/index.htm>.
2. Executive level staff and the ISC will decide to pursue a single command response, with one Incident Commander, or a unified command response, with multiple Incident Commanders working as a team. A Unified Command approach is designed to be used in multi-agency or multi-jurisdiction responses.
 - b. Initial Unified Command Meeting: If a unified command approach is used the Incident Commanders in the Unified Command should meet to discuss and concur on important issues prior to starting the first operational period planning meetings.

Step 3. Implement the ICS Planning Cycle

- a. Begin to utilize the ICS planning cycle to document the current status of the response, identify objectives, strategies, specific task assignments and operational period. See http://www.uscg.mil/hq/g-m/mor/media/Chapter_3.pdf for a description of the ICS Planning Cycle.
 1. During every ICS planning cycle, an Incident Action Plan is developed for the following operational period. It contains objectives, safety measures, staff contact information, status of the incident and assignments for each organizational element that will be active during the next operational period. The plan must be approved by the Incident Commander(s).
 - a) The plan is comprised of standard ICS forms that are available in electronic form. Once the initial set of forms is completed, the Incident Action Plan can rapidly be revised and updated.

Step 4. Develop the Organization

- a. Command Post: Establish a command post capable of supporting the space, logistic, communication and other technology needs for managing the operation. It may or may not be a high priority to have the command post located close to the infested site, based on the characteristics of a particular incident. Potential command posts will be listed in the AIS Rapid Response Resource Directory discussed in the Planning Section of this document.
- b. Logistics and Finance: The Logistic and Finance Section Chiefs will establish the fundamental tools and means to run the organization, such as setting up the check-in routine, necessary ICS forms, communication services, spending authorizations, and tracking of resources.
- c. Assemble Organizational Elements: Using the ICS system, develop an organization that is suitable for the size and complexity of the incident.
 1. Directory of Approved Staff: To staff the organizational elements (e.g. sections, branches, units) the Incident Command and upper level General Staff will utilize (but are not limited to) staff directories of people approved to be assigned to rapid response efforts.
 2. ICS training materials suggest that "it is better to initially overestimate the need for a larger organization than to underestimate it, as it is always possible to downsize the organization." (National Wildfire Coordinating Group, 1994, p.3-19).

3. Logistics Section staff will utilize the Resource Directory discussed in the Planning Section of this document in their effort to procure the necessary equipment and supplies among cooperating agencies and organizations during a rapid response procedure.
- d. Consider the need to assemble a science advisory panel that may include experts outside of the ICS organization to provide input on such topics as AIS biology, sampling techniques, eradication or control measures.

Step 5. Safety Plan

- a. The standard ICS organization includes a Safety Officer who reports to the Incident Commander/Unified Command. One of the duties of the Safety Officer is to develop a Safety and Health Plan that assesses potentially hazardous situations that could exist throughout the operation for responders and the public, and outlines the safety measures that should be taken.

Step 6. Outreach

- a. Outreach Plan: The incident's Information Officer develops an Outreach Plan for the incident that addresses short and long-term proactive communication objectives and strategies to be employed with relevant groups such as the media, government agency representatives outside of the ICS response, stakeholders, interest and community groups and the general public.
 1. Develop policy with the Incident Commander(s) and the Liaison Officer regarding protocols for disseminating information.
 2. Besides disseminating information the outreach plan should address obtaining input from stakeholder groups and other interested individuals.
- b. The Media: Typically, the Information Officer is assigned to be the contact person for inquiries from the media.
 1. Typical tasks include preparation of press releases, briefings, public meetings, etc.
 2. The Information Officer reports to the Incident Commander.
- c. Government Agencies: Typically, a Liaison Officer is assigned to be the point of contact for inquiries from government agencies that have an interest in the response.
 1. The Liaison Officer provides relevant updates on the response to representatives from these agencies.
 2. The Liaison Officer reports to the Incident Commander.
- d. Stakeholder and Interest Groups: Outreach to these groups can be crucial, especially if their activities can result in spread of the AIS. Outreach to non-governmental groups needs to be assigned to the Information Officer or the Liaison Officer. A large stakeholder group for a large incident may warrant their own Assistant Liaison Officer or Assistant Information Officer to maximize cooperation from this group and be aware of concerns they may have.
- e. General Public: Assign who will be responsible for responding to inquiries from individual members of the public. Determine whether it is advisable to establish and publicize a toll-free call-in number for the incident.

Step 7. Training

- a. Develop a Training Plan: There is often a need to establish a training branch within the ICS. As the incident begins to unfold, the Training Director will be responsible for working with managerial level staff to assess and find appropriate means to provide the types of training that are needed, both for staff within the ICS and for cooperating agencies, organizations and volunteers.
 - 1. A training manual should be developed that contains any specialized protocols and associated training materials (e.g. survey or decontamination protocols).

Step 8. Regulatory Compliance

- a. The Planning Section is typically responsible for addressing regulatory compliance with environmental laws, with input from the Legal Specialist assigned to the incident. The issues that are most likely to arise are related to water quality and effects on state or federally listed species during survey or control activities.

Step 9. Containment Actions

- a. Take action to prevent the spread of the AIS. Examples of containment actions that might be taken include:
 - 1. Inspections: Working with public and private managers of infested and potentially infested waterbodies and waterways, locate and inspect potentially contaminated facilities, shorelines, boats, vehicles and equipment to the extent possible. Prioritize a list of potential sites that should be inspected. Some of this work is part of the rapid assessment described below.
 - a) Survey boaters about previous and subsequent waterways visited and provide them with information about the AIS problem.
 - b) If regulations allow, require, or otherwise, request that aquatic plant and animal material be removed from the watercraft, motor and trailer and for any remaining water to be drained.
 - c) Request that boats and equipment be rinsed with high pressure or hot water and dried before launching. The time needed for drying is species specific.
 - d) Boats that are found to be contaminated with a legally restricted species per F&G Code Sec. 671 cannot be launched until they are certified by DFG to be decontaminated.
- b. Introductions from Out-of-State: Coordinate with California Department of Food and Agriculture's Border Protection Station Program, federal, and other state and national agencies if the introduction is known to have come from out of state or has potential to have come from out of state.
- c. Prevent Spread from California: Coordinate with federal and state agencies on preventing spread from California into other states (especially states that border CA), Canada or Mexico.
- d. Temporarily quarantine body(ies) of water that contain subject AIS.
 - 1. Establish a quarantine utilizing one of the methods discussed in legal authority section.
 - 2. In addition to sites known to contain the subject AIS, consider whether it is appropriate to quarantine areas where the AIS may have been introduced.

Step 10. Rapid Assessment

- a. Extent of the Infestation: Get a qualitative “snapshot” of the extent of the infestation and identify potential vectors for spreading the AIS.
 1. Planning and Operations Section staff can work together to identify short vs. longer-term information needs and plan how various types of information should be gathered.
 - a) Samples may need to be collected for gathering basic demographic information or more in-depth taxonomic work. Establish protocol for collecting, transporting, and storing samples. Develop appropriate permits for possession and transportation of specimens.
 - b) In addition to noting the presence or absence of the AIS, consider whether it's appropriate to systematically get some basic information about the habitat at this point, collect samples of substrate or water, etc.
 - c) Determine whether there are known occurrences of, or potential habitat for, state or federally listed species in the area that needs to be surveyed, and whether surveys may require consultation with DFG, the U.S. Fish and Wildlife Service or NOAA Fisheries.
- b. Data collection is typically done by the Operations Section of the ICS, with the Logistics and Finance Sections providing assistance with the procurement of equipment, vehicles, travel, etc.
- c. Impacted Parties: Obtain contact information for pertinent landowners, land managers, holders of water rights, water users and jurisdiction over the body(ies) of water involved. If it is necessary to enter private property to conduct rapid response work, assign an ICS member to obtain permission to enter.

Step 11. Plan Eradication or Control Measures

- a. If appropriate, develop a plan to eradicate the AIS from CA or a control plan to prevent the spread of the AIS. It may not be feasible to finalize the plan during the rapid or ICS phase of the response. Some planning may occur after the ICS is demobilized.
 1. During the assessment phase of the response, the Planning Section can gather and review information on potential eradication or control techniques and confer with experts (Step 4D).
 2. As information is gained from the rapid assessment, and possibly from subsequent detailed sampling, a more refined version of an eradication or control plan can be prepared, discussing the specific measurable objectives, locations and methods for eradication or control, methods for evaluating the effectiveness of the plan, and the potential costs, benefits and impacts.
 3. Conduct any regulatory processes and obtain any regulatory permits that may be necessary prior to implementation of the plan.

Step 12. Implement the Eradication or Control Plan

- a. Implementation of the eradication or control plan may place during the “rapid” part of a response; however, if this is not the case, eradication or control measures might be implemented during a later “post –ICS” phase of the response.
- b. Document implementation of the eradication or control plan. Note any deviations from the plan and why those occurred.

Step 13. Prevent Reinfestation

- a. Develop specific recommendations for actions that can be recommended to prevent reinfestation such as:
 1. Long-term monitoring
 2. Continued outreach and education
 3. Partnerships with business and interest groups
 4. Strengthening relevant regulations
 5. Identify staffing needs
 6. Identify research needs
- b. Ensure the potential for introduction from nearby commercial operations (shipping, bait shops, aquaculture, aquarium shops) is removed or minimized to the extent possible.

Step 14. Prepare Demobilization Plan

- a. During the response, the Planning Section is responsible for preparation of a Demobilization Plan and having it approved by the Incident Commander(s). The purpose of the Demobilization Plan is to assure that all participants understand their role in an orderly, safe and efficient demobilization of incident resources as rapid response procedures are completed. Equipment and supplies must be returned to appropriate locations, time and cost accounting reports must be completed within required timeframes, and any other required progress and final reports must be prepared and submitted.

Step 15. Monitor the outcome of the Rapid Response

- a. Evaluate Eradication or Control Efficacy: If eradication or control actions were taken during the response, monitor and evaluate the efficacy of the treatment(s) used and conduct environmental monitoring that may be necessary to meet regulatory compliance requirements. Prepare a monitoring report and submit a copy to the ISC. If the control or eradication measures require months or years to implement, these evaluation reports may take the form of periodic progress reports.
 1. If the treatments were not successful or an acceptable level of progress is not being achieved, evaluate the potential for remedial measures to improve the results. If there is a strong possibility for improvement, propose possible remedial actions as part of the monitoring report.
- b. If eradication or control measures were not taken, there may be a decision to conduct monitoring of the AIS population and provide monitoring reports to the DFG Invasive Species Program.

Step 16. Undertake remedial actions and long-term follow up

- a. Remedial Action Approval: If there is efficacy monitoring prior to the demobilization of the incident and remedial actions are recommended, the Incident Commander(s) can approve the implementation of a remedial action plan and utilize the assembled rapid response personnel, assuming any environmental regulatory and/or fiscal issues are addressed.
- b. Remedial Action Monitoring: Remedial actions and their results will require subsequent monitoring.
- c. Follow-Up Actions: If longer-term actions are necessary, the Planning Section, with input from other rapid response personnel and outside expert input as necessary, will develop a follow-up plan that will be submitted to the DFG Invasive Species Program.

Step 17. Implement the Demobilization Plan

- a. Implement the demobilization plan described in Step 14. The work will be carried out by the Incident Teams and Specialists with oversight and coordination from the Incident Command Staff. Reports will be submitted to the ISC for approval and appropriate distribution.

IV. PLANNING FOR RAPID RESPONSE

This section suggests 11 basic task areas necessary to plan for rapid response and completion of this plan.

Task 1. Collaborate to complete plan

Representatives from public agencies and other organizations that are currently involved in rapid response work, or likely to be involved in the foreseeable future, should collaborate to finalize the Rapid Response Plan (see Task 4). The goal is to have a plan that can be the basis for interagency agreements (Task 2). Note that not every item in Task 4 needs to be complete in order to have a plan that supports such agreements. This group could also prioritize and carry out parts of additional planning tasks listed below. The collaboration necessary to carry out the tasks in this section could occur through a technical advisory panel to the CAAIST or AISWG (collaborative groups described in the CAISMP), through the California Biodiversity Council (CBC) Rapid Response Working Group, or through executive or upper management staff of cooperating agencies assigning staff to an interagency Rapid Response Planning Team.

Task 2. Enter into cooperative agreements

DFG Invasive Species Program staff will work with cooperating agencies and organizations to produce a list of entities that should be invited to sign Memoranda of Understanding, Implementation Agreements or similar instruments to cooperate on rapid response to AIS. Existing information in the CAISMP and information collected by CBC Rapid Response Working Group will be used, among other sources, to generate this list. The proposed list and a conceptual outline for these agreements will be presented to CBC and/or directly to relevant agency executives.

Task 3. Secure funding

This Plan cannot be implemented without adequate, stable and dedicated funding. Agencies signatory to the Rapid Response agreement(s) should coordinate efforts to pursue funding options for Rapid Response program development, training and implementation.

Organizations and industries that have a vested interest in successful early detection and rapid response systems could participate in the development of funding sources.

- a. Funding Analysis: Consider the following types of funding sources:
 1. A permanent funding source(s) maintained solely for rapid response actions. Without this, rapid response may not occur or may only occur by redirecting funds on short notice from other important programs.
 2. A user-fee system based on vectors for AIS introductions. This would be similar in concept to fees paid by the shipping industry for ballast water inspections or fees paid by the petroleum industry for an oil spill response program. Methods used by states that already have dedicated funding for rapid response can be emulated.
 3. Private/public partnerships for supporting rapid response efforts in the form of equipment, supplies, personnel or funding.
 4. One-time grants for specific planning or research projects related to rapid response.
- b. Taxonomy Funds: Develop funding for taxonomic work to identify potential AIS specimens. In some cases, this will include genetic analysis (e.g. to determine presence or absence of microscopic larvae of AIS species, or help determine the origin of an introduction). Expert taxonomic work will bolster confidence that subsequent management decisions are based on solid information. There should also be funding to maintain specimens. The proper maintenance and documentation of specimens is especially important in cases where infestations are the subject of law enforcement actions and may also be beneficial for future AIS identification needs and research.
- c. Professional Cost Analysis: Consider whether a detailed, professional analysis of rapid response costs to support funding requests is necessary (Task 10b).
- d. Funding Development: Consider using funding for development purposes (i.e. grant writing).

Task 4. Finalize the Rapid Response Plan

Work that needs to be done to finalize the Rapid Response Plan includes:

- a. Implementation Criteria: Develop the process and criteria for the State to use in determining the course of action to take for any new AIS introductions. Circulate for peer review.
- b. Likely Species & Scenarios: Identify likely species and/or early detection scenarios for AIS. Run these scenarios through the criteria developed for Task 4a to fine-tune the criteria.
- c. Agency Preparation: Develop information needed to help cooperating agencies designate and train, in advance, potential responders to AIS introductions.
- d. Alternate Staff: Develop a procedure to designate and prepare potential alternate staff. This could avoid gaps in getting work done and minimize managerial time spent searching for substitutes during a response.
- e. Personnel Directory: Develop a statewide Rapid Response Personnel Directory. These people could be called upon to participate during rapid response activities, and into an ICS response. Ideally the Directory should include staff that represent the full spectrum of knowledge and skills that might be necessary during rapid response activities (e.g. ICS

implementation, logistics, finance, legal and various technical experts). The development of this list and staff participation in Rapid Response planning and training will likely require support of executive level staff from the cooperating agencies.

- f. Resource Directory: Develop and maintain a directory among cooperating agencies for equipment, operations centers, supply sources and associated contact people so that resources can be mobilized as quickly as possible during a response.
- g. Taxonomic Experts: A list of taxonomic experts and protocols for requesting and using their services needs to be developed and periodically reviewed and updated. This would be a list of experts who have agreed to identify specimens for AIS Rapid Response efforts and appropriately preserve and catalog them.
- h. Local Assistance Protocol: Develop a protocol for responding to a private entity or local government agency that wants to conduct a rapid response under its own direction but requests assistance or permits from one or more agencies signatory to the statewide Rapid Response Plan. Include this protocol in the rapid response training program.
- i. Notification List: Develop a list of whom, outside of those directly involved, needs to be notified when rapid response procedures are being planned and implemented.
- j. Database Compatibility: Consider whether information should be collected in a particular manner in order to be compatible with existing AIS databases. For example, the North American Weed Management Association has a list of required elements for weed mapping projects (www.nawma.org).

Task 5. Streamline permit processes for rapid response

DFG Invasive Species Program staff will coordinate with staff from relevant agencies to investigate and pursue possibilities for streamlining the regulatory permit processes that might be required for rapid response measures. General measures or best management practices necessary to comply with streamlined permitting can be incorporated into the Rapid Response Plan.

Task 6. Revise the Rapid Response Plan

- a. Incorporate New Information: Periodically revise the Plan and incorporate anything learned by evaluating the Plan's effectiveness and consulting current scientific research and related technological developments. Revisions may also be necessary due to changes in funding, agency restructuring and environmental regulations. The interagency agreement(s) to cooperate on rapid response should include a procedure for making revisions to the Plan.
- b. Notification of Plan Changes: DFG Invasive Species Coordinator should ensure that adopted changes to the Plan are circulated to people listed in the Rapid Response Personnel Directory and other appropriate staff among the cooperating agencies and organizations. Changes should be addressed in training activities.
- c. Update Directories: DFG Invasive Species Program staff, with assistance and input from cooperating agencies and organizations, will be responsible for the periodic update and circulation of the Rapid Response Personnel Directory, the Rapid Response Resource Directory and the list of taxonomic experts.

Task 7. Develop species- or location-specific rapid response plans

Identify and prioritize certain species, groups of species or certain locations for the development of specific rapid response plans. Detailed technical information can allow this type of response plan to be implemented more efficiently than a generic response plan. The development of species- or location-specific rapid response plans is called for in Action 4A3 of the CAISMP. The process of prioritizing which species warrant the development of rapid response plans will also help guide the development of outreach materials for early detection efforts.

Task 8. Train employees, participants and team members

- a. Training Program: Agencies that agree to cooperate on AIS rapid response need to participate in the development of a training program and train the employees likely to be involved in rapid response activities. Potential rapid response participants need to be familiar with the Rapid Response Plan, Incident Command System (ICS), and may need specialized training related to their likely duties during a rapid response. ICS training is available on-line at: <http://training.fema.gov/IS/>.

There may be a need to develop supplemental training materials and presentations for information specific to California, AIS or other topics.

- b. Drills: Ensure that training includes AIS rapid response drills using a variety of scenarios and locations around the state. This will also assist in fine-tuning the Rapid Response Plan.

Task 9. Conduct education and outreach

- a. Outreach Planning: Outreach specialists from participating agencies and organizations should develop a plan of potential methods and protocols for conducting outreach to local communities, interest groups and the media during rapid response procedures. This could include sharing contact information for key groups such as boaters, anglers and marina owners.
- b. Disruption of Regular Work: Within the cooperating agencies, supervisors of employees who are in the Rapid Response Personnel Directory should be made aware that rapid response work can supersede other projects on very short notice. Supervisors and employees who are on rapid response teams could discuss in advance how they plan to handle this potential source of disruption.

Task 10. Conduct research necessary for improved rapid response

- a. Response Research: Academic institutions, government agencies and other organizations that agree to cooperate on rapid response should work together through various AIS working groups, professional and environmental organizations and commercial interests to promote research that can specifically improve or promote rapid response efforts.
- b. Cost Research: Research the costs of rapid response, possible funding mechanisms (Task 3) and, if feasible, study the environmental and economic benefits and costs of conducting rapid response efforts versus not conducting rapid response. This may help governments decide how much to invest in rapid response measures.

Task 11. Develop interim rapid response protocols

This section addresses the question: What steps can be taken to prepare to implement a rapid response effort while a formal plan is going through the review and approval processes?

- a. Memorandum of Understanding (MOU): The Directors of the appropriate agencies could sign an interim MOU directing their staff to participate in rapid response planning and implementation if a new AIS introduction occurs prior to the approval of the final plan.
- b. Interim Funding: Management staff could identify and pursue interim funding sources for implementing a rapid response program.
- c. Interim Strategy: Management level staff from cooperating agencies could informally agree upon an interim strategy regarding roles and responsibilities should an AIS introduction occur.
- d. Permitting: Management level staff from cooperating agencies could discuss how, in the absence of a formal streamlined permitting process, their staff could work within the existing regulatory permit programs to facilitate a rapid response operation and direct staff to follow through on these interim measures.
- e. Employee Assignment: Management level staff of cooperating agencies could assign employees to an interim core rapid response team or working group. This team could participate in some advance preparation and planning. In the event of a rapid response, this team would need to be augmented by additional staff based on the location of the response and the necessary areas of expertise.

V: Supporting Materials

Rapid Response Form 1. Suspect AIS Sighting Report

The reporter may not be able to provide all of the information requested below, but please fill in as many of the information fields as possible.

Report Tracking Number: _____ Date of Sighting: _____

Reporter's First and Last Name: _____

Reporter's Phone Numbers: Home: _____ Work: _____

Cell: _____

Reporter's E-Mail Address: _____

Reporter's Mailing Address:

Type of Organism (as specific a descriptive label as possible (e.g. submerged plant, shellfish, etc.):

Description of size, color, shape and other distinguishing characteristics:

Approximate number of individuals or area they occupy:

Location of sighting:

Directions and description of nearby landmarks:

Were any photographs taken or specimens collected? If so, where can they be obtained?

Landowner or Land Manager:

Possible Source of Introduction:

Name and Contact Information of Person Filling Out This Form:

Rapid Response Form 2. AIS Alert Report 1

To be filled out by Species Identification Team member following up on a preliminary report of a possible AIS sighting (Form 1). The AIS Report will be expanded to two pages in the final draft to reduce the crowding on this form.

Species Name: _____ Report Tracking # _____
Name of Person Filling out Form: _____ Phone Number(s): _____
Agency: _____
Address: _____ E-mail address: _____

Reporter's Name: _____
Reporter's Phone Number(s): _____
Reporter's e-mail: _____

Date of Pest Sighting: _____
If the identification was verified by expert, who provided the verification?
Verifier's phone number(s): _____ E-mail: _____

Location of voucher specimens: _____

Sighting Location (if possible attach a map showing the location):
County: _____ Body of water: _____

Landowner/Manager: _____
Describe location
(Relationship to nearby road intersection, pier, mile marker, buoy, other landmarks)

If possible, please provide map information (You choose the system):

T____ R____ Sec____, _____ 1/4 of _____ 1/4, Meridian: H__ M__ S__
T____ R____ Sec____, _____ 1/4 of _____ 1/4, Meridian: H__ M__ S__

Quad Name: _____ Source of Coordinates (GPS, topo map & type): _____
GPS Make and Model: _____ Horizontal Accuracy _____ meters/feet

Datum: NAD27____ NAD83____ WGS84____
Coord. System Zone 10 ____ Zone 11 ____ or Geographic Latitude/Longitude _____

Describe pest species population (approximate number of individuals or stems, area they occupy)

Describe any evidence of reproduction (flowering, juvenile animals, egg masses, etc.)

Describe habitat: (e.g. plant community, associated plant species, host species, water depth, distance from bank, substrate characteristics (e.g. gravel, large rocks, silt, sand), etc.)

Photographs can be accessed at:

1 Based on California Department of Fish & Game, California Natural Diversity Data Base, "Native Species Field Survey Form" and the "Maui County Report A Pest Online Report Form," Maui County, HI.

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APPENDICES B-D

Introductory Notes

These appendices provide a detailed description of the primary federal and state laws, regulations and public policies that empower and direct different government agencies to manage AIS in California. They also describe the primary activities of government agencies – state, federal and regional – involved in AIS management, as well as most of the major committees and boards set up to coordinate and oversee such activities. These details are provided to support and expand on the information contained in the Management Framework provided in Chapter 4 and the Summary of Laws provided in Chapter 5 of this plan (as such, there is some repetition of information). While these appendices attempt to be comprehensive, there is inadequate space to present every single AIS program, law or activity in the state and nation. Through the web links provided below and further information in the appendices, more details on legal authorities and AIS stakeholders is available to all interested parties. A key to the acronyms used in these appendices can be found in the Acronym Glossary in the introductory pages of this plan. (*Note: Some laws and policies refer to ANS, aquatic nuisance species, rather than AIS, aquatic invasive species.*)

APPENDIX B: FEDERAL AUTHORITIES, LEGISLATION & AGENCIES

FEDERAL AUTHORITIES

No single federal agency has comprehensive authority for all aspects of aquatic invasive species management. Federal agencies with regulatory authority over the introduction and transport of aquatic species that may be invasive or noxious include the U.S. Department of Agriculture Animal Plant Health Inspection Service, the U.S. Department of Agriculture Agricultural Marketing Service, the U.S. Fish and Wildlife Service (USFWS), the U.S. Department of Commerce (DOC), and the U.S. Coast Guard (USCG). Many other agencies have programs and responsibilities that address components of AIS, such as importation, interstate transport, exclusion, control and eradication.

The primary federal authorities for managing and regulating AIS derive from the National Environmental Policy Act, the Nonindigenous Aquatic Nuisance Prevention and Control Act (NANPCA, 1990), the National Invasive Species Act (NISA, 1996), the Lacey Act, the Plant Pest Act, the Federal Noxious Weed Act, and the Endangered Species Act. An Executive Order signed by President William J. Clinton on February 3, 1999 expanded federal efforts to address AIS. The order created a National Invasive Species Council charged with developing a comprehensive plan to minimize the economic, ecological and human health impacts of invasive species.

Brief descriptions of the President's Executive Order, NANPCA and NISA are provided below, followed by an explanation of how federal activities are now coordinated through the national Aquatic Nuisance Species Task Force (ANSTF) and the National Invasive Species Council (NISC), and by descriptions of some of the earlier acts and laws still enforced in AIS management.

Primary Federal AIS Authorities

1990 – Nonindigenous Aquatic Nuisance Prevention and Control Act (NANPCA; Title I of P. No.101-646, 16 U.S.C. 4701 et seq.)

<http://www.anstaskforce.gov/default.php>

The Nonindigenous Aquatic Nuisance Prevention and Control Act of 1990 (NANPCA) established a federal program to prevent the introduction and control the spread of introduced aquatic nuisance species. The act provides an institutional framework that promotes and coordinates research, develops and applies prevention and control strategies, establishes national priorities, educates and informs citizens, and coordinates public programs. The act calls upon states to develop and implement comprehensive state management plans to prevent introduction and control the spread of aquatic nuisance species (ANS). Section 1002 of NANPCA outlines five objectives of the law, as follows:

1. Prevent further unintentional introductions of nonindigenous aquatic species;
2. Coordinate federally funded research, control efforts, and information dissemination;
3. Develop and carry out environmentally sound control methods to prevent, monitor and control unintentional introductions;
4. Understand and minimize economic and ecological damage; and
5. Establish a program of research and technology development to assist state governments.

Section 1201 of the act established the national ANSTF, co-chaired by the USFWS and the National Oceanic and Atmospheric Administration. The Task Force is charged with coordinating governmental efforts related to ANS prevention and control. The ANSTF consists of 10 federal agency representatives and 12 ex officio members representing nonfederal governmental agencies (see Other AIS Interests, Appendix D).

**1996 – National Invasive Species Act
(NISA; P. No.104-332)**

In 1996, the National Invasive Species Act (NISA) amended the NANPCA of 1990 to mandate ballast water exchange for vessels entering the Great Lakes and to implement voluntary ballast water exchange guidelines for all vessels with ballast on board that enter U.S. waters from outside the U.S. Exclusive Economic Zone (U.S. EEZ). Though the act did not make exchange mandatory, it did require all vessels to submit a report form to the USCG documenting specific ballast water management practices. It also authorized the USCG to toughen requirements if compliance proved unsatisfactory, which it did in 2004 (see below). NISA authorized funding for research on aquatic nuisance species prevention and control in Chesapeake Bay, the Gulf of Mexico, the Pacific coast, the Atlantic coast, and the San Francisco Bay-Delta Estuary. In addition, NISA required a ballast water management program to demonstrate technologies and practices to prevent ANS from being introduced into and spread through ballast water in U.S. waters. It modified both the composition and research priorities of the ANSTF and requirements for the zebra mussel demonstration program.

**1999 – Executive Order 13112
(64 Fed. Reg. 6183)**

<http://www.invasivespeciesinfo.gov/council/main.shtml>

President William J. Clinton signed Executive Order 13112 on Invasive Species on February 3, 1999. The order seeks to prevent the introduction of invasive species, provide for their control and minimize their impacts through improved coordination of federal agency efforts under a National Invasive Species Management Plan developed by the newly created National Invasive Species Council (NISC). The order directs all federal agencies to address invasive species concerns, as well as to refrain from actions likely to increase invasive species problems.

The NISC has three co-chairs: the secretaries of Agriculture, Commerce, and the Interior. Members also include the secretaries of State, Defense, Homeland Security, Treasury, Transportation and Health and Human Services, as well as the administrators of USEPA, the U.S. Agency for International Development, the U.S Trade Representative and the National Aeronautics and Space Administration. The NISC released the first National Invasive Species Management Plan in 2001. The NISC is currently working to establish federal and non-federal task teams to implement the plan's action items.

The NISC actively works with the Invasive Species Advisory Committee (ISAC), also established under the order. The ISAC is composed of stakeholder representatives from state governments, industry, conservation groups, academia and other interests. Its role is to advise the federal government on the issue of invasive species.

To help coordinate the work of the NISC and the ANSTF, the Department of Commerce (DOC) Policy Liaison to the NISC also serves as the DOC representative to the ANSTF. In addition, NISC and the ANSTF have formed joint working groups on each of the following topics: pathways, risk analysis and screening.

The ANSTF and the NISC are similar in that they perform coordinating functions but differ in their responsibilities: the NISC addresses all invasive species, while the ANSTF focuses on aquatic invasive species. Although many of the same principles apply to managing aquatic and terrestrial invasive species, many management issues are unique to the aquatic environment and need to be addressed separately.

1993-2005 – Coast Guard Regulations under NISA (33 CFR 151)

The USCG has promulgated a number of ballast water management regulations based on the authority given to it by NANPCA in 1990 and NISA in 1996. As directed by NANPCA, in 1993, the USCG implemented regulations requiring vessels entering the Great Lakes and the Hudson River to conduct ballast water management after operating outside the U.S. EEZ.

To comply with the NISA, the USCG established regulations and guidelines to control the introduction of ANS via ballast water discharges in U.S. waters other than the Great Lakes. Compliance with the resulting voluntary ballast management and mandatory reporting program was only 30%, according to a 2002 Report to Congress. Therefore, under the authority of NISA, the USCG established mandatory ballast water management requirements and penalties for non-compliance. The mandatory program requires ships to use one of three ballast water management methods: 1) retaining ballast water on board, 2) conducting a mid-ocean exchange, and/or 3) using an approved ballast water treatment method. All vessels are required to submit ballast water management reports (failure to submit a report can now result in penalties). These mandatory regulations came into effect on September 27, 2004. Federal regulations also require vessels to maintain a ballast water management plan that is specific for that vessel and assigns responsibility to the master or appropriate official to understand and execute the ballast water management strategy for that vessel.

Under NANPCA/NISA, states are specifically permitted to regulate ballast water on ships. Several states have elected to do so to various degrees. In addition to reporting requirements, California, Oregon and Washington have ballast water exchange requirements and California will soon specify a ballast water discharge standard (see California Authorities section).

Other Federal Authorities

Animal Damage Control Act (1931)

<http://www.aphis.usda.gov/>

Under the Animal Damage Control Act, the U.S. Department of Agriculture (USDA), Animal and Plant Health Inspection Service has authority to control wildlife damage on federal, state, or private land, including damage from invasive species. The act protects field crops, vegetables, fruits, nuts, horticultural crops and commercial forests; freshwater aquaculture ponds and marine species cultivation areas; livestock on public and private range and in feedlots; public and private buildings and facilities; civilian and military aircraft; and public health.

Animal Health Protection Act (2002)

(7 U.S.C Sec. 8301, et seq.)

<http://www.aphis.usda.gov/>

The Animal Health Protection Act provides a flexible statutory framework for protecting domestic livestock from foreign pests and diseases. This act authorizes the USDA to promulgate regulations and take measures to prevent the introduction and dissemination of pests and diseases of livestock. The scope of such regulatory authority extends to the movement of all animals, domestic and wild, except humans. The fact that a pest or disease primarily affects animals other than livestock, including humans, does not limit USDA's authority to regulate a species, so long as it carries a pest or disease of livestock. Further, the act defines "livestock" to mean all farm-raised animals, clarifying the USDA's authority to conduct animal health protection activities in connection with farm-raised aquatic animals.

Clean Water Act

<http://www.epa.gov/r5water/cwa.htm>

<http://unds.bah.com/default.htm>

Various sections of the Clean Water Act (CWA) regulate discharges of pollutants (such as AIS and ballast water) and fill material to waters of the United States. Section 402 of the act authorizes the National Pollutant Discharge Elimination System (NPDES), a permit program intended to reduce and eliminate the discharge of pollutants from point sources that threaten to impair beneficial uses of water bodies. The act defines point sources to include vessels (Section 502(14)) and prohibits all point source discharges of pollutants into U.S. waters unless a permit has been issued either under Section 402 (NPDES) or Section 404 (dredge and fill activities).

California's Waste Discharge Requirements, issued by the state's Regional Water Quality Control Boards (RWQCBs), incorporate the authority of the federal NPDES permitting program for discharges of wastes to surface waters. In addition, under Section 303(d) of the each of the RWQCBs has the requirement to establish "a total maximum daily load for those pollutants which the (Environmental Protection Agency (USEPA)) Administrator identifies under Section 304(a) (2) as suitable for such calculation." This section of the CWA was developed to support a water quality-based system of effluent limits for chemical pollutants; the interpretation of what an allowable load of invasive species is has not been defined.

Under Section 305(b) of the CWA, California's nine RWQCBs are required to assess water bodies for attainment of beneficial uses every two years and report to the USEPA. In cases where beneficial uses of water bodies are shown to be impaired, Section 303(d) requires the Regional Boards to list the impaired water bodies and "establish a priority ranking for such waters, taking into account the severity of the pollution and the uses to be made of such waters." Section 502(6) defines "pollutant" as dredged spoil, solid waste, incinerator residue, sewage, garbage, sewage sludge, munitions, chemical wastes, *biological materials*, radioactive materials, heat, wrecked or discarded equipment, rock, sand, cellar dirt, and industrial, municipal and agricultural waste discharged into water. Ballast water is considered to be a pollutant in discharges based on the above definition and definitions in the State Water Code.

Endangered Species Act of 1973

(ESA; 16 U.S.C.A. §§ 1531 to 1544)

<http://www.fws.gov/angered/>

The ESA aims to protect endangered and threatened species. When non-native invasive species threaten endangered species, this act could be used as basis for their eradication or control by the USFWS or by the National Oceanic and Atmospheric Administration – National Marine Fisheries Service (NOAA-Fisheries Service) The potential to harm a federally-listed species and the need to obtain a permit from the USFWS or NOAA-Fisheries Service should be taken into consideration when selecting methods to manage AIS.

Lacey Act (1900; amended 1998)

<http://www.fws.gov/laws/lawsdigest/lacey.html>

As the first federal act that tried to control migrations and importations of nonindigenous species, the Lacey Act prohibits the importation of a list of designated species and other vertebrates, mollusks and crustaceans that are "injurious to human beings, to the interests of agriculture, horticulture, forestry, or to wildlife or the wildlife resources of the United States." Under this law, it is unlawful to import, export, sell, acquire, or purchase fish, wildlife or plants taken, possessed, transported, or sold: 1) in violation of U.S. or Indian law, or 2) in interstate or foreign commerce involving any fish, wildlife, or plants taken possessed or sold in violation of State or foreign law.

The Lacey Act allows for the import of species for scientific, medical, education, exhibition or propagation purposes. The USFWS is the lead agency for enforcing the Lacey Act's prohibition of fish and wildlife imports.

National Environmental Policy Act of 1970
(NEPA; 42 U.S.C.A. §§ 4321 to 4370e)
<http://www.epa.gov/compliance/nepa/index.html>

NEPA requires the consideration of environmental impacts for any federal action, including direct federal activities, permitting and federal funding of activities by another entity. NEPA environmental documents may include a “finding of no significant impact (FONSI),” an “environmental assessment (EA),” or a full “environmental impact statement (EIS).” Potential impacts of invasive species, both direct and indirect, may be among the issues that should be considered under NEPA.

Noxious Weed Act
(1974; 7 U.S.C. § 360)

Section 15 of the Federal Noxious Weed Act requires federal land management agencies to develop and establish a management program for control of undesirable plants that are classified under state or federal law as undesirable, noxious, harmful, injurious or poisonous, on federal lands under the agency’s jurisdiction (7 U.S.C. 2814(a)). The act also requires the federal land management agencies to enter into cooperative agreements to coordinate the management of undesirable plant species on federal lands where similar programs are being implemented on state and private lands in the same area (7 U.S.C. 2814(c)). The Secretaries of Agriculture and the Interior must coordinate their respective control, research and educational efforts relating to noxious weeds (7 U.S.C. 2814(f)). USDA’s Departmental Regulation 9500-10 sets forth departmental policy relating to the management and coordination of noxious weeds activities among the agencies within USDA and other entities.

Plant Protection Act
(2000; 7 U.S.C. 7701)
<http://www.aphis.usda.gov/>

The Plant Protection Act (PPA) authorizes the USDA to prohibit or restrict the importation or interstate movement of any plant, plant product, biological control organism, noxious weed, article or means of conveyance if the Secretary of Agriculture determines that the prohibition or restriction is necessary to prevent the introduction into the United States, or the dissemination within the United States, of a plant pest or noxious weed.

The PPA specifically authorizes USDA to develop integrated management plans for noxious weeds for the geographic region or ecological range where the noxious weed is found in the United States. In addition, the act authorizes the USDA to cooperate with other federal agencies or entities, states or political subdivisions of states, national governments, local governments of other nations, domestic or international organizations or associations, and other persons to carry out the provisions of the act.

FEDERAL AGENCIES

Numerous federal agencies, presented here in alphabetical order, have authority to implement the laws and policies described above. Other federal agencies have mandates impacted by AIS and thus engage in research, monitoring, prevention or control programs. Still others delegate primary responsibility for implementation to state and regional agencies (see next section). The following descriptions attempt to provide a general introduction to the scope of each agency’s work, as well as a brief review of the agency’s recent (as of 2006) major AIS-related activities.

Bureau of Reclamation

<http://www.usbr.gov/>

The Bureau of Reclamation is involved in several important projects related to this issue. The Bureau has partnered with the DFG, USFWS and others to investigate the Chinese mitten crab infestation in the Sacramento-San Joaquin Delta. The agency participates in the Giant *Salvinia* Task Force's efforts to limit the spread of this invader in the Colorado River (see Appendix D), has a detection program for water hyacinth and participates in activities related to the New Zealand mudsnail infestation in Putah Creek. The agency also participated in DFA's *Hydrilla* Eradication Program.

National Oceanic and Atmospheric Administration (NOAA)

<http://www.noaa.gov/>

NOAA is the primary federal agency charged with management of marine resources. NOAA is the co-chair of the ANSTF and has been designated the Department of Commerce lead as co-chair of the National Invasive Species Council. Within NOAA, a number of national, state and regional agencies and programs are actively involved in AIS issues in California. These include: National Estuarine Research Reserve System (NERRS), a network of protected areas established for long-term research, education and stewardship; National Marine Fisheries Service, which works to protect fisheries habitat, commercial fisheries and endangered fish; National Marine Sanctuaries, the nation's system of marine protected areas, and Sea Grant, a nationwide network of 30 university-based programs that work with coastal communities and conduct scientific research and education projects designed to foster science-based decisions for the use and conservation of U.S. aquatic resources.

National Estuarine Research Reserve System (NOAA – NERRS)

<http://nerrs.noaa.gov/>

<http://sfbaynerr.org>

<http://www.elkhornslough.org/>

<http://nerrs.noaa.gov/TijuanaRiver/>

There are three reserves in California that provide a platform to increase communication between scientists, decision-makers, land managers, and the public in order to better deal with AIS issues. The San Francisco Bay reserve protects two large, relatively pristine, tidal wetlands: China Camp State Park in Marin County and Rush Ranch Open Space in Solano County. These sites are part of an AIS early detection and assessment study and detailed vegetation maps are being created to serve as a baseline to evaluate future invasions. China Camp serves as an uninvaded reference site for marshes invaded by *Spartina* hybrids in San Francisco Bay. Rush Ranch is a site of active research on invasive fish and invertebrates. The Elkhorn Slough reserve protects approximately 1,400 acres, including Elkhorn Slough, one of the few coastal wetlands remaining in California. Elkhorn estuarine habitats have over 60 species of non-native invertebrates, over 20 species of non-native plants and a few non-native fish and algae. All of these are currently widespread, so eradication seems impossible. Efforts are focused on early detection and eradication of species identified as "least wanted" invaders such as Chinese mitten crabs and *Caulerpa*. The reserve launched an early detection program for aquatic non-native invaders in 2002. The Tijuana River reserve's 2,500 acres encompass beach, dune, mudflat, salt marsh, riparian, coastal sage and upland habitats surrounded by the growing cities of Tijuana, Imperial Beach and San Diego. Critical invasive species issues include: tamarisk, ice plant and other exotic plants displacing native species in the salt marsh and upland habitats; ongoing surveys to understand the dynamics of AIS; and efforts to understand ecosystem recovery following eradication of invasives.

National Marine Fisheries Service (NOAA – Fisheries Service)

<http://www.nmfs.noaa.gov/>

NOAA-Fisheries Service is in charge sustaining the nation's fisheries, many of which are being directly impacted by AIS, and is involved in many AIS projects in California. It has a key role on the Southern California Caulerpa Action Team. NOAA-Fisheries Service is also involved with a variety of other collaborative research projects including: ballast water exchange, AIS risk evaluation research and hull fouling research funded by the Port of Oakland; analysis of biofouling communities and community effects; and surveys and experimental treatments of several invasive species in San Francisco Bay. NOAA-Fisheries Service also participates on several AIS advisory and coordinating committees including: the Pacific Ballast Water Group, Non-Native Invasive Species Advisory Council and the West Coast Ballast Outreach Project Advisory Team.

National Marine Sanctuaries (NOAA – NMS)

<http://sanctuaries.noaa.gov/>

<http://channelislands.noaa.gov/>

<http://cordellbank.noaa.gov/>

<http://farallones.noaa.gov/>

<http://montereybay.noaa.gov/>

California has four sanctuaries – Channel Islands NMS, Cordell Banks NMS, Gulf of Farallones NMS and Monterey Bay NMS. The latter two sanctuaries are in the process of developing aquatic invasive species management plans and have conducted monitoring programs for AIS.

National Sea Grant (NOAA – Sea Grant)

<http://www.seagrants.noaa.gov/>

<http://www.csgc.ucsd.edu>

<http://ballast-outreach-ucsgep.ucdavis.edu/>

The National Sea Grant Program is a partnership between the nation's universities and NOAA (under the Office of Oceanic and Atmospheric Research) that began in 1966. The California Sea Grant program is the largest of these programs. Sea Grant began the West Coast Ballast Outreach Project in 1999 (co-sponsored by the CALFED Bay-Delta Program) to address concerns that ballast water discharges could be introducing foreign marine species into the state's coastal and estuarine ecosystems. The project educates the maritime industry about the ecological seriousness of aquatic exotic species by publishing the newsletter "Ballast Exchange," maintaining an educational Web site and coordinating workshops. In addition, California Sea Grant provides two major services to the state. First, the research arm of California Sea Grant, operating out of the Scripps Institute for Oceanography in La Jolla, funds critical coastal and marine research through an annual request for proposal and a National Strategic Initiative (NSI) program. Through both of these avenues, the college program funded approximately \$2.6 million in research on invasive species between 1995 and 2003. Second, Sea Grant and the University of California Cooperative Extension jointly fund a network of eleven advisors and specialists who work on applied research and outreach projects throughout the state, including those related to AIS. Sea Grant funding has supported a wide variety of research projects on key invasive species, such as the Chinese mitten crab, European green crab, an exotic Australian isopod, several invasive seaweeds, and *Spartina* hybrids. Sea Grant sponsored research led to the eradication of the South African *sabellid* worm at the site near Cayucos, California, where it had become established.

National Park Service (NPS)

www.nps.gov

NPS strives to preserve the unimpaired natural and cultural resources of the national park system for the enjoyment, education and inspiration of this and future generations. The Park Service cooperates with partners to extend the benefits of natural and cultural resource conservation and outdoor recreation throughout this country. The NPS has several invasive species monitoring, control, research and eradication programs in California. Eradication and control are supported by two programs. The first is the (California) Exotic Plant Management Team (EPMT), which travels around the state to national parks that have requested assistance in removal and control projects. The EPMT has traditionally focused on terrestrial non-natives but could work on aquatic invaders. Through the second program, individual parks can request funds from Washington or the NPS Western Region for control and eradication projects. Natural resource inventories and monitoring activities occur in all of the National Parks in California, and these programs are well positioned to alert state managers to emerging and growing threats from invasive species. Information from these programs could be shared among the California AIS plan partners and benefit the state's early detection efforts. Finally, the NPS actively supports and hosts research projects on impacts of invasive species on ecological communities. National Parks in California, that participate with the EPMT, conduct invasive species inventories, monitoring and research on lands totaling about 2.4 million acres and include hundreds of miles of coastline. Significant education and outreach occurs at all of these sites.

U.S. Army Corps of Engineers (COE)

<http://www.usace.army.mil/>

The COE provides engineering, construction and environmental project services for the military and local governments. Congress authorizes the COE to assist local governments with water resource development needs, which include flood control, navigation, ecosystem restoration and watershed planning. For ecosystem restoration, this includes research on invasive species. Specific programs addressing invasive species issues include the Aquatic Nuisance Species Research Program, the Aquatic Plant Control Research Program and the Water Operations Technical Support Program. COE is also responsible for permitting aquaculture projects, including oyster farms, which often involves AIS considerations.

U.S. Coast Guard (USCG)

<http://www.uscg.mil/hq/g-m/mso/bwm.htm>

The USCG has established a mandatory program aimed at keeping aquatic nuisance species out of U.S. waters using ballast water management methods. USCG activities focus on enforcement and monitoring to ensure compliance with the program, which includes regular on-board inspections. USCG coordinates with California's State Lands Commission, manager of the state's ballast water program. In 2004, USCG issued "Ballast Water Management for the Control of Aquatic Nuisance Species in the Waters of the United States," a guidance document concerning ballast water management.

USCG activities related to AIS are diverse. The agency is working on the development of chemical and engineering methods to verify that a mid-ocean ballast water exchange has occurred. It is also evaluating technologies for the treatment of ballast water. USCG has determined that due to difficulties in establishing the effectiveness of ballast water exchange as it varies across ship types, voyages and from tank to tank, treatment technologies are best evaluated through a ballast water discharge standard (a benchmark for maximum numbers of organisms that may be discharged in ballast water). Such a standard will not only be helpful in evaluating the effectiveness of treatment technologies but also clearly establish when the ballast water no longer contains quantities of organisms that pose a significant risk. A Programmatic Environmental Impact Statement, detailing the evaluation of environmental impacts to the U.S. by several potential ballast water discharge standard alternatives, is currently in development.

USCG has also initiated several projects designed to provide information on the state of development of treatment technologies and the basic characteristics of treatment processes. These efforts have included scientific audits that tested and evaluated three approaches: filtration, ultraviolet light and hydro cyclonic separation. In addition, USCG developed and launched the Shipboard Technology Evaluation Program (STEP) in 2004 to encourage ship owners and operators to participate in evaluating technologies for shipboard application (see also CAISMP Action 7C3). This program allows for the review of experimental plans and treatment technology installations aboard ships. If they perform largely as designed and show promise for reducing the risk of introductions, treatment technology installations will be granted an equivalency with regulations for ballast water management and the Ballast Water Discharge Standard.

U.S. Department of Agriculture (USDA)

<http://www.aphis.usda.gov/>

<http://www.ars.usda.gov/main/main.htm>

<http://www.invasivespeciesinfo.gov>

USDA provides leadership on food, agriculture, natural resources and related issues. USDA conducts a number of programs and activities related to invasive species. USDA's Animal and Plant Health Inspection Service's (APHIS) deals with invaders like the South American wetland rodent, nutria, in the Mississippi Delta region and has also worked on other invasive animal, fish and crab problems around the country. APHIS has done extensive noxious weed work, including exclusion, permitting, eradication of incipient infestations, surveys, data management, public education, and (in cooperation with other agencies) integrated pest management of introduced weeds, including biological control. Aquatic weeds are included in the federal noxious weed list through the APHIS Cooperative Agricultural Pest Survey (CAPS).

The USDA's Agricultural Research Service (ARS) has three Exotic and Invasive Weed Research (EIWR) units in the west: at Davis and Albany, California, and at Reno, Nevada. Scientists at these facilities are responsible for research, the transfer of technology for improvement of management and control, and eradication of invasive aquatic and riparian weeds affecting agriculture and natural resources. These projects address three current ARS program priorities: 1) the reduction of dependence on pesticide use (specifically herbicides); 2) implementation of Executive Order 13112 (see above subsection on this order); and 3) water-quality improvement.

Research is conducted on the biology, reproduction, ecology, management or eradication of several important invasive aquatic weeds. The program provides technology transfer for the eradication and management of several problem species. The EIWR units are also involved in aquatic and riparian weed education for public, state and federal stakeholders.

U.S. Environmental Protection Agency (USEPA)

http://www.epa.gov/owow/invasive_species

USEPA leads the nation's environmental science, research, education and assessment efforts. It develops and enforces regulations, offers financial assistance, performs environmental research, sponsors voluntary partnerships and programs, furthers environmental education and publishes information. USEPA is responsible for enforcing the Clean Water Act (CWA). USEPA released its *EPA Authorities for Natural Resource Managers Developing Aquatic Invasive Species Rapid Response and Management Plans* in December 2005. This document provides an overview of USEPA authorities that apply to state or local AIS rapid response and control actions. The document summarizes relevant sections of the CWA and the Federal Insecticide, Fungicide and Rodenticide Act (FIFRA); summarizes how to apply for CWA Section 404 permits to discharge dredged or fill material; summarizes how to apply for FIFRA Section 18 emergency exemptions and FIFRA Section 24(c) special local need registrations; and describes case studies in which state and local natural resource managers successfully obtained FIFRA emergency

exemptions and special local need registrations for AIS eradication or control actions.

Within USEPA, there are three members of the National Estuary Program in California whose activities encompass AIS management.

National Estuary Program (USEPA – NEP)

<http://www.epa.gov/nep>

San Francisco Estuary Project: <http://www.abag.org/bayarea/sfep/sfep.html>

Morro Bay National Estuary Program: <http://www.mbnep.org/index.php>

Santa Monica Bay Restoration Commission: <http://www.santamonica.org/>

Congress established the National Estuary Program in 1987 to protect and improve the water quality and natural resources of estuaries nationwide. There are three programs in California. The San Francisco Estuary Project (SFEP) was formed in 1987 as a cooperative federal/state/local program to promote effective management of the San Francisco Bay-Delta Estuary, and created a consensus-based management plan for the Estuary including concrete actions related to invasive species. More recently, SFEP identified invasive species as the number-one priority issue in estuary restoration. SFEP holds an ex officio seat on the ANSTF and is a member of the Western Regional Panel.

The Morro Bay National Estuary Program was established in July 1995. The estuary contains the most significant wetland system along California's south-central coast. It supports many species of internationally-protected migratory birds, offers rare wetland habitat to a number of threatened native plant and animal species, and provides a protected harbor for marine fisheries. There are plans to suppress or eliminate at least two aquatic invasive species present in the estuary: giant cane and Sacramento pikeminnow. Efforts to eliminate a pioneer population of giant cane growing along Chorro Creek, a major estuary waterway, and its tributaries, are ongoing; eradication is expected by 2008. Efforts to suppress the pikeminnow to the point where native steelhead populations can begin recovery are expected to begin in 2007.

The Santa Monica Bay Restoration Project was established in 1988 to ensure the long-term health of the 266-square-mile Santa Monica Bay and its 400-square-mile watershed. In 2003, this project became an independent state organization, the Santa Monica Bay Restoration Commission. In terms of invasives, the commission has focused most recently on coastal bluff, wetland and riparian vegetation, funding extensive removal and replanting programs as well as outreach on "California friendly" gardens. The newest threat is the arrival of the New Zealand mudsnail in some Santa Monica mountains streams. The commission has convened experts to strategize how to slow the snail's spread.

U.S. Fish and Wildlife Service (USFWS)

<http://www.fws.gov/>

<http://www.100thmeridian.org>

USFWS has multiple programs that address AIS management. USFWS serves as co-chair of the Federal ANSTF and is the agency that provides federal funding for the implementation of Task Force approved state AIS management plans. USFWS also provides technical assistance to states regarding AIS management. USFWS administers the Lacey Act, which prohibits importation and interstate delivery of listed species. USFWS prevention programs include the 100th Meridian Initiative (see Appendix D), which focuses on preventing the western spread of zebra mussels. In cooperation with the ANSTF, the USFWS has developed planning documents for Chinese mitten crab, European green crab, New Zealand mudsnail and *Caulerpa*. USFWS refuges support invasive species control programs as part of their overall habitat restoration activities.

U.S. Geological Survey (USGS)

<http://www.usgs.gov>

<http://nas.er.usgs.gov/>

USGS acknowledged its role in non-native species management in a White Paper on Invasive Species, which identifies the goal of developing new strategies for the prevention, early detection and prompt eradication of new invaders. The USGS further identifies information management and documentation of invasions as a priority for the agency. In keeping with this objective, the USGS developed and maintains an extensive, spatially referenced database of non-native species, which is accessible online.

APPENDIX C: STATE AUTHORITIES, LEGISLATION & AGENCIES

In California, many state agencies have authority over and regulatory roles for managing natural resources. While diverse agencies have some authority to regulate AIS, there has been no centralized authority or management structure to coordinate AIS activities before this plan. The legal frameworks that apply to control of aquatic invasive species introductions are broad and varied. This section describes the existing authorities that various state agencies and entities have for managing AIS in California, and overlaps somewhat with information presented in Chapters 4 and 5. For help with acronyms, see the Acronym Glossary in the introductory pages of this plan.

CALIFORNIA AUTHORITIES

California Environmental Quality Act (CEQA) (CA Public Resources Code §§ 21000 et seq.)

<http://ceres.ca.gov/ceqa/>

The California Environmental Quality Act (CEQA) requires public disclosure of all significant environmental effects of proposed discretionary projects. If a project would cause significant effects, final documents in the CEQA process show: 1) what mitigation measures will be required to reduce particular effects to a less significant level; and 2) provide justifications for the approval of the project with particular significant effects left unmitigated (i.e. a finding of overriding consideration). CEQA also contains lists of project types exempt from this process. A “significant” impact is a “substantial, or potentially substantial, adverse change in any of the physical conditions within the area affected by the project including land, air, water, minerals, flora, [and] fauna . . .”. The documented adverse impacts associated with invasive species can fit this broad definition.

California Porter-Cologne Water Quality Control Act (CA Water Code §§ 1300 et seq.)

http://www.swrcb.ca.gov/water_laws/docs/portercologne.pdf

Under California’s Porter-Cologne Water Quality Control Act, “any person discharging waste, or proposing to discharge waste, within any region that could affect the quality of the waters of the state” must file a report of the discharge with the appropriate Regional Water Quality Control Board (RWQCB). Pursuant to the act, the RWQCB then prescribes “waste discharge requirements” related to control of the discharge. The act defines “waste” broadly, and the term has been applied to a diverse array of materials. The San Francisco Bay RWQCB, for example, has determined that “ballast water and hull fouling discharges cause pollution as defined under the Porter-Cologne Water Quality Control Act.”

The act, (California Water Code, Division 7), lists a number of types of pollutants that are subject to regulation by the State Water Resources Control Board (SWRCB). Section 13050, for example, specifically includes the regulation of “biological” pollutants by defining them as relevant characteristics of water quality subject to regulation by the Board: AIS are an example of this kind of pollutant if they are discharged to receiving waters. The SWRCB also regards the application of pesticides to control AIS in waters of the state as a discharge of a pollutant requiring an NPDES permit. Several of the Regional Boards have taken legal policy and enforcement actions related to AIS (see also CWA in Appendix B and SWRCB in California Agencies).

Fish and Game Code and Title 14 of the California Code of Regulations

<http://www.fgc.ca.gov/html/regs.html>

<http://www.dfg.ca.gov/ospr/organizational/scientific/exotic/exotic%20report.htm>

The Fish and Game Code consists of the laws passed by the state legislature that pertain to fish and wildlife resources. Under statutes in the Fish and Game Code, the California Fish and Game Commission has the responsibility for the adoption of regulations that provide details on how certain Fish and Game laws are to be implemented. These regulations are published in Title 14 of the California Code of Regulations. A summary is provided below of Fish and Game Code Sections that address invasive species issues or may relate to control actions.

F & G Code §§ 2080 – 2089 DFG regulates the take of species listed under the California Endangered Species Act. In addition to the instructions in the Fish and Game Code, guidelines for this process are located in Title 14, Division 1, Subdivision 3, Chapter 6, Article 1 of the California Code of Regulations. These statutes and regulations should be consulted if AIS control measures have the potential to impact State-listed species.

F & G Code §§ 2118, 2270-2300: DFG is responsible for enforcement of importation, transportation and sheltering of restricted live wild animals; places importation restrictions on aquatic plants and animals; and prohibits nine species of *Caulerpa*.

F & G Code §§6400-6403: It is unlawful to place live fish, fresh or saltwater animals or aquatic plants in any waters of this state without a permit from DFG.

F & G Code §§15000 et seq.: DFG is responsible for regulations pertaining to the aquaculture industry, including disease issues.

Harbors & Navigation Code

The Harbors & Navigation Code, Article 2, Section 64, authorizes the Department of Boating and Waterways to manage aquatic weeds affecting the navigation and use of the state's waterways.

Ballast Management for Control of Nonindigenous Species Act (AB 703) of 1999

This act charged the California State Lands Commission (SLC) with oversight of the state's first program to prevent nonindigenous species (NIS) introductions through the discharge of ballast water from commercial vessels of over 300 gross registered tons (GRT). The 1999 act required that vessels originating from outside the United States Economic Exclusive Zone (U.S. EEZ) carry out mid-ocean exchange or use an approved ballast water treatment method, before discharging in California state waters. The SLC was tasked with: receiving and processing ballast management reports from all such vessels, monitoring ballast management and discharge activities of vessels through submitted reports, inspecting vessels for compliance and assessing vessel reporting rates and compliance. The activities and analyses of the first few years of the program are detailed in the 2003 biennial report of the California Ballast Water Management Program. Upon the sunset of the act, the Marine Invasive Species Act (AB 433) was passed in 2003, revising and widening the scope of the program to more effectively address the invasion threat (see below).

**Marine Invasive Species Act (AB 433) of 2003
(Public Resources Code, Sections 71200-71271;
Title 2, California Code of Regulations, Section 2271)**

The Marine Invasive Species Act, passed in 2003, revises and recasts the state's law pertaining to control of nonindigenous species and ballast water management (AB 703). It imposes additional requirements upon vessel masters, owners, operators and persons in charge of vessels to prevent the introduction of nonindigenous species into waters of the state or waters that may impact the waters of the state. The bill deletes exemptions for specified vessels from compliance with the act and revises the qualifications for the vessels subject to the act.

Ballast water management is required of all vessels greater than 300 gross registered tons (GRT) that intend to discharge ballast water in California waters, though the regulations differ depending on voyage origin. All qualifying vessels coming from ports within the Pacific Coast region must conduct near-coast exchange (in waters at least 50 nautical miles offshore and 200 meters deep) or retain all ballast water and associated sediments. There are exceptions that address safety concerns and for vessels that transit wholly within defined shared waters (San Francisco-/Stockton/Sacramento Delta, and Los Angeles/Long Beach/EI Segundo Complex).

All vessels must complete and submit a ballast water report form upon departure from each port of call in California. They must also comply with the good housekeeping practices, ranging from avoiding discharge near marine sanctuaries to rinsing anchors and removing fouling organisms from the hull. They must maintain a ballast water management plan prepared specifically for the vessel; keep a ballast water log outlining ballast water management activities for each ballast water tank on board the vessel, and make the separate ballast water log available for inspection; conduct training of vessel master, person in charge, and crew regarding the application of ballast water and sediment management and treatment procedures; and pay a fee for each qualifying voyage at their first port of call in California.

In addition to requirements imposed upon vessels operating in state waters, the SLC was charged with the development of several legislative reports offering policymaking guidance on commercial vessel AIS issues including: a Report on Commercial Vessel Fouling in California, Analysis, Evaluation and Recommendations to Reduce Nonindigenous Species Release from the Non-Ballast Water Vector; a Report on Performance Standards for Ballast Water Discharges in California Waters; and a Report on the California Marine Invasive Species Program. These efforts have resulted in the development of regulations to stem transport of AIS in the ballast water of vessels operating with the Pacific Coast Region; and legislation directing SLC to adopt regulations on performance standards for ballast water discharges.

Finally, the legislation also requires DFG to conduct a series of biological surveys to monitor new introductions to coastal and estuarine waters of the state and to assess the effectiveness of the management provision of the Act. AB 703, passed in 1999, required a baseline survey of the state's ports, harbors and bays. AB 433 expanded the baseline to include outer coast sites and required continued monitoring of all sites to determine if the ballast control measures have been successful in reducing the number of new introductions.

**Coastal Ecosystems Protection Act of 2006
(Public Resources Code, Sections 71204.7 – 72423)
(Revenue and Taxation Code, Section 44008)**

The Coastal Ecosystems Protection Act, passed in 2006, adds to the state's law pertaining to the discharge of ballast water (AB 433). It requires the SLC to adopt regulations that require an owner or operators of a vessel carrying, or capable of carrying, ballast water that operates in the waters of the state to implement certain interim and final performance standards for the discharge of ballast water.

California Ocean Protection Council Strategic Plan

http://resources.ca.gov/copc/strategic_plan.html

<http://resources.ca.gov/copc>

The California Ocean Protection Council, formed to coordinate the activities of ocean-related state agencies and improve state efforts to protect ocean resources, among other mandates (see California State Agencies), adopted a five-year strategic plan in 2006. The strategic plan supports the completion and implementation of both the state rapid response plan and this California Aquatic Invasive Species Management Plan, as well as the California Noxious and Invasive Weed Action Plan.

Delta Protection Act

www.delta.ca.gov

California's 1992 Delta Protection Act recognizes the natural resource significance of the 738,000 acre-Sacramento-San Joaquin Delta. The act seeks to preserve and protect Delta resources for the use and enjoyment of current and future generations and recognizes the threat posed by urban encroachment to the Delta's agriculture, wildlife habitat and recreation uses. Pursuant to the Act, a Land Use and Resource Management Plan for the Primary Zone (Management Plan) was completed and adopted by the Commission in 1995. The Management Plan sets out findings, policies and recommendations resulting from background studies in the areas of environment, utilities and infrastructure, land use, agriculture, water, recreation and access, levees and marine patrol boater education/safety programs. As mandated by the act, the policies of the Management Plan are incorporated in the General Plans of local entities having jurisdiction within the Primary Zone. Some of the plan sections relevant to AIS management include: Environment, Finding 8 and Recommendations 3 & 4; Water, Policy 2; and Marine Patrol, Boater Education & Safety, Policy 6 (see also Delta Protection Commission, Appendix D).

CALIFORNIA STATE AGENCIES

San Francisco Bay Conservation and Development Commission (BCDC)

<http://www.bcdc.ca.gov/>

The Bay Conservation and Development Commission is dedicated to the protection and enhancement of San Francisco Bay and to the encouragement of the Bay's responsible use. Any person or government agency wishing to place fill, extract materials or make any substantial change in use of any water, land or structure within the area of the Commission's jurisdiction requires a Commission permit or federal consistency determination. The Commission's jurisdiction includes San Francisco Bay, including tidal flats, subtidal areas and marshlands lying between mean high tide and five feet above mean sea level and a 100 foot shoreline band measured inland from the Bay shoreline, as defined by Section 66610 of the McAtteer-Petris Act. The Commission recognizes the threat of non-native invasive species to the Bay's ecosystem and the *San Francisco Bay Plan* contains policies regarding the monitoring, control and eradication of aquatic invasive species in the Bay.

California Department of Boating and Waterways (DBW)

<http://www.dbw.ca.gov/>

DBW works to help develop convenient public access to California waterways, promote on-the-water safety and keep waterways free of navigational problems. General activities include boating law enforcement, boater education, improvements to boating facilities and vessel sewage management. In addition, DBW manages the state's largest and oldest aquatic weed control program, working with other public agencies to control water hyacinth, and more recently Brazilian elodea, in the Sacramento-San Joaquin Delta, its tributaries and the Suisun Marsh. DBW also leads the California Clean Boating Network, a collaboration of government, business, boating and academic organizations working to increase and improve clean boating education efforts, including invasive species education, across the state.

California Coastal Commission (CCC)

<http://www.coastal.ca.gov/>

The CCC is mandated to protect and enhance public access, recreation, wetlands, visual resources, agriculture, commercial activity, industrial activity and environmentally sensitive habitats within the coastal zone through coastal development permits, local coastal programs and federal consistency review. The CCC has responsibility to protect both the biology of aquatic ecosystems and the special uses associated with the marine environment, such as commercial fishing and recreation. The CCC regulates development activities in state waters under its coastal development permit authority and is responsible for working with local governments within the coastal zone. The CCC is also the designated coastal management agency administering the federal Coastal Zone Management Act (CZMA) over Pacific waters offshore of California (outside of San Francisco Bay). As such, the Coastal Commission exercises federal consistency review authority over all federal activities and federally licensed, permitted or funded activities affecting the coastal zone, regardless of whether the activity occurs within, landward, or seaward of the coastal zone boundary. Federal agency activities, including permits and plans, are subject to the consistency determination process, and must be "consistent to the maximum extent practicable" with the state's coastal management program, in this case, the Chapter 3 policies of the California Coastal Act (15 CFR § 930.32).

California Department of Fish and Game (DFG)

<http://www.dfg.ca.gov/>

<http://www.dfg.ca.gov/ospr/>

DFG has jurisdiction over the conservation, protection and management of fish, wildlife, plants and habitat necessary for biologically sustainable populations of those species. DFG conducts a number of programs related to aquatic invasive species, including serving as the lead agency in developing this statewide AIS management plan, as well as a rapid response plan for invasions (see Appendix A). DFG is responsible for enforcement of regulations concerning the aquaculture industry; the importation and transport of live wild animals, aquatic plants and fish into the state; and the placement of any such animals in state waters. The agency is also responsible for conducting biological surveys to assess the amount and types of AIS present in state waters, and the degree of success of ballast water management activities. Starting in 1999 with ballast management legislation, these surveys have been undertaken by DFG's Office of Spill Prevention and Response (DFG/OSPR). DFG/OSPR also manages the California Aquatic Non-Native Organism Database (CANOD) and is working to establish consistency among the various major databases being used to analyze similar types of AIS-related information. Lastly, DFG has been an active manager or partner in numerous AIS eradication and control programs, especially for those AIS that threaten at-risk species or the conservation and restoration of aquatic or riparian ecosystems.

California Department of Food and Agriculture (DFA)

<http://www.cdffa.ca.gov/>

DFA is the lead agency for regulatory activities associated with aquatic weeds. This regulatory authority includes quarantine, exterior pest exclusion (border protection stations and inspections), interior pest exclusion (pet/aquaria stores, aquatic plant dealers and nurseries) and detection and control/eradication programs. In addition, the DFA Plant Pest Diagnostic Center identifies plant species and assigns plant pest ratings. DFA maintains a rated list of noxious weed species. "A"-rated pests require eradication, containment, rejection or other holding actions at the state-county level. Quarantine interceptions are to be rejected or treated at any point in the state. For "B"-rated pests, eradication, containment, control or other holding actions are taken at the discretion of the agricultural commissioner. State-endorsed holding actions and eradication of "C"-rated pests occur only when these pests are found in a nursery. Action is taken to retard spread outside of nurseries at the discretion of the commissioner. Rejection occurs only when found in a crop seed for planting or at the discretion of the commissioner. "Q" ratings are temporary "A" ratings pending determination of a permanent rating. DFA is also responsible for the *Hydrilla* eradication program (see Chapter 2).

County Agricultural Commissioners (CACs)

<http://www.cdffa.ca.gov/exec/cl/cacasa.htm>

CACs have long been at the forefront in the battle against invasive species throughout the state. They work collaboratively with DFA and other agencies to exclude, detect and eradicate or manage a wide range of pest species. CACs perform numerous inspections of incoming plant materials, checking for compliance with quarantine requirements and for noxious weeds and other pests. Nurseries and pet stores are also inspected. The CACs have worked with DFA to obtain additional resources to fund more effective programs. Once plant materials enter the state, it is generally the CACs who perform inspections and carry out most of the weed eradication and management activities. While the CACs are not a "state" agency, they form a statewide system, represented at the state level by California Agricultural Commissioners and Sealers Association (CACASA) and have specific authorities granted by state law to carry out pest prevention programs.

California Department of Parks and Recreation (PARKS)

<http://www.parks.ca.gov/>

PARKS manages more than 270 park units and approximately 1.4 million acres, of which more than 280 miles is coastline and 625 miles of lake and river frontage. Management objectives of individual properties within the system depend on a unit's classification and range from a preservation mandate to a recreation emphasis. Units of the state park system can be established in either the terrestrial or underwater environment. Management to restore natural processes is basic to many types of state park units. This management includes removal of exotic species and is expected to extend below the waterline in units that are primarily terrestrial.

California Department of Pesticide Regulation (DPR)

<http://www.cdpr.ca.gov/>

DPR is vested with primary responsibility to enforce federal and state pesticide laws and regulations pertaining to the proper and safe use of pesticides in California. The Department regulates pesticides under a comprehensive program that includes enforcement of pesticide use in agricultural and urban environments, prevention of environmental contamination, environmental monitoring for emergency eradication projects and other related functions. DPR conducts monitoring of emergency eradication projects to ascertain that the public and the environment are being protected and the correct amounts of pesticides are being applied. DPR conducts sampling in consultation with the County Agricultural Commissioners, Department of Fish and Game, the RWQCBs and other stakeholders. DPR works cooperatively with other government agencies sharing information and monitoring results.

California Department of Water Resources (DWR)

<http://www.water.ca.gov/>

DWR addresses invasive species issues that impact water supply, water delivery and flood control. In general, DWR administers programs involving flood control for the Central Valley, dam safety for more than 1,200 dams statewide, design and construction of water facilities, water quality improvement and water supply data collection and studies. DWR also operates and maintains the State Water Project (SWP).

Recent activities related to invasive species are diverse. DWR conducts monthly monitoring of benthic (bottom-dwelling) invertebrates, zooplankton and phytoplankton throughout the upper San Francisco Estuary and reports trends in invertebrate abundance and community composition, including newly introduced species, to the State Water Resources Control Board. DWR is documenting the distribution of the invasive algal species *Microcystis spp.* in the upper San Francisco Estuary, investigating which strains (toxic versus non-toxic) are present and examining effects on the aquatic food web. DWR is also investigating the impacts of the Chinese mitten crab on the benthic invertebrate community in the Sacramento-San Joaquin Delta and co-authored a white paper on its life history.

On the prevention front, DWR implemented the California Zebra Mussel Watch Program until June 2005 (which included risk assessment, early detection, public outreach, the development of a rapid response plan for the Central Valley watershed and a centralized reporting system for mussel sightings). The future of this program depends on funding. At Lake Davis, DWR has been coordinating with DFG on northern pike control and downstream protection (including the installation of a structure to prevent pike escape over the dam). DWR contributes to programs aimed at controlling invasive weeds along eroding Sacramento River banks, within flood control and water conveyance structures and along urban streams. The agency coordinates its activities with other state and federal agencies as a member of the CALFED Non-native Invasive Species Advisory Council (NISAC).

California Ocean Protection Council (OPC)

<http://www.coastalconservancy.ca.gov/>

The OPC, created in 2004, is a state cabinet level council consisting of the Secretaries for Resources and the California Environmental Protection Agency, the chair of the State Lands Commission and two members of the Legislature. The OPC is a policy making body and also prioritizes the expenditure of various funds appropriated to other State departments for ocean protection purposes. The OPC has authorized funding for the completion of this AIS plan and is considering inclusion of implementation of this plan in its strategic plan as a major objective over the next five years. OPC's policies are administered by the Coastal Conservancy with direction from an Executive Policy Officer housed at the Resources Agency.

California State Lands Commission (SLC)

<http://www.slc.ca.gov>

SLC manages the mandatory, statewide, multi-agency Marine Invasive Species Program. This program works to implement regulations governing ballast water management for vessels operating on the West Coast of North America. Commission inspectors board approximately 25% of all vessels that arrive in California to verify compliance with regulations and to disseminate outreach materials to vessels and crews new to California. In addition to its regulatory activities, the Commission facilitates scientific research and technology development to enhance management efforts of the program and to inform policymakers. Limited funding is provided for research that targets priority information gaps and to technologies that show exceptional promise for the treatment of ballast water. In recent years, the SLC has also prepared a number of reports for the state legislature documenting commercial vessel fouling in California, proposing performance standards for ballast water discharges, and summarizing vessel ballast water activities and compliance in California (see also Ballast Water Management, California Authorities, and Chapter 5). In addition to the mandated Marine Invasive Species Program, the SLC has been coordinating interagency efforts to manage invasive aquatic plants such as Eurasian watermilfoil in Lake Tahoe (see Case Study, Chapter 8).

State Coastal Conservancy (SCC)

<http://www.coastalconservancy.ca.gov/>

SCC has been involved for over twenty years in the control and eradication of aquatic invasives, pursuant to Division 21 of the Public Resources Code. SCC developed, funded and operates the Invasive *Spartina* Project in San Francisco Bay that shows great promise in eradicating nonindigenous species of *Spartina* and their associated hybrids. SCC is also involved in efforts to control *Arundo* in many coastal watersheds. SCC directly develops projects and provides grant funds related to resources enhancement and restoration, including control and elimination of invasives. SCC is also a partner in developing this management plan.

The San Francisco Estuary Invasive *Spartina* Project (ISP)

<http://www.spartina.org/>

SCC established the ISP in 2000. Its overall goal is to develop and implement a regionally coordinated project to eradicate the four introduced and highly invasive *Spartina* species in the San Francisco Estuary. The ISP is comprised of a number of components, including outreach, research, permitting, mapping, monitoring and allocation of funds for efforts to eliminate populations of nonindigenous *Spartina*. In 2005 the Conservancy and ISP began full-scale implementation of the regionally coordinated *Spartina* Control Program (SCP), employing an aggressive treatment strategy to target nearly all infested sites in the San Francisco Estuary. Initial results show on average about 85% efficacy at treated sites. SCC will continue to coordinate the regional control effort through the ISP, and to allocate funds to land owners and managers around the San Francisco Bay for aggressive treatment activities consistent with the SCP. If funding

continues, it's expected that invasive *Spartina* will be effectively eradicated from the San Francisco Estuary between 2009 and 2011 (see also Case Study, Chapter 8).

State Water Resources Control Board (SWRCB)

<http://www.swrcb.ca.gov/>

The SWRCB's mission is to preserve, enhance and restore the quality of California's water resources, and ensure their proper allocation and efficient use for the benefit of present and future generations. The Board has joint authority over water allocation and water quality protection. Under the State Board are nine Regional Water Quality Control Boards (RWQCBs). The SWRCB and regional boards have been working in support of, and in an advisory capacity to, other state agencies on various AIS activities, such as hull fouling and ballast water management. Invasives come under water board purview as part of the state's efforts to implement and enforce the Clean Water Act (CWA, see also Appendix B). A 2005 federal court ruling defined non-indigenous species as "pollutants" present in discharges from vessels and found that such discharges are not exempt from permitting requirements (NPDES, see also CWA, Appendix B).

In terms of AIS management activities, some of the regional boards have also sought to place specific water bodies within their regions on the CWA's 303(d) list, as impaired by exotics. S.F. Bay was listed in 1998. In 2006, the State Board placed the Delta, the Cosumnes River and a portion of the San Joaquin River on the 303 (d) list. Once on the 303(d) list, the regional boards are required to develop discharger/source based programs for managing pollutants, including the determination of "total maximum daily loads" (TMDLs)), which in the case of exotics have proved somewhat difficult to develop. Trying to allocate loads or goals for zero loads, among dischargers, water users and municipalities is challenging when most of the water bodies in question are already heavily invaded. Despite the implementation challenges, the S.F. Bay Water board's work on the state's first exotics TMDL did, however, widely publicize the problem and led to other successful AIS management and legislative programs.

Other regional boards have become involved in AIS-related water quality issues through watershed management projects, non-point source pollution management programs and wetland mitigation and restoration programs (raising issues about the use of non-native aquatic plant species for these programs, and the control of invasives, for example). The State Board has also participated in AIS management activities concerning the use of aquatic pesticides.

University of California (UC)

www.universityofcalifornia.edu

www.ipm.ucdavis.edu/

UC conducts extensive research on invasive species issues and has a substantial pool of scientists devoted to biological invasions and management. UC faculty serve on NGO, and state and federal government panels and committees charged with invasive species management. They also provide expertise and management for a variety of cooperative government units such as UC's Division of Agricultural and Natural Resources' (ANR) Integrated Pest Management Program and the Center for Invasive Species Research (UC Riverside). This center has managed the Exotic Species Research Program for USDA for almost five years. UC ANR also has Marine Advisors in most coastal counties in the state as part of the Sea Grant extension program. This provides a direct academic presence for extension outreach and applied research collaboration with agencies and campus faculty (see also National Sea Grant, Appendix B). UC also has formal graduate training programs on invasive species, such as the Integrative Graduate Education and Research Traineeship, based at UC Davis, in which the students intern with DFG, USFWS and other government agencies.

APPENDIX D: OTHER AIS INTERESTS

COORDINATING COMMITTEES, EDUCATIONAL INITIATIVES & SPECIAL INTEREST GROUPS

AIS spread across so many jurisdictions and impact so many different types of human activities and environmental priorities that diverse efforts have been made to promote coordination among AIS-involved agencies, organizations and stakeholders. Some of these, such as CALFED or the Western Regional Panel serve important functions in implementing federal and state mandates for coordination. Others provide ongoing forums for information sharing and priority setting among different agencies, organizations and interest groups, or among those attempting to restore or preserve specific waterways.

COORDINATING COMMITTEES & PARTNERSHIPS

Aquatic Nuisance Species Task Force

www.anstaskforce.gov

Federal legislation established the national Aquatic Nuisance Species Task Force (ANSTF), co-chaired by the USFWS and NOAA. ANSTF is charged with coordinating governmental efforts related to ANS prevention and control. ANSTF consists of 10 federal agency representatives and 12 ex officio members representing nonfederal governmental agencies.

Adopt-A-Riverway Program

This program is a government-volunteer partnership established in 2003. Participation in the program includes management of noxious and invasive weeds. Authorized program activities include planting and establishing native seedling trees, shrubs, native grasses, wildflowers, and removing litter and weeds, consistent with an integrated weed management plan. AB 66, a state bill, established an Adopt-A-Riverway Fund for proceeds donated, appropriated, transferred or otherwise received for purposes pertaining to the Adopt-A-Riverway Program.

Association of Fish and Wildlife Agencies (AFWA)

<http://www.fishwildlife.org/>

AFWA represents the government agencies responsible for North America's fish and wildlife resources. It promotes sound management and conservation and speaks with a unified voice on important fish and wildlife issues. AFWA was awarded a recent grant to create communications strategies on issues related to unwanted invasive aquatic species. This project will help states develop comprehensive programs to address aquatic nuisance species issues within their states and will collectively help the Regional Associations and the AFWA nationally develop a stronger voice and greater capabilities when addressing regional and national aquatic nuisance species efforts.

CALFED Bay-Delta Program (CALFED)

<http://calwater.ca.gov/>

CALFED is a cooperative effort of more than 20 state and federal agencies working with local communities to improve the water quality and reliability of California's water supplies and restore the San Francisco Bay-Delta ecosystem. One goal of CALFED's Ecosystem Restoration Program (ERP) has been to "prevent establishment of and reduce impacts from non-native species." The goal includes 10 specific objectives, such as eliminating further introductions of new species in ballast water of ships and preventing the invasion of the zebra mussel into California. CALFED has also developed a strategic plan for managing non-native invasive species in the San Francisco Bay-Delta Estuary and the Sacramento and San Joaquin Rivers and associated watersheds. To date, CALFED has funded 31 projects that address preventing the establishment of, or reducing the impacts from, non-native invasive species in California.

CALFED also created a Non-native Invasive Species Advisory Council (NISAC), a council of agency and technical stakeholders to advise the program on non-native invasive species.

California Horticultural Invasives Prevention (Cal-HIP)

www.suscon.org/invasives

This partnership develops strategies to reduce introductions of invasive plants through horticulture. Partners include environmental NGOs, agency representatives, and nursery and landscaping trade organizations. Sustainable Conservation, a nonprofit organization, facilitates the partnership.

California Interagency Noxious & Invasive Plant Committee (CINIPC)

http://www.cdfa.ca.gov/phpps/ipc/CINWCC/cinwcc_hp.htm

This committee, formerly known as California Interagency Noxious Weed Coordinating Committee (CINWCC), was formed in 1995, with a memorandum of understanding among 14 federal and state agencies. The committee changed its name again in 2006. Its mission is to facilitate, promote and coordinate the establishment of an integrated pest management partnership between public and private land managers toward the eradication and control of noxious weeds on federal and state lands and on private lands adjacent to public lands.

California Invasive Plant Council (Cal-IPC).

www.cal-ipc.org

This Council is a nonprofit organization that works to protect California wild lands from invasive plants through research, restoration and education. Cal-IPC proposes and facilitates solutions to problems caused by invasive plants. Membership includes public and private land managers, ecological consultants, researchers, planners, volunteer stewards and concerned citizens. Cal-IPC is recognized as an authoritative source of new information on all aspects of wild land weed management.

California Invasive Weed Awareness Coalition (CALIWAC)

www.cal-ipc.org/policy/state/caliwac.php

This coalition, made up of primarily industry stakeholders, was formed in 2001 to increase awareness of the invasive weed issue in California. The coalition's goals are to support the development of a statewide management plan for invasive weeds; provide a public forum to increase awareness of the detrimental environmental and economic effects of invasive weeds and contribute to solutions for invasive weed issues; promote increased funding for management of invasive weeds; and influence state and national policy on invasive weeds

California Weed Science Society (CWSS)

<http://www.cwss.org/>

This Society was founded in 1948 to promote environmentally sound proactive research and develop educational programs in weed science; support undergraduate/graduate students seeking a career in weed science; and encourage and support educational activities to promote integrated weed management systems.

County Weed Management Areas (WMA)

A Weed Management Area (WMA) is a local organization that brings together landowners and managers (private, city, county, state, and federal) in a county, multi-county or other geographical area for the purpose of coordinating and combining action and expertise in combating common invasive weed species. The WMA Support Program in DFA provides coordination and training opportunities and allocates state funding earmarked for WMAs.

Delta Protection Commission (DPC)

www.delta.ca.gov

California's 1992 Delta Protection Act created a Delta Protection Commission in recognition of the natural resource significance of the 738,000 acre-Sacramento-San Joaquin Delta. The Act seeks to preserve and protect Delta resources for the use and enjoyment of current and future generations and recognizes the threat posed by urban encroachment to the Delta's agriculture, wildlife habitat and recreation uses (see also Appendix C, State Authorities). The 19-member Delta Protection Commission provides for stakeholder representation in the areas of agriculture, habitat, and recreation. A land use and resource management plan for the primary zone of the Delta, completed in 1995 and updated in 2002, acknowledges the impacts of exotic species on Delta resources and makes recommendations for preventing impacts on native fish, and on aquatic, channel island and seasonal wetland habitats (including mosquito abatement projects).

Pacific Ballast Water Group (PBWG)

<http://www.psmfc.org/ballast/>

This group was formed by representatives from the shipping industry, state and federal agencies, environmental organizations, and others who recognized the need for a cooperative and coordinated regional approach to ballast water management to prevent the introduction of invasive species on the West Coast. The PBWG meets regularly and is currently addressing the development of ballast water discharge standards and inter-jurisdictional issues related to ballast water management on the West Coast.

Pacific States Marine Fisheries Commission (PSMFC)

<http://www.psmfc.org/>

PSMFC is one of three interstate commissions dedicated to resolving fishery issues. Representation includes the states of California, Oregon, Washington, Idaho and Alaska. The PSMFC does not have regulatory or management authority; rather, it serves as a forum for discussion, works towards coast wide consensus on state and federal authorities and addresses issues that fall outside state or regional management jurisdiction. Over the past four years, the Pacific States Marine Fisheries Commission's AIS program has concentrated on four species of aquatic invaders: Chinese mitten crab, European green crab, zebra/quagga mussel and Atlantic salmon. Program activities include research and monitoring, educational outreach, interjurisdictional planning and coordination, and funding and contracting services for numerous partners.

Western Governors' Association

<http://www.westgov.org/>

The Western Governors' Association is developing a new program to address undesirable nonindigenous aquatic and terrestrial species in the west. In 1998, the Western Governors passed a resolution on Undesirable Aquatic and Terrestrial Species to develop and coordinate western strategies and to support management actions to control and prevent the spread and introduction of undesirable species; support the use of integrated pest management concepts; encourage broad-based partnerships; and urge adequate support for the U.S. Department of Agriculture's Animal and Plant Health Inspection Service. The Association has formed a working group of state and federal agencies, industry, non-governmental organizations and academia to develop western strategies to limit the spread of these species.

Western Regional Panel (WRP)

<http://www.fws.gov/answest/>

This panel on Aquatic Nuisance Species was formed as a committee of the ANSTF after the passage of NISA to help limit the introduction, spread and impacts of aquatic nuisance species into western North America. This panel includes representatives from federal, state, and local agencies, Native American tribes, and private environmental and commercial interests, as well as a representative from Canada.

The general goals of the WRP are to prevent nuisance species introductions, coordinate activities of the western states among federal, local and tribal agencies and organizations, and minimize impacts of already established nuisance species. The purposes of the WRP, as described in NISA, are to: identify western region priorities for responding to aquatic nuisance species; make recommendations to the ANSTF regarding an education, monitoring (including inspection), prevention, and control program to prevent the spread of the zebra mussel west of the 100th meridian; coordinate other aquatic nuisance species activities in the west not conducted pursuant to the act; develop an emergency response strategy for federal, state, and local entities for stemming new invasions of aquatic nuisance species in the region; provide technical assistance to public and private stakeholders for preventing and controlling aquatic nuisance species infestations; and submit an annual report to the ANSTF describing activities related to ANS prevention, research and control.

MAJOR NATIONAL EDUCATION CAMPAIGNS

100th Meridian Initiative, USFWS

<http://www.100thmeridian.org>

The primary goal of the 100th Meridian Initiative is to prevent the further spread of zebra mussels. At the time it was formed, the western limit of the zebra/quagga mussel roughly coincided with the 100th meridian. It is the first large-scale, cross-jurisdictional effort to combat the spread of an aquatic invasive species. Participating entities include federal, state, local and tribal governments, potentially affected industries such as commercial boat haulers and other stakeholders. The initiative has produced an extensive public information and education campaign aimed at marina users, anglers and recreational boaters. It sponsors the production of posters, informational flyers and signs educating boaters about the risks of zebra mussels and other AIS. Its members conduct voluntary boat inspections and boater surveys to identify boats at highest risk for harboring AIS. Collected boater travel patterns are being used to model potential pathways for the mussel's spread. The initiative has supported the establishment of mussel monitoring stations across the west, as well as the development of regional rapid response plans should the mussel establish new populations. Recent programs include the Lewis and Clark Initiative, a program aimed at increasing outreach efforts to recreational boaters retracing the path of the historic expedition during its bicentennial. Among other accomplishments, the effort resulted in the establishment of more AIS monitoring stations and a mussel monitoring database for the Columbia River Basin region.

Habitattitude

www.habitattitude.net

Habitattitude is an ANSTF collaboration of the Pet Industry Joint Advisory Council (PIJAC), the U.S. Fish & Wildlife Service, the NOAA National Sea Grant College Program, and the nursery and landscape industry. It was established in 2004 to educate aquarium hobbyists, backyard pond owners, water garden enthusiasts, and others on how to prevent the spread of potential aquatic nuisance species. Its web site includes information on how non-native fish and plants can harm ecosystems, suggests environmentally sound alternatives to releasing unwanted aquatic plants and animals in the wild and offers tips on how to prevent accidental releases. The site offers promotional materials, signage and decals for participating retailers and manufacturers. The initiative offers a means for industry and the USFWS to work together to promote their shared interests in preventing AIS impacts.

Stop Aquatic Hitchhikers

www.protectyourwaters.com

The Stop Aquatic Hitchhikers web site is part of the ANSTF public awareness campaign. It is sponsored by the USFWS and the USCG. It functions as a reputable, central source of information about aquatic nuisance species affecting the United States. Resources include photos and descriptions of common nuisance species, how they impact ecosystems, boaters and anglers, and tips for preventing their spread. A news page features stories from major news outlets as well as government news releases related to AIS. Video and audio clips geared toward traveler information centers are available for download as are outreach materials such as posters, flyers, stickers for tackle boxes, banners and signs. Clubs, state and government agencies, and private entities are encouraged to join the campaign and pledge to prevent the spread of AIS. In California, partners include the DFG, California Trout, the City of Davis, Heal the Bay (Santa Monica), and the Santa Ana Zoo, among others.

SPECIES- & PLACE-SPECIFIC COALITIONS, INITIATIVES & NONPROFITS

100th Meridian Initiative, USFWS

(see Major National Education Campaigns)

California Sea Grant

(see Appendix B, NOAA – Sea Grant)

Channel Islands National Marine Sanctuary

(see Appendix B, NOAA – NMS)

Cordell Banks National Marine Sanctuary

(see Appendix B, NOAA – NMS)

Elkhorn Slough National Estuarine Research Reserve

(see Appendix B, NOAA – NERR)

Gulf of the Farallones National Marine Sanctuary

(see Appendix B, NOAA – NMS)

Invasive Spartina Project

(see Appendix C, State Coastal Conservancy)

Lower Colorado River Giant Salvinia Task Force

<http://lcrsalvinia.org/salviniahome.asp>

On August 4, 1999, the USFWS found giant salvinia in the Imperial National Wildlife Refuge on the Colorado River. Plants were also seen floating down the Colorado River, on the Cibola National Wildlife Refuge, and in Pretty Water and Three Finger lakes. Subsequent investigation determined that the source of the infestation was the West Side/Outfall Drain of the Palo Verde Irrigation District near Blythe, California. To ensure a coordinated response to the infestation, a task force was formed. Teams focused on accomplishing steps to control and/or eradicate giant salvinia in the lower Colorado River. Teams address issues relating to research, monitoring, rapid response, field implementation, regulation and compliance, outreach, and financial and international issues.

Monterey Bay National Marine Sanctuary

(see Appendix B, NOAA – NMS)

Morro Bay National Estuary Program (USEPA National Estuary Program)

(see Appendix B, USEPA – NEP)

San Francisco Bay National Estuarine Research Reserve

(see Appendix B, NOAA – NERR)

San Francisco Estuary Institute

www.sfei.org/bioinvasions

SFEI was founded as a non-profit organization in 1986 to foster the scientific understanding needed to protect and enhance the San Francisco Estuary. It is governed by a board composed of Bay Area scientists, environmentalists, regulators, local governments and industries. SFEI's Biological Invasions program conducts scientific and policy research and provides information and analyses on the introduction of exotic organisms into marine and freshwater ecosystems. In the last decade, the program has been actively working to improve understanding and management of invasive species, to document the status of invasive species in San Francisco Bay and the increasing rate of invasions. The program is also involved in

helping develop regulatory standards for ballast water discharges. Most recently, SFEI is chairing the scientific advisory panel that is providing guidance from the research community to the government agencies responding to the recent discovery of quagga mussel in California and performing some of the research identified by the quagga mussel incident command.

San Francisco Estuary Project (USEPA National Estuary Program)

(see Appendix B, USEPA – NEP)

Santa Monica Bay Restoration Commission (USEPA National Estuary Program)

(see Appendix B, USEPA – NEP)

Southern California Caulerpa Action Team (SCCAT)

<http://www.sccat.net/>

SCCAT was established to respond quickly and effectively to the discovery of *Caulerpa* in Southern California. The group consists of representatives from local, state, and federal governmental entities and from private organizations. SCCAT's goal is to completely eradicate all infestations in Agua Hedionda Lagoon and Huntington Harbour and to prevent new infestations (see also Chapter 8, Case Study)

Tahoe Basin Weed Coordinating Group

(775) 784-4848

This group is coordinated through the University of Nevada Cooperative Extension to address the increasing aquatic weed problem in the two-state Lake Tahoe Basin. This group and local agencies have undertaken mechanical removal of Eurasian watermilfoil and efforts are now being expanded, incorporating a variety of removal methods (see also Case Study, Chapter 8).

Team Arundo

<http://www.sawpa.org/arundo/>

Team Arundo was formed in Orange County, California, in 1991 to control *Arundo* along the Santa Ana River, and has since become a statewide program. Chapters exist in the Bay Area, San Luis Obispo and surrounding counties, Greater Los Angeles County, and San Diego County.

Team Arundo Del Norte

<http://ceres.ca.gov/tadn/>

Team Arundo Del Norte is a forum of local, state and federal organizations dedicated to the control of *Arundo* in rivers, creeks and wetlands in Central and Northern California. The organization formed in the summer of 1996 and meets several times per year in the Sacramento area to explore opportunities for information exchange and partnerships in support of the ongoing work of eradicating *Arundo*.

Tijuana River National Estuarine Research Reserve

(see Appendix B, NOAA – NERRS)

APPENDIX E: AIS PLAN DEVELOPMENT & PROCESS

An initial draft of this plan was developed for DFG several years ago with stakeholder input (see below). At that time the plan was not completed due to funding and staffing issues. In 2006, additional funding was awarded to SFEP from the OPC, through the SCC, to finish and begin implementation of the plan.

2006 Draft & Final Plan Process

The 2006 draft of the plan incorporated much of the text, research and public comments provided by the original 2004 draft (see below).

In early 2006, agency staff reviewed the 2004 version and suggested updates. The resulting draft was circulated two times for review and comment by AIS program managers within lead state and federal agencies. Two internal meetings – one in June, and one in July – were held to discuss the draft and documented in meeting notes. Revisions were made accordingly.

The resulting draft plan was posted for public review on August 22nd, 2006. Three public meetings were held in August and September 2006 in Oakland, Sacramento and Long Beach to review the draft plan. Public comments were reviewed and incorporated to the extent possible.

Attendees at one or more of 2006 internal interagency meetings included:

Susan Ellis, DFG
Abe Doherty, SCC
Julie Horenstein, DFG
Dan Wilson, DFG
Paul Ryan, DBW
Geoff Newman, DBW
Terri Ely, DBW
Marian Ashe, DFG/OSPR
Jeffrey Herod, USFWS
Marcia Carlock, DBW
Suzanne Gilmore, SLC
Tanya Veldhuizen, DWR
Lynn Takata, SLC
Ben Becker, NPS
Karen McDowell, SFEP
Maurya Falkner, SLC
Pat Akers, DFA

2006 Public Meetings Summary

Background

A Draft AIS Plan was publicly released in late August 2006 and three public meetings were held in August and September to solicit input. The following pages summarize the presentation used at all three meetings and present comments and questions raised by meeting attendees. In addition, the results from a “prioritization” exercise conducted at each meeting are presented.

Meeting Overview

The meetings were called to order by Austin McInerny, facilitator, from the Center for Collaborative Policy, California State University, Sacramento. After McInerny provided an overview of the meeting agenda, participants and staff involved in preparing the Draft AIS Plan, introduced themselves. Project staff participating in the meetings included:

- Susan Ellis, Invasive Species Coordinator, DFG
- Julie Horenstein, DFG
- Karen McDowell, Project Coordinator, SFEP
- Abe Doherty, Project Manager, SCC
- Paula Trigueros, SFEP (note taker)
- Debbi Egter Van Wissekerke, SFEP (logistics manager)

Karen McDowell provided a brief background and overview of the plan’s development process and explained the need to complete the plan to qualify for federal funding. She further clarified that the plan is to provide a management framework for agency coordination and that the anticipated adoption timeline is very aggressive. She reviewed the required components of the plan and explained the proposed management framework and the Technical Advisory Panels. She highlighted the objectives, strategies and action items for implementation and noted the priority section would be completed following the public review process. She explained the appendices including the Rapid Response Plan. Next steps included posting updates on the website and including the public comments as an Appendix also to be posted on the web. The complete presentation is available online at:

http://sfep.abag.ca.gov/projects/invasive_species.html

Following the presentation, a short question and answer period was held to address questions on how the plan was developed. Then, meeting attendees provided feedback, comments, and questions regarding the Draft AIS Plan. Lastly, meeting attendees were asked to review the proposed Action Items proposed in the plan and identify what they believed were both “high” and “low” priority action items.

Comment forms were provided and copies of the Draft AIS Plan were available for review.

Public Meeting #1 (Sacramento) Summary

The meeting was held August 28 in the auditorium of the California Department of Food & Agriculture and had nearly 30 attendees. The following comments and questions were raised:

- Woody Schon, Sacramento/Yolo Mosquito & Vector Control District: Expressed concern with Action 2E4 regarding use of mosquito fish for mosquito control. His district uses fish to control mosquitoes in degraded habitats such as rice or agricultural fields that are not flowing into streams, rivers or vernal pools and does not want to see these fish excluded as a tool for mosquito control.
- Raynor Tsuneyoshi, Director, DBW: Would like to see Collaborative Center for AIS at a university. Concern with hull cleaning for small boats – it is 9 times more expensive to haul a boat out of the water for hull cleaning than to clean in the water. There is in-water technology for anti-fouling for large boats but not for small. Regarding cleaning stations, who would fund, and how would they be distributed around the state? Recommended the development of remedies for specific behaviors – fishing boats, trans-Pacific yacht racing. Recommends going slow to curtail copper based hull paint as it slows down hull fouling.
- Dave Breninger- General Manager Placer Co Water Agency; Director ACWA; Director RBOC: Concern with water quality issues (agricultural water and the delta). Need to link water agency and boating concerns (Objective 2I). His water district is plagued with non-natives. Need to eradicate in waterways. Likes use of native plants. Need to make recreational boaters part of the solution. *Egeria* should be eradicated. Need a positive way to put money into solution.
- Duane L. Schnabel, Primary State Biologist, DFA: Although the plan cites NEPA/CEQA in Appendix B there is no discussion of when an EIR will be done for the plan. People need to know if the actions will do more harm than good.
- Ted Grosholz, Dept. of Environmental Science & Policy, UC Davis: He is a cooperative extension researcher who developed the initial plan. The plan as written has an absence of university and research institution participation. The plan ignores non-agency participants in AIS work. Action 1A6 calling for a data base of AIS projects ignores already existing National Biological Species nodes at UC Davis and UC Santa Barbara. Actions 6A3-11 ignores cooperative extension and sea grant work in progress for years. Actions 7A1-2 to complete AIS studies ignores work under development at the universities. The plan needs to bring the University of California into the management plan. The Ocean Protection Council endorsed a university inter-agency center for AIS and the center is not included as part of this plan. The center needs to be part of the plan and needs to be stated explicitly.
- Rick Grosberg, Center for Population Biology, UC Davis: The threat of AIS was identified by the research community and not state agencies. The document completely ignores the contributions of the research community. UC Davis formed an AIS council that is not included or even mentioned. The management framework includes only agency leaders who will meet (When? For What?). The Document needs to integrate geographically and biologically. It does not provide a management framework for integration at all levels. There is a missing objective for coordination of research problems, ecological problems, biological problems; the structure for coordination is not listed as an objective. Document does a good job identifying problems but fails in coordination and development of policy.
- Rebecca Verity - UCOP: UCOP supports the University of California and CSU's disappointment at being left out of the plan. The state constitution designated the University of California as the research arm of the State of California. The university was

told there would be an AIS Center for coordination of research, surveys and development of new tools. All faculty were told the bones of the center would be in the management plan. They are very disappointed it is not.

- Jodi Cassell, Sea Grant: Has been involved in outreach and applied research on AIS. Jodi herself has been involved for 8 years. They are also a funding source having funded \$1,800,000 in AIS projects and outreach. They are very disappointed the plan ignores all non-state agency work related to AIS. Sea Grant is not mentioned at all in the plan although they have done extensive outreach on ballast water management, newsletters, research on hull fouling, transport vectors, establishing a network of advisors, etc. She feels the agency role should be to coordinate ongoing programs. DFG is not in outreach; outreach is not a strong component of their mandate. She felt the plan needed to use existing resources and not push them out of the management plan.
- Elaine Sledge, National Paint and Coatings Association: The association concurs with the plan findings on the threat of AIS. They support prevention vs. control and eradication. Coatings must have copper for anti-fouling. Inter-coastal vessels transport AIS. There are also non-ballast vectors. Non-biocide coatings are preferred. Written comments will provide additional information.
- Ron Eng, DFA: Action 211 proposes adding staff and hours at DFA Border Protection Stations with no indication of how this would be funded.
- Clint Meyer, Project Manager, Michael Brandman Associates: There is already a good regulatory program through CEQA. CEQA should be updated to address terrestrial and aquatic invasive species.

Public Meeting #2 (Oakland) Summary

The meeting was held August 30 in the Association of Bay Area Governments / Metropolitan Transportation Commission's conference room and had nearly 25 attendees. The following comments and questions were raised:

- Karl Malamud-Roam, Mosquito Ecologist, Contra Costa Mosquito VCD: He stated the regulatory aspects on control of public health were good. AIS present a huge problem. Insects and the diseases that come with them require continual surveillance and rapid response which the districts have in place. There is confusion in tone in the introduction; the plan treats non-native species and invasives as synonymous. The definition of invasives is not clear; the federal definition emphasizes harm (as stated in first paragraph) but the second paragraph treats all non-natives as invasive. It should not assume that non-native is detrimental; there are benefits of non-natives. The mosquito fish comments need correcting. There is a presumption that mosquito fish are known to harm; be careful of context of usage. They are a tool for resource management.
- Steve Hajik, Lake Co. Dept. of Agriculture: Spraying requires a permit from the regulatory water agency. County only allows licensed sprayers and inspects all applicators. His county passed an ordinance that lists banned weeds. He commented the plan should not forget agricultural commission offices.
- Caitlin Sweeney, SF Bay Conservation & Development Commission (BCDC): There is a critical omission of BCDC in the management plan. They have enforceable policies on fill, dredging, tidal marsh restoration projects and require eradication permits in their jurisdiction.
- Doug Johnson, California Invasive Plant Council: Plan needs to emphasize the impacts of chemical treatment as well as the impacts of all treatments. High level coordination

under Strategy 1A should include agricultural and environmental groups; should be strengthened to advocate for AIS council not partitioned as aquatic, but all inclusive.

- Cathy McGowan, Office of Research, UCOP: Lawrence Coleman, Vice Provost will submit detailed comments in writing. Cathy read from a 4-page document (attached) with preliminary comments. Solutions must be cross-cutting; researchers, policy makers and managers must work together. There must be formation of a California Center for Invasive Species; UC supports this strongly and wants it added to the plan. The plan needs to include members of UC and Sea Grant on the CAAIST (1A2). The section on Education and Outreach needs to include the UC Riverside Aquatic Center and Sea Grant Extension outreach. The education of ongoing researchers needs to be added. Section 7 provides an excellent start but needs to be expanded to include an academic research center.
- Mike Connor, Executive Director, San Francisco Estuary Institute (SFEI): SFEI has been working on biological invasions for over a decade. The rate of invasions is increasing; at present they are working on a multi-agency rapid response effort to eradicate invasive oysters in the South Bay. The report needs three things; 1) transparency; 2) peer review; and 3) competitive funding. First, transparency, the public cannot figure out who is working on what and therefore cannot determine overall success. Second, there is no call for outside peer review, which is necessary to insure that implementation is up to date. This is crucial for incorporation into the report. Third, there should be provision for competitive funding of line items in the document. Funding should go through a competitive process to insure transparency and the best quality work.
- Cathy Roybal, Contra Costa Dept. of Agriculture: Local county agricultural offices need to be involved.
- Karl Malamud-Roam, Mosquito Ecologist, Contra Costa Mosquito VCD: Department of Health Services needs to be added to agencies; the Health & Safety Code needs to be added to statutes. Use of vector should be carefully defined; conventional use includes mosquito control. The Society of Wetland Scientists was the first concerned with invasive cord grass. Strong kudos for rapid response.
- Arthur Berlowitz, U.S. Department of Agriculture (USDA): Goal is to prevent invasive species if we can. USDA reviews plants for the aquarium trade. He does not see how USDA can interface with the plan; it is not clear how USDA fits in. Thinks a center is a great idea. Document should show who has jurisdiction over what part of invasive species control.
- Sarah Mannell, Mill Valley, CA: She wants to know who does the public contact about invasive species. There are large carp in Corte Madera Creek; a protected creek; with steelhead fry in their guts.

Public Meeting #3 (Long Beach) Summary

The meeting was held at the Port of Long Beach Board Room on September 1 and had eight attendees. While no comments were presented, meeting attendees did raise the following questions:

1. How does the plan articulate agricultural invasive plants?
Answer: DFA is on the coordinating committee. The committee also worked with DPR.
2. For the Technical Advisory Committee, will there be one for the state, or will there be regional panels to focus on the issues for that region?
Answer: Having regional coordinating panels is a good suggestion and will be considered during finalization and/or implementation of the plan.

3. How much public outreach was there for these public meetings? He did not see a full press announcement.
Answer: There was targeted outreach to the OPC mailing list, stakeholder groups, web sites and DFG did a press release.
4. Is this a modification of an existing plan or a new plan?
Answer: It is restructured and rewritten from an earlier draft.
5. How is the SFEP associated with the project?
Answer: SFEP was contracted for one year by the SCC with funding from the Ocean Protection Council to finish the state AIS Plan.
6. Has there been outreach to shipping companies?
Answer: SLC, which is in charge of the ballast water program, has been keeping shipping up to speed. The ballast water recommendations were taken from the proposed actions. The plan basically looks at vectors other than shipping.

Prioritization Exercise Results

At all three meetings, posters were provided on the walls for attendees to indicate which action items (as described in the Draft AIS Plan) they believed should be “high” and “low” priority. After the close of the public comment period, meeting attendees held informal conversations with project staff and added to the posters. The posters were brought to each subsequent meeting to allow attendees to see which action items other individuals had prioritized.

One action was identified as extremely important: 8A3. Pursue the authority for DFG and DFA to establish a Rapid Response Program.

The following information was collected. Some of the action numbers changed as comments were addressed and the draft plan was finalized. The action numbers below were updated to reflect the new numbers; some of the original actions were deleted or moved in the editing process. Some of the action language has been edited since this summary was made. A few actions are listed as both high and low priorities because of differing opinions among participants. For final priorities identified see Chapter 8.

Objective 1: Coordination & Collaboration

High Priority Actions

- 1A1. Develop an executive level consultation process.
- 1A2. Form the California Agencies AIS Team (CAAIST).
- 1A7. Identify lead state agencies for particular AIS, water bodies and invasion vectors.
- 1A8. Identify agency personnel required for AIS management.
- 1A9. Improve state websites related to AIS.
- 1A10. Assess effectiveness of and gaps in AIS programs.
- 1B4. Expand participation in local AIS efforts and task forces.
- 1B5. Expand participation in regional, national and international AIS task forces.
- 1B7. Participate in national and international conferences.
- 1C2. Establish stable, long-term funding to help implement this plan.
- 1C3. Provide state funding for the AIS positions.
- 1C4. Provide state funding for a rapid response program.
- 1C5. Hire a funding development specialist.
- 1C6. Provide new funding mechanisms.

Low Priority Actions

All remaining actions for this objective not shown as high priority above.

Objective 2: Prevention

High Priority Actions

- 2B1. Quantify the ballast water and hull fouling vectors and assess invasion risk.
- 2B2. Continue and improve state ballast water inspection and enforcement program.
- 2B3. Implement discharge standards for treated ballast water.
- 2B4. Identify and address gaps in the Marine Invasive Species Program.
- 2B7. Quantify and assess the role of commercial fishing vessels as AIS vector.
- 2C1. Quantify and assess the role of recreational boating as an AIS vector.
- 2C2. Develop a recreational boating outreach and management program.
- 2C3. Develop a watercraft inspection program for high priority boat launch sites.
- 2C4. Quantify and assess the role of recreational fishing as an AIS vector.
- 2C5. Develop a recreational fishing outreach and management program.
- 2C6. Develop guidelines for: disposal of invasive species, cleaning of gear disposal of live bait.
- 2D1. Quantify and assess live bait as an AIS vector.
- 2E1. Quantify and assess fisheries enhancement as an AIS vector.
- 2I1. Increase staffing and hours of operation at DFA Border Protection Stations.
- 2I2. Develop guidelines for border inspections.
- 2I3. Increase DFG enforcement of current regulations on prohibited and restricted species.
- 2I4. Ensure adequate staffing and cargo inspection guidelines at ports and airports.
- 2I5. Continue disease sampling for shipments and stocks of live aquatic species.
- 2I6. Identify mail order, online vendors selling CA prohibited and restricted species.

Low Priority Actions

- 2B3. Implement discharge standards for treated ballast water.
- 2B4. Identify and address gaps in the Marine Invasive Species Program.
- 2C. All actions mandating hull cleaning and/or inspections.
- 2C10. Link activities to the national Stop Aquatic Hitchhikers campaign. (Action later deleted).
- 2E4. Weigh benefits of mosquito-fish introductions.

Objective 3: Early Detection & Monitoring

High Priority Actions

- 3A1. Assess current monitoring of the state waters for early detection opportunities.
- 3A3. Develop statewide approach to early detection.
- 3A4. Outreach to those regularly sampling state waters.
- 3A5. Create and train a statewide citizen monitoring network.
- 3B1. Assess long-term AIS monitoring of state waters.
- 3B3. Monitor locations with high invasion rates.
- 3B7. Review the efficacy of long-term monitoring systems.

Low Priority Actions

- 3B6. Include maps of existing AIS in California waters in DFG BIOS system.

Objective 4: Rapid Response & Eradication

High Priority Actions

- 4A1. Develop and implement a statewide rapid response plan.
- 4A2. Evaluate and coordinate existing systems for reporting AIS sightings.
- 4A3. Clarify among the agencies and organizations involved who is responsible for which areas and/or species. (This action from August '06 draft has been deleted. It will be addressed through current actions 4A1 and 4A3).
- 4A4. Explore permanent funding to implement rapid response.
- 4B1. Review effectiveness of eradication programs.

- 4B2. Continue and complete current eradication efforts.
- 4B3. Standardize criteria for identifying priority species for eradication.
- 4B4. Develop a method to prioritize sites of AIS invasion concern.

Low Priority Actions

All remaining actions for this objective not shown as high priority above.

Objective 5: Long-Term Control & Management

High Priority Actions

- 5B All strategy action items; limit the dispersal of established AIS to new water bodies.
- 5C2. Coordinate entities to meet AIS protection and restoration objectives.
- 5C6. Assess guidelines for preventing AIS spread in habitat restoration and shoreline landscaping projects. (See 6C5)

Low Priority Actions

- 5B1. Establish boat washing stations and disposal facilities at infested water bodies.
- 5B3. Use volunteer monitors to conduct AIS inspections.

Objective 6: Education & Outreach

High Priority Actions

- 6A1. Inventory education and outreach efforts. Develop a state AIS communication strategy.
- 6A2. Partner with ongoing outreach campaigns.
- 6A4. Develop posters, brochures and articles for industry sectors and user groups.
- 6A5. Develop permanent interpretive displays at marinas, boat ramps, and fishing sites.
- 6A6. Work directly with industry trade shows to deliver the AIS message.
- 6A7. Present AIS information at public gatherings.
- 6A8. Include AIS information in state hunting, fishing and boating regulations and licenses.
- 6A9. Include AIS information in fishing and recreational publications.
- 6A10. Develop and distribute AIS identification cards
- 6A11. Encourage industries to offer noninvasive alternatives to AIS.
- 6A12. Partner with stakeholders and interest groups to broaden education efforts.
- 6A13. Educate waterfront and shoreline property owners about AIS.
- 6A14. Develop and offer AIS management classes for professional organizations.
- 6A15. Continue state education measures concerning ballast water.
- 6C2. Educate researchers on AIS containment, disposal methods and legal restrictions.
- 6C5. Disseminate guidelines to promote the use of native plants. (See 5C6)

Objective 7: Research

High Priority Actions

Note: suggestion was made to add “increase coordination of researchers and develop research agenda based on high priority research needs.”

- 7A1. Host workshops to develop AIS research priorities and identify gaps.
- 7A2. Assess, continue and complete current studies.
- 7A3. Develop a strategy to communicate and support research needs.
- 7C4. Identify opportunities for interagency funding of AIS management research.

Low Priority Actions

- 7C3. Consider test center to evaluate ballast water treatment technologies.

Objective 8: Policy

High Priority Actions

- 8A1. Establish a regulatory review committee.
- 8A2. Identify the potential for improved regulatory coordination.
- 8A3. Pursue the authority to establish an interagency rapid response program.
- 8A4. Explore the need for additional state authority for AIS management.
- 8A6. Review current system for regulating plant and animal importations. .

Low Priority Actions

None indicated

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*Mtg. = meeting locations

Sac – Sacramento, August 28, 2006

Oak – Oakland, August 30, 2006

LB – Long Beach – September 1, 2006

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*Mtg. = meeting locations

Sac – Sacramento, August 28, 2006

Oak – Oakland, August 30, 2006

LB – Long Beach – September 1, 2006

2004 Draft Plan Process

The first draft of the AIS management plan included the valuable input of many dedicated individuals with expertise on a wide variety of topics relating to AIS in California and the region. Contributors ranged from local, state and federal agencies, to industry representatives, NGOs and other stakeholders.

Funding for the development of the first draft was provided by the DFG and USFWS. Susan Ellis, the Statewide Invasive Species Coordinator, developed a contract with the University of California, Davis, to develop an Aquatic Invasive Species Plan following the general outline provided by the Aquatic Nuisance Species Task Force. Ted Grosholz was the Principal Investigator for the contract. The deliverables for the contract included facilitated meetings to ensure that agency and stakeholder input was incorporated in the Plan.

In August of 2002, representatives of 14 agencies with a role in managing aquatic invasive species came together to participate in a State AIS Planning Workshop in Davis, CA. Results of that meeting included a draft set of goals and objectives for an AIS Plan and a brief summary of current AIS activities for some of the participating agencies. There was agreement that a state plan could help identify AIS of concern, and provide a framework for how to address AIS prevention, eradication, research, management and education and outreach in a more coordinated and comprehensive fashion.

Additional information for the plan was gathered from other state and federal plans, various websites, published papers, internal agency documents and through personal communication (phone and email).

The Plan's Review Committee (members listed below) commented on a first draft of the plan, which was then distributed to a broader group of Agency reviewers and for public review.

Review Committee for the 2004 Draft Plan

Lars Anderson, United States Department of Agriculture, Agricultural Research Service
Robert Leavitt, California Department of Food and Agriculture
Dale Steele, California Department of Fish and Game
Mark Sytsma, Portland State University
Erin Williams, United States Fish and Wildlife Service

Participation by Other Agencies and Groups

Courtney Albrecht, California Department of Food and Agriculture
Marcia Carlock, California Department of Boating and Waterways
Marina Carzola, California Coastal Commission
Jason Churchill, Lahontan Regional Water Quality Control Board
Nate Dechoretz, California Department of Food and Agriculture
Joseph DiTomaso, University of California, Davis
Maurya Falkner, California State Lands Commission
Connie Ford, State Water Resources Control Board
Joann Furse, California Sea Grant
Eric Gillies, California State Lands Commission
Bob Hoffman, National Marine Fisheries Service
Christina Johnson, California Sea Grant
Jaime Kooser, California Coastal Commission
Steve Lonhart, Monterey Bay National Marine Sanctuary
Karen McDowell, California Sea Grant
Cindy Messer, California Department of Water Resources
Julie Owen, California Department of Boating and Waterways
Bill Paznokas, California Department of Fish and Game
Stephen Phillips, Pacific States Marine Fisheries Commission
Carolyn Pizzo, U.S. Department of Agriculture

Jim Rains, California Department of Food and Agriculture
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Linda Sheehan, Pacific Regional Office, The Ocean Conservancy
Basia Trout, Bureau of Reclamation
Tanya Veldhuizen, California Department of Water Resources
Kim Webb, United States Fish and Wildlife Service
Katherine Zarembo, Invasive Spartina Project

2002-2003 Stakeholder Meeting Comments

Incorporating recommendations from a broad array of stakeholders contributes to a better and more responsive AIS plan for the State of California. In an effort to get input on concerns and perspectives regarding AIS during the plan's development, scoping meetings were held to get input from many organizations, businesses, industry representatives and individuals. A northern California stakeholder meeting was held in Sacramento on November 19, 2002. A southern California stakeholder meeting was held on March 20, 2003. Participants provided valuable comments, most of which have been incorporated into the management plan.

Northern California Stakeholder Comments

Invitations were sent to over 200 individuals and included representatives of many industries including the pet, aquarium, and nursery/landscaping trades, live bait and seafood dealers, and ports and marinas. The following individuals attended:

Drew Alden, Growers in Tomales Bay
John Berg, Pacific Merchant Shipping Association
Thomas Confal, IPM Specialist, Bitterroot Restoration, Inc.
John Cruger-Hansen, Harbor Master, City of Antioch
Daniel Garcia, Public Affairs, Marine Aquarists Roundtable of Sacramento
Jeff Hart, President, Habitat Assessment and Restoration Team, Inc.
James Kidder, President, Colombo Bait, Inc.
Karen McDowell, Project Coordinator, West Coast Ballast Outreach Project
James Mills, Vice President and Regional Manager, Westree Marinas
Fleur O'Neill, Policy Education Coordinator, Save Our Shores
John O'Sullivan, Curator of Field Operations, Monterey Bay Aquarium
Roger Phillips, Applied Research Manager, Monterey Bay Aquarium
Kirsten Upson, The Nature Conservancy
M.K. Veloz, Administrative Director, Northern California Marine Association

Mike Fraidenburg of Dynamic Solutions Group of Olympia, Washington facilitated the meeting. Susan Ellis (State Invasive Species Coordinator) explained the different roles and responsibilities of state agencies and current management activities for aquatic invasive species in California. Ted Grosholz (UCD) and Holly Crosson (UCD) discussed the process for the plan's development including future stakeholder and agency meetings as well as the current status of the plan. Mark Sytsma (Portland State University, Portland, Oregon) discussed Oregon's experience with writing a state management plan for aquatic invasive species as well as the uses and limits of state plans. The rest of the meeting was spent listening to concerns and suggestions presented by the stakeholders. Most of the comments could be divided into the categories of Education, Prevention, Best Management Practices, Regulation, State Invasive Species Council and General AIS Management Plan development suggestions.

EDUCATION

- Education about AIS should be a top priority.
- Educational tools should be used instead of legislation and regulations.
- A list of AIS experts should be made available to stakeholders.

- AIS information should be available at all bait shops, marinas, boat access areas, etc.
- It may take 20 years, but *all* of the public needs to be educated about AIS (example used was educational programs for dealing with issues such as recycling, littering, etc.).
- The public needs to know why they should care about AIS (i.e., the consequences of invasions).
- The public as well as industry needs to know the economic cost of AIS (cost/benefit analysis).
- Stakeholders are a resource and can help with education, such as public service announcements.
- Multiply educational efforts by identifying what industry sectors can do to help with AIS education and outreach (i.e., using Wal-Mart, Home Depot, PetSmart etc. to educate their customers about AIS).
- A database is needed that focuses on providing information about AIS outreach, education and research-based grants. Information on who is doing what on AIS should also be available and include efforts by NGO's, universities and industry.
- AIS hazards that exist in particular areas need to be identified and publicized before they spread.
- Cross-education between interest groups and government would help understanding of the issues and concerns for both groups.
- Education in the K-12 classroom is important; biologists should go into schools to talk about AIS.
- Aqua-culturists need current information to help avoid AIS introduction problems of the past.
- There should be guidelines developed to help groups "self-police" and educate their constituents.
- Coordination needs to be improved between state, regional and federal groups.
- Identify all educational and technical resources currently available and make them easily accessible.
- Identify where the information gaps are.

PREVENTION (including Early Detection and Rapid Response)

- A Rapid Response program requires extensive coordination but is critical.
- An AIS "hotline" is needed so new sightings can be reported immediately.
- Management of introduction pathways is important for AIS prevention.
- We should have the ethic of not transporting California's AIS elsewhere; include this in the plan.
- The largest percentage of funds should be spent on prevention since it is the most cost-effective.
- Early detection is key to successful AIS eradication and management.
- Each vector/pathway that is identified in the plan should have a lead agency listed as well as a stakeholder group.
- Look into whether funds from anti-terrorism sources could be tapped into (i.e. to address the intentional introduction of a devastating foreign, water-borne organism).

BEST MANAGEMENT PRACTICES (BMPs)

- Each industry should be actively involved in the development of the BMPs that relate to them.
- BMPs can be a tool for industry to understand and meet their obligations.
- Consider using a neutral third party or group (scientific panel) to offer advice and develop recommendations for BMPs instead of leaving development to agencies or industry alone.
- Investigate how "management" of a landscape (or lack thereof) affects the likelihood of invasion.

REGULATION

- The public and industry need to have an understanding of AIS laws and their history before they go into effect.
- We need more education and outreach on laws already passed so the public can abide by them.
- AIS laws and penalties need to be publicized in the DFG regulations right up front.
- Regulatory agencies need to "get on the same page"; inconsistencies confuse the public.
- There should be more opportunity for stakeholder input when new regulations are being written, especially when livelihoods are at stake (*Caulerpa* in southern California was example used).
- A patchwork of regulations makes coordination between state, regional and federal levels difficult.

- Inter-jurisdictional coordination needs improvement to make compliance easier.
- Guidelines need to be developed for meeting NPDES permit requirements.
- A process needs to be developed to authorize within-state transfer of approved live aquatic species.
- Laws, regulations and permits need to be more clear, consistent and effective.
- Enforcement needs to be more vigilant and consistent.
- Stakeholder input should be solicited when permitting procedures are being written.
- New legislation should be written with the help of stakeholders (ballast water example was used).
- Methods for complying with aquaculture regulations need to be clearer.
- Some stakeholders feel like they are working in a vacuum; they need guidelines to help them determine if the right thing is being done.
- Develop a mechanism for mandatory reporting of listed AIS.
- Make sure regulations that affect industry are feasible (shipping example was used).
- Use existing Department of Boating and Waterways (DBW) laws to make AIS introductions illegal.
- Create a single, central clearing house for information on all AIS laws and regulations.

STATE AQUATIC INVASIVE SPECIES COUNCIL (ISC)

- The ISC needs to have broader public representation; consider expanding it to include more stakeholder groups.
- Each industry should decide who will represent them on the ISC.
- The number of industry representatives should be equal to or higher than the number of government representatives on the ISC.
- DBW should not represent all boating interests on the ISC.

GENERAL AIS MANAGEMENT PLAN DEVELOPMENT

- Make the plan short and simple.
- Funding priorities in the plan should be delineated by the ISC or another representative group.
- Work together; don't have government on one side and resource users on the other.
- Stakeholders are interested in practical solutions.
- Use common names in addition to scientific names for AIS to make the plan more user-friendly.
- Limit use of acronyms or fully explain them.
- Prioritization of species within the plan is necessary.
- Develop a system to prioritize aquatic invasive species using the ISC or another representative group.
- Use assigned "Management Classes" as Oregon did rather than prioritizing species.
- Consider using DFA's ABC List of Noxious Weeds as a model.
- Develop a process to determine which method gets used to control or eradicate a species.
- Limit administrative overhead.
- Develop a process to resolve disputes.
- Make sure all groups are represented (include tribes, irrigation districts, bass anglers, boaters, etc.).
- The planning effort should take into account the target species as well as the environment.
- There is a concern that some may try to sidetrack the plan or use the plan to push their own agenda.
- Consider using AIS instead of ANS (the word "invasive" is perhaps better than "nuisance").
- Write into the plan that state and federal agencies coordinate through formal written agreements.
- High profile species should not take over concern for lesser-known problem species.
- Support for current AIS programs should be continued.
- Make sure limited resources go to on-the ground projects rather than getting lost in the bureaucracy.

Southern California Stakeholder Comments

Invitations were sent to over 450 individuals and included representatives of local water agencies and irrigation districts, tribes, various industries including the pet, aquarium, aquaculture and nursery/landscaping trades, live bait and seafood dealers, ports, marinas and shippers, and others with an interest in aquatic invasive species. The following individuals attended:

Douglas Ball, Los Angeles Department of Water and Power
Mark Baumann, Live Cargo Reptile and Fish/ San Diego Fish Society
Paul Brown, Project Analyst, Port of San Diego
Thomas Buckowski, Lake Biologist, Lake Mission Viejo Association
Larry Chapp, Vice President, Divisional Merchandise Manager, PETCO
Hugh Cobb, Pacific Coast Bait and Tackle
Tom Gass, Manager, El Pescado Caliente
Chris Graham, Lake Biologist, Lake Mission Viejo Association
Miguel Hernandez, Watermaster, Natural Resources Office, Pauma Band of Mission Indians
Annaliese Hettinger, The Diving Locker
Steve Lonhart, Monterey Bay National Marine Sanctuary
Marshall Meyers, Executive Vice President, Pet Industry Joint Advisory Council
Craig Parsons, Live Fish, Reptile, Bird and Small Animal Buyer, PETCO
Russell Moll, Director, California Sea Grant/ Scripps Institute of Oceanography (SIO)
Anandra Ranasinghe, Southern California Coastal Water Research Project
Freda Reid, San Dieguito Lagoon Committee and Research Associate (SIO)
Andi Shluker, The Nature Conservancy of Hawaii
Ed Smith, General Manager, Palo Verde Irrigation District

Mike Fraidenburg of Dynamic Solutions Group (DSG) of Olympia, Washington facilitated the meeting. Ted Grosholz (UCD) discussed the ecological and economic costs of aquatic invasive species and introduced the goals and purpose of the meeting. Susan Ellis (State Invasive Species Coordinator) explained the different roles and responsibilities of state agencies and current management activities for aquatic invasive species in California, and provided an update on the formation of the California Aquatic Invasive Species Council. Mark Sytsma (Portland State University, Portland, Oregon) discussed Oregon's experience with writing a state management plan for aquatic invasive species as well as the uses and limits of state plans. Holly Crosson (UCD) discussed the process for the California plan's development and progress on the plan thus far. The rest of the meeting was spent discussing concerns and suggestions presented by the stakeholders. Most of the comments could be divided into the categories of Education, Prevention, Best Management Practices, Regulation and General AIS Management Plan development. Below is a summary of specific comments made under each of these categories.

EDUCATION

- A comprehensive strategy for AIS Education and Outreach should be developed.
- Education should be used instead of new legislation and regulation.
- More AIS information needs to reach the public, retail stores, industry, schools, etc.
- Prioritize educational efforts based on risk associated with a given pathway.
- Piggyback onto current Agency educational programs.
- Consider "green labeling" to help consumers make the right choice; peer pressure will encourage appropriate behavior/decisions of others.
- Educational efforts need to take into account the multi-cultural nature of CA (signs, etc. need to be published in other appropriate languages besides English).
- Marketing experts should be used to get a single, common AIS message out across the region.
- The AIS message has to touch people personally (an impact on the quality of life or the pocketbook).
- Educational materials should be tailored to specific industry sectors (aquaculture, boaters, bait shops, pet/aquarium retailers, etc.).

- The public as well as industry needs to know the economic cost of AIS (pay now or pay more later).
- Stakeholders are a resource and can help with educational efforts (i.e., using Recreational Fisherman's Alliance, American Sportfishing Association, Diving or Tropical Fish Clubs, etc.).
- Multiply educational efforts by identifying what industry sectors can do to help with AIS education and outreach; partner with pet/aquarium and other industries.
- Develop better ways to get the AIS message out, for instance, don't just have a booth at trade shows but work directly with promoters of shows (example – Fred Hall Show).
- Publish articles in Western Outdoor News and similar magazines.
- Train people to use the AIS "Traveling Trunk" and have them take it "on the road".
- A comprehensive AIS species list should be developed and publicized with appropriate contacts listed for experts associated with each species.
- There should be guidelines developed to help groups "self-regulate" and educate their constituents.

PREVENTION (including Early Detection and Rapid Response)

- An AIS Prevention Program is key to success but is not foolproof.
- AIS Screening and Risk Assessment Programs should not be overly simplistic or arbitrary. They need to be based on the best available information and sound science.
- Volunteers can be an important piece in monitoring efforts for early detection of AIS.
- Training volunteers takes a lot of organization and keeping them motivated over the long term can be challenging
- Interaction with Watershed Councils is important.
- An AIS "hotline" is needed so new sightings can be reported immediately.
- Determine the economic consequences of pathway prevention.
- Look into funds available through "homeland security".

BEST MANAGEMENT PRACTICES (BMPs)

- Develop guidelines for acceptable, humane and environmentally safe ways to deal with unwanted aquatic organisms (whether it be proper disposal, returning the organism to the retailer, or being “adopted” by someone else).
- Industry and individuals need to accept a degree of economic liability and responsibility for their actions regarding AIS introduction and spread.
- Create industry standards to regulate and penalize the bad actors.
- Each industry should be actively involved in the development of their own BMP’s. Weak industry initiative yields weak BMPs.
- Industry documentation is needed to support accountability.
- Determine if BMPs should be regulatory.
- Develop BMPs for Bass Tournaments.
- BMPs need to maintain some flexibility and an acknowledgement that “one size does not fit all”.
- BMPs can help achieve buy-in, create institutional memory, give an outsider a way to monitor activities and are already an accepted process in industry (similar to ISO example).

REGULATION

- Enforce the laws and regulations we already have, rather than pass new ones.
- Provide positive incentives to encourage self-regulation.
- Provide better information about what AIS laws are currently in place and how to comply with them.
- A few bad apples are causing regulatory problems for all involved.
- Determine more effective ways to catch violators of current laws, including interstate transport.
- Improve current regulations. Piranhas and snakeheads were used as examples of species that are regulated but still are imported and released. We should learn from these experiences and attempt to prevent similar situations.

GENERAL AIS MANAGEMENT PLAN DEVELOPMENT

- Coordinate with the National Marine Sanctuaries on Plan development.
- Work with California Sea Grant to achieve success in plan implementation, especially with education and outreach strategies and actions.
- Be creative with funding and partnerships.
- Leverage resources by doubling up on surveys, inspections, etc. that are already being done.
- Continually evaluate and update the plan and make sure the plan’s goals are being realized (develop a scorecard).
- Make sure the functioning of the California Aquatic Invasive Species Council is evaluated so it does not outlive its useful purpose. If changes are needed to make the council more effective, they should be able to be promoted through other agencies and the general public.
- Take steps to minimize the loss of dollars through overhead.
- Do not set the stage for failure by creating a timeline that cannot be met.
- Involve economists if possible (can a dollar figure be put on habitat/resources?).
- Make it clear who will determine priorities in the plan and what gets funded.
- Incorporate Watershed Councils in the planning effort.
- Make the relationship between the plan and AIS policy clear.
- Determine how plan implementers will interact with on-the-ground managers.
- Write the plan so that it facilitates funding for implementation. The plan should be user-friendly.
- Plans should promote accountability so that managers have an incentive to perform and meet commitments.

APPENDIX F: EXECUTIVE SUMMARY OF *BIOLOGICAL INVASIONS:* *RECOMMENDATIONS FOR U.S. POLICY AND MANAGEMENT*

Position Paper of the Ecological Society of America

Biological Invasions: Recommendations for U.S. Policy and Management

David M. Lodge, Susan L. Williams, Hugh MacIsaac, Keith Hayes, Brian Leung, Sarah Reichard, Richard N. Mack, Peter B. Moyle, Maggie Smith, David A. Andow, James T. Carlton and Anthony McMichael, 2006

Executive Summary

The spread of nonindigenous (non-native) species introduced into the United States is a significant and growing national problem, costing taxpayers hundreds of billions of dollars in environmental degradation, lost agricultural productivity, increased health problems and expensive prevention and eradication efforts. Some nonindigenous species are introduced intentionally and are highly valued by humans, e.g., agriculture, aquaculture, and ornamental species. Many other species are introduced as by-products of human activity, especially through the increasing global transportation of humans and commercial goods. A subset of introduced species spread widely, become abundant and cause harm. The definition of “harm” is a function of human values, which often differ in different regions and may change temporally. Nevertheless, harm is often unambiguous and the species from elsewhere that causes harm are referred to as invasive nonindigenous species. They are the focus of policy and management concern because of their serious and complex contributions to diseases of plants, animals and humans; reductions in native species; changes in ecosystem function; and financial losses.

Well known examples of invasive nonindigenous species include the vine kudzu (*Pueraria lobata*) in the southeastern U.S., cheat grass (*Bromus tectorum*) in the western U.S., and zebra mussel (*Dreissena polymorpha*) in the central U.S. More recent arrivals with large net negative impacts on the environment, agriculture, forestry, industry and human health include West Nile virus, the seaweed *Caulerpa* (*Caulerpa taxifolia*), Asian long-horn beetle (*Anoplophora glabripennis*), emerald ash borer beetle (*Agilus planipennis*), sudden oak death (*Phytophthora ramorum*), monkeypox virus, and the SARS virus. Without management, the populations of these species grow and spread such that damages accelerate over time. In contrast to many other forms of pollution, such widespread invasions become irreversible because the technology often does not exist to selectively eradicate species. Relative to the economic and ecological costs of other forms of environmental pollution, the costs of nonindigenous species are therefore of particular concern because they are likely to be borne over very long time frames.

Despite the great diversity of invasive species and their impacts, an identified group of pathways transport species, and a common set of biological processes – introduction, establishment, spread, and impact – operate in all invasions. Policy and management solutions become clearer when these common pathways and processes are recognized. Nevertheless the possible management responses diminish as any invasion progresses. Prevention is possible only before a species arrives or at the point of entry. Thereafter, a narrow window of opportunity for eradication exists before some species spread so widely that it is impossible or infeasible to locate and kill all populations. Once a species is too widespread for eradication, only three management options remain: controlling populations in selected locations; active mitigation of impacts; or simply bearing the cost of the changes caused by the invader. U.S. policy, often by default, has largely adopted the last option, i.e., acceptance of often irreversible environmental and economic damage.

The only study to attempt a nationwide estimate of the economic costs to the U.S. of nonindigenous species concluded that annual costs exceed \$120 billion (Pimentel et al. 2005), which

we regard as an underestimate because the majority of invasive species were not included in the study. Even this underestimate equates to costs of \$1,100 per U.S. household per year, costs that will continue to grow unless prevention and management of invasive species improves. Yet, the U.S. has allowed invasions to continue and damages to increase.

A more cost-effective approach would include greater investments in prevention and other active management steps, including early detection, eradication and control. Recent scientific advances in our understanding of biological invasions make it clear that more effective options exist for these threats. Here, on behalf of the Ecological Society of America, we make six recommendations for government action that, if implemented, would substantially reduce the current and future damages to the U.S. from invasive species. We include proposals for cost-effective government actions that will address these problems with the understanding that other measures are important to complement governmental responses. Key challenges that require urgent government action include prevention, detection, eradication and control of harmful non-native species, and the coordination of these efforts at the state, federal and international levels. Table 1 summarizes the major recommendations, data and techniques for implementation, and proposed lead organizations.

Prevention

Recommendation 1. Use a combination of existing and new technologies, education strategies, industry codes of conduct, and government oversight to prevent introductions from pathways that already are well known to be major sources of nonindigenous species, and to monitor other pathways into the United States to better assess the degree of risk they pose.

Recommendation 2. Screen live organisms proposed for importation into the U.S. for environmental, economic and human health risk before a decision is made to allow entry. Risk analysis tools should be repeatable, transparent, supported by current scientific findings and applied to all pathways, across all agency jurisdictions.

Early Detection, Eradication and Control

Recommendation 3. Use new technology to improve active surveillance of invasive species to increase the success of rapid response and eradication efforts, in cooperation with existing web-based information networks in universities, herbaria, museums and state agencies.

Recommendation 4. Make legal authority and emergency funding available for eradication and control to proceed rapidly once a newly established potentially invasive species is detected. Current legal mechanisms and funding for responses to agricultural pests and parasites, and to human pathogens, should be extended to all potentially invasive species in all habitats, and employed commensurate with the threat.

Recommendation 5. Provide on-going funding and incentives for slowing the spread of established invasive species on public and private lands, in cooperation with the states and tribal governing bodies.

Establishing a National Center for Invasive Species Management

Recommendation 6. Expand existing authority of the National Invasive Species Council (NISC), including the establishment of a National Center for Invasive Species Management under NISC, to better coordinate policies among government agencies and with other countries. Current U.S. examples of intergovernmental cooperation include the National Interagency Fire Center and the Center for Disease Control and Prevention. Unless these or conceptually similar recommendations are adopted, the rate of damages to our environment, economy and health caused by invasive species will accelerate. These damages are spread across many stakeholders, and no strong, nationwide group has emerged to encourage industries that are pathways of introduction to reduce the threat. Hence the federal government must assume greater leadership to coordinate efforts by all

levels of government. We recognize that the problem is complex and interdisciplinary, includes many pathways, a tremendous diversity of organisms that are invasive, and the vulnerability of all terrestrial, marine and freshwater ecosystems. Despite this complexity, and the consequent overlapping and sometimes conflicting state, federal and international policies involved, the six recommendations described in this paper provide sound guidance for the future. Recent scientific and interdisciplinary advances provide a strong basis for rapid implementation of these cost-effective solutions.

APPENDIX G: LIST OF REGULATED SPECIES IN CALIFORNIA

Aquatic invasive species are regulated by a number of state and federal regulations. The aquatic plant and animal species restricted in California, and the regulations that apply to each, are listed below.

ANIMALS

In California, the animal species considered detrimental to native wildlife, state agriculture or public health and safety are listed in California Administrative Code Title 14, Section 671. Importation, transportation and possession of the restricted animals on this list are unlawful except under permit issued by the California Department of Fish and Game. Animal species restricted by the federal government are considered “injurious wildlife” and named in the Lacey Act (50 CFR 16.11-16.15). The U.S. Fish and Wildlife Service has responsibility for regulating the live importation or shipment of these animals.

California’s list of Restricted Animals

<http://www.dfg.ca.gov/licensing/pdffiles/fg1518.pdf>

Click on the following link: “Search for a Specific Regulatory Section”

Title: 14

Section: 671

Injurious Wildlife Species List (PDF)

U.S. Fish and Wildlife Service

<http://www.invasivespeciesinfo.gov/laws/main.shtml>

PLANTS

Certain aquatic invasive plants are listed as Noxious Weed Species in Title 3, Section 4500 of the California Administrative Code. Their eradication, control, and containment are regulated by the California Department of Food and Agriculture (DFA). Each species has been given a “pest rating” based on the economic risks it poses to the state. In addition, Division 3, Chapter 3.5, Section 2300 of the California Fish and Game Code restricts all species of the marine alga genus *Caulerpa*. Federally restricted invasive plants are listed in Noxious Weed Act P.L. 93-629.

CDFA Weed List

http://www.cdfa.ca.gov/phpps/ipc/encycloweedia/pdfs/noxiousweed_ratings.pdf

Federal Noxious Weed List (PDF)

<http://www.aphis.usda.gov/ppq/weeds/weedlist2006.pdf>

Appendix G

State and/or Federal Regulated Aquatic Invasive Animals

Scientific Name	Common Name	Group	Habitat	Regulated By
Mustelidae (Family)	All species except <i>Amblyonyx cinerea</i> , Oriental small-clawed otter, <i>Aonyx capensis</i> , African clawless otter, <i>Pteronura brasiliensis</i> , giant otter and all species of genus <i>Lutra</i> , river otters.	Mammals	F	CA
Amiidae (Family)	bowfins	Fish	F	CA
Anguilla (Genus)	freshwater eels	Fish	F	CA
Aplodinotus grunniens (Species)	freshwater drum	Fish	F	CA
Astyanax fasciatus (Species)	banded tetra	Fish	F/B	CA
Belonesox belizanus (Species)	pike killifish	Fish	F	CA
Carcharhinus (Genus)	freshwater sharks	Fish	F	CA
Cetopsidae (Family)	whalelike catfishes	Fish	F	CA
Channidae (Family)	snakeheads	Fish	F	CA, US
Clariidae (Family)	labyrinth catfishes	Fish	F	CA*,US
Ctenopharyngodon idella (Species)	grass carp (permits may be issued for possession of triploid grass carp)	Fish	F	CA
Cyprinodon variegatus (Species)	sheepshead minnow	Fish	F/B	CA
Dorosoma cepedianum (Species)	gizzard shad	Fish	F	CA
Esocidae (Family)	piques	Fish	F	CA
Heteropneustidae (Family)	airsac catfishes	Fish	F	CA
Hoplias malabaricus (Species)	tiger fish	Fish	F/B	CA
Hypophthalmichthys molitrix (Species)	silver carp	Fish	F	CA
Hypophthalmichthys nobilis (Species)	bighead carp	Fish	F	CA
Ictiobus (Genus)	buffalo suckers	Fish	F/M	CA
Lepisosteidae (Family)	gars	Fish	F	CA
Leuciscus idus (Species)	Ide	Fish	F	CA
Morone americana (Species)	white perch	Fish	F	CA
Morone chrysops (Species)	white bass	Fish	F	CA
Perca flavescens (Species)	yellow perch	Fish	F	CA
Potamotrygonidae (Family)	river stingrays	Fish	F/M	CA
Petromyzontidae (Family)	lampreys - all nonnative species	Fish	F/M	CA
Salmo salar (Species)	Atlantic salmon - restricted in the Smith River watershed	Fish	F/M	CA

* Only members of the Clarias, Dinotopterus, and Heterobranchus genera are prohibited by Title 14 section 671

Key

B	Brackish	CA	CDFG Restricted Species, Title 14, Section 671
F	Freshwater	US	USFW Lacey Act 50 CFR 16.11-16.15
M	Marine		

Appendix G

State and/or Federal Regulated Aquatic Invasive Animals

Scientific Name	Common Name	Group	Habitat	Regulated By
Salmonidae (Family)	live or dead uneviscerated salmonid fish, live fertilized eggs, or gametes of salmonids are prohibited unless accompanied by a certification that ensures they are free of <i>Onchocorhynchus masou</i> virus and the viruses causing viral hemorrhagic septicemia and infectious hematopoietic necrosis, and meet the conditions in 50 CFR 16.13	Fish	F/M	US
Serrasalmus (Genus)	piranhas (including genera <i>Pygocentrus</i> and <i>Pygopristis</i> , and invalid genera <i>Serrasalmo</i> , <i>Taddyella</i> , <i>Rooseveltiella</i>)	Fish	F	CA
<i>Stizostedion vitreum</i> (Species)	walleye	Fish	F	CA
<i>Tilapia aurea</i> (Species)	blue tilapia	Fish	F/M/B	CA
<i>Tilapia nilotica</i> (Species)	Nile tilapia	Fish	F/M/B	CA
<i>Tilapia sparrmani</i> (Species)	banded tilapia	Fish	F/M/B	CA
<i>Tilapia zillii</i> (Species)	redbelly tilapia (permits may be issued to a person or agency for importation, transportation, or possession in the counties of San Bernardino, Los Angeles, Orange, Riverside, San Diego, and Imperial)	Fish	F/M/B	CA
Trichomycteridae (Family)	parasitic catfishes	Fish	F	CA
<i>Ambystoma</i> (Genus)	tiger salamanders	Amphibian	F	CA
Bufo (Family)	toads (including <i>Bufo marinus</i> , cane toad, giant toad or marine toad; and invalid species, <i>Bufo paracnemis</i> , Cururu toad, and <i>Bufo horribilis</i> , other large toads from Mexico and Central and South America)	Amphibian	F/M	CA
<i>Xenopus</i> (Genus)	clawed frog	Amphibian	F	CA
Crocodylia (Order)	crocodiles, caimans, alligators and gavials	Reptile	F/M	CA
Chelydridae (Family)	snapping turtles	Reptile	F	CA
Cambaridae (Family)	crayfish - all species except <i>Procambarus clarkii</i> and <i>Orconectes virilis</i>	Invertebrate	F/M	CA
<i>Eriocheir</i> (Genus)	crabs	Invertebrate	F/M	CA, US
<i>Dreissena</i> (Genus)	zebra and quagga mussels	Invertebrate	F	CA, US **
<i>Potamopyrgus antipodarum</i> (Species)	New Zealand mudsnail	Invertebrate	M	CA
Transgenic Aquatic Animals	Freshwater and marine fishes, invertebrates, crustaceans, mollusks, amphibians and reptiles		F/M	CA

** Only the species *Dreissena polymorpha* is prohibited by the Lacey Act

Key

B	Brackish	CA	CDFG Restricted Species, Title 14, Section 671
F	Freshwater	US	USFW Lacey Act 50 CFR 16.11-16.15
M	Marine		

Appendix G

State and/or Federal Regulated Aquatic Invasive Animals

Scientific Name	Common Name	Habitat	Applicable Regulations/Pest Rating
<i>Alternanthera philoxeroides</i>	alligatorweed	F	A
<i>Arundo donax</i>	giant reed	W/U/R	B
<i>Azolla pinnata</i>	mosquito fern, water velvet	F	US
<i>Cabomba caroliniana</i>	fanwort	F	Q
<i>Caulerpa taxifolia</i>	Caulerpa	M	US, DFG
<i>Caulerpa cupressoides</i>	Caulerpa	M	DFG
<i>Caulerpa mexicana</i>	Caulerpa	M	DFG
<i>Caulerpa sertularioides</i>	Caulerpa	M	DFG
<i>Caulerpa floridana</i>	Caulerpa	M	DFG
<i>Caulerpa ashmeadii</i>	Caulerpa	M	DFG
<i>Caulerpa racemosa</i>	Caulerpa	M	DFG
<i>Caulerpa verticillata</i>	Caulerpa	M	DFG
<i>Caulerpa scapelliformis</i>	Caulerpa	M	DFG
<i>Eichhornia azurea</i>	anchored water hyacinth	F	US
<i>Hydrilla verticillata</i>	hydrilla	F	US, A
<i>Hygrophila polysperma</i>	Miramar weed	F	US
<i>Ipomoea aquatica</i>	Chinese water spinach	F	US
<i>Lagarosiphon major</i>	oxygen weed	F	US
<i>Limnobium spongia</i>	spongeplant	F	Q
<i>Limnophila indica</i>	ambulia	F	Q
<i>Limnophila sessiliflora</i>	ambulia	F	US, Q
<i>Lythrum salicaria</i>	purple loosestrife	W/U	B
<i>Melaleuca quinquenervia</i>	broadleaf paper-bark tree	W	US
<i>Monochoria hastata</i>	monochoria	F	US
<i>Monochoria vaginalis</i>	heartshape false pickerelweed	F	US
<i>Nymphaea mexicana</i>	banana water lily	F	B
<i>Ottelia alismoides</i>	duck lettuce	F	US
<i>Pistia stratiotes</i>	water lettuce	F	B
<i>Polygonum amphibium</i>	swamp smartweed	F	C
<i>Polygonum cuspidatum</i>	Japanese knotweed	W/U/R	B
<i>Sagittaria sagittifolia</i>	arrowhead	F	US
<i>Salvinia auriculata</i>	salvinia	F	US, A
<i>Salvinia biloba</i>	salvinia	F	US, A*
<i>Salvinia herzogii</i>	herzog salvinia	F	US, A*
<i>Salvinia molesta</i>	giant salvinia	F	US, A*
<i>Sparganium erectum</i>	exotic bur-reed	F	US
<i>Tamarix chinensis</i>	Chinese tamarisk	U/R	B
<i>Tamarix gallica</i>	French tamarisk	U/R	B
<i>Tamarix parviflora</i>	smallflower tamarisk	U/R	B
<i>Tamarix ramosissima</i>	salt cedar	U/R	B

*DFA considers these species a synonym of *Salvinia auriculata*

Appendix G
State and/or Federal Regulated Aquatic Invasive Animals
Key for State and/or Federally Regulated Aquatic Invasive Plants

DFG Regulated by CDFG Division 3, Chapter 3.5, Section 2300

F Freshwater

M Marine

R Riparian

SM Saltmarsh

U Upland

US **Regulated by the Federal Noxious Weed Act, P.L. 93-629.**

For more details, see the discussion of the Noxious Weed Act in the subsection titled "Other Federal Authorities" in Appendix B of the California Aquatic Invasive Species Management Plan..

W Wetland

Noxious Weed Ratings per California Department of Food and Agriculture Plant Industry Policy Letter 89-2, May 1, 1989. <http://www.cdffa.ca.gov/cdfa/pendingregs/docs/PlantPestRatings.pdf>

A An organism of known economic importance subject to enforced action involving eradication, containment, rejection, or other holding action at the state-county level. Quarantine interceptions to be rejected or treated at any point in the state.

B An organism of known economic importance subject to eradication, containment, control or other holding action at the discretion of the commissioner. OR an organism of known economic importance subject to state holding action and eradication only when found in a nursery.

C An organism subject to state endorsed holding action and eradication only when found in a nursery; action to retard spread outside of nurseries at the discretion of the commissioner; reject only when found in a cropseed for planting or at the discretion of the commissioner.

Q An organism requiring a temporary "A" action pending determination of a permanent rating. It is suspected to be of economic importance, but its status is uncertain because of incomplete identification or inadequate information.

D Organisms determined to be of little or no economic importance

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Inhibition of Giant Kelp Recruitment by an Introduced Brown Alga

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Abstract

Following a natural disappearance of the native giant kelp *Macrocystis pyrifera*, an invasion of the introduced brown alga *Sargassum muticum* appeared to prevent giant kelp recruitment. Experimental removal of adult *S. muticum* resulted in a significant increase in giant kelp recruitment compared to nonremoval areas. Two requirements for the successful establishment of giant kelp appeared to be: 1) an absence of *S. muticum*, and 2) the presence of nearby giant kelp adults to supply sufficient spores. Higher densities of giant kelp in removal areas persisted throughout the period when *S. muticum* dies back to a perennial holdfast. Shading at a critical time in the giant kelp life cycle is suggested as a possible mechanism for the inhibition of giant kelp recruitment.

Introduction

The large kelps (Phaeophyta, Laminariales) comprise the majority of the subtidal canopy algal species in temperate marine communities (Mann 1973). Their ecological importance has long been recognized (Darwin 1860, North 1971). Dayton (1975) and Hruby (1976) have found that recruitment, growth, and development of several laminarians can be inhibited by competition with native algal species. In this paper, we examine the interactions between the introduced brown alga *Sargassum muticum* (Yendo) Fensholt (Order Fucales) and the native giant kelp *Macrocystis pyrifera* (L.) C. A. Agardh.

Sargassum muticum was introduced to the west coast of North America from Japan in the late 1940's (Scagel 1956). Initially restricted to the Pacific Northwest, it spread southward to Baja California by 1971 (Setzer and Link 1971) and was found for the first time in England in 1973 (Farnham, Fletcher and Irvine 1973). In spite of the rapid invasion of *S. muticum* and speculation of its potential influence on algal communities (Druehl 1973, Jones and Farnham 1973, Farnham and Jones 1974, Fletcher and Fletcher 1975, Norton 1976), its actual impact on native algae has not been studied. Our evidence suggests that *S. muticum* inhibits recruitment of the giant kelp *Macrocystis pyrifera* on Santa Catalina Island, California, USA.

Prior to 1976, a very dense bed of *M. pyrifera* existed near the east end of Bird Rock, a small island near Santa Catalina Island. This kelp bed, along with many others at Santa Catalina Island, disappeared in the summer of 1976 (see Coyer 1979). Elevated temperatures from

mid-June to November 1976 at Santa Catalina (Coyer 1979) were the probable cause of the decline, since kelp beds in southern California deteriorate when water temperatures exceed 20 °C (North 1971, Abbott and North 1972). During the following winter, *S. muticum* appeared in high densities at Bird Rock for the first time, ultimately covering the entire east end. Even though *S. muticum* dies back to a perennial holdfast in early summer and does not resume vegetative growth until fall (Ambrose and Nelson, pers. obs.), there was essentially no *M. pyrifera* recruitment at Bird Rock in 1977 or 1978. Kelp abundance in other Santa Catalina kelp beds, where *S. muticum* had not become established, increased to predecline levels by August 1977 (Coyer 1979). These observations suggested that *S. muticum* might have been preventing the re-establishment of *M. pyrifera* at Bird Rock. We cleared several areas of *S. muticum* in 1979 to examine the possibility that *S. muticum* inhibited *M. pyrifera* recruitment.

Methods

Research was conducted near the east end of Bird Rock, a small island 0.5 km N of Big Fisherman's Cove on Santa Catalina Island (33° 27' N, 118° 29' W), 30 km S of Los Angeles, California, USA. The study site depth varies from 3 to 7 m below MLLW. Water temperature varies from 12°–14 °C in the winter to 20°–22 °C in the summer.

Sargassum muticum was removed by pulling up individual plants by hand along a corridor 2 m wide; no at-

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tempt was made to remove the perennial holdfast. Two areas were studied. Area 1 removal (15 m long) and its control (13 m long) were 2 m and 4 m away from a sharp dropoff to approximately 30 m ("the wall"), respectively. *Sargassum muticum* does not extend onto the wall, where many *Macrocystis pyrifera* persist. In Area 2 there were two removal areas (a and b) and one control, each 10 m long and located 13 m away from the wall. Removal areas were cleared monthly from January to June 1979, although not all areas were cleared every time; generally, an area was cleared as soon as *S. muticum* began to grow back noticeably. Nonmanipulated areas maintained high *S. muticum* cover until early June when *S. muticum* dies back to a perennial holdfast. *Sargassum muticum* densities were estimated by censusing 20 haphazardly placed 0.25 m² quadrats. Removal and control areas were censused by counting all juvenile *Macrocystis pyrifera* in contiguous 1 m² quadrats. The study site was censused in late June, approximately one month after the natural die-back of *S. muticum*, and in late September, one month before *S. muticum* resumed vegetative growth, to distinguish between the following hypotheses: 1) *Macrocystis pyrifera* recruitment is delayed until *S. muticum* dies back, and 2) *S. muticum* prevents the recruitment of *M. pyrifera*. Differences between areas were examined using the Wilcoxon two-sample test.

Results

In 1979 the *Sargassum muticum* density near the east end of Bird Rock averaged 56.9 plants/m² (range 8/m² to 136/m²); plant height was 3–4 m. In June, following the *S. muticum* die-back, juvenile *M. pyrifera* density was significantly higher in removal areas 1 and 2a than their respective controls ($p < 0.001$; Tab. I). Density was not significantly greater in removal 2b than its control. In general, areas near the wall had higher recruitment than their counterparts away from the wall. Recruitment in removal area 1 was significantly higher than removal areas 2a and 2b ($p < 0.01$). The greater juvenile *M. pyrifera* density in control area 1 versus control area 2 was not significant.

Supplementary observations in June of an isolated adult *M. pyrifera* plant 10 m away from the wall, where *S. muticum* was not removed, failed to reveal any juvenile *M. pyrifera* around the adult.

Juvenile *M. pyrifera* density in September 1979, although lower than in June, was still significantly higher in removal areas 1 and 2a than their respective controls ($p < 0.001$ and $p < 0.01$, respectively). Density was marginally significantly greater in removal area 2b than its control ($0.05 < p < 0.10$). Recruitment in removal area 1, near the wall, was significantly higher than removal areas 2a and 2b ($p < 0.01$ and $p < 0.001$, respec-

tively). The difference in juvenile *M. pyrifera* density between control area 1 and control area 2 was marginally significant ($0.05 < p < 0.10$).

Tab. I. Juvenile *Macrocystis pyrifera* density in *Sargassum muticum* removal quadrats and in unmanipulated controls. Censuses were conducted in June 1979, one month after the natural die-back of *S. muticum*, and in September 1979, one month before *S. muticum* resumed growth. N = 30 for area 1 removal, N = 26 for area 1 control, and N = 20 for area 2 removals and control. ** and * indicate removal significantly different from control at $p < 0.001$ and $p < 0.01$, respectively. + indicates removal marginally significantly different from control at $0.05 < p < 0.10$.

Area	Dis- tance from wall (m)	<i>Macrocystis</i> Density Mean (s. d., range) (Plants/m ²)		
		<i>Sargassum</i> Removal		Control
June 1979				
1	2	5.0 (5.91, 0–32)	**	0.4 (0.63, 0–3)
2a	13	1.9 (1.38, 0–12)	**	0.2 (0.37, 0–1)
2b	13	0.4 (0.75, 0–3)		
September 1979				
1	2	1.9 (2.06, 0–7)	**	0.1 (0.31, 0–1)
2a	13	0.4 (0.60, 0–2)	*	0.0 (–)
2b	13	0.1 (0.31, 0–1)	+	

Discussion

The *Macrocystis pyrifera* juveniles growing in the *S. muticum* removal areas represent the first significant *M. pyrifera* recruitment near the east end of Bird Rock since the establishment of *S. muticum* 3 years earlier. However, *S. muticum* removal did not invariably result in greater numbers of juvenile *M. pyrifera*. One treatment, removal area 2b, was statistically indistinguishable from its nonremoval control. Low recruitment in area 2b may result from the limited distance of *M. pyrifera* dispersal (Anderson and North 1966). Only one adult *M. pyrifera* was within a few meters of removal area 2b. Limited dispersal may also explain the differences between areas 1 and 2, which were 2 m and 13 m away from the wall, respectively. Tidal currents could more easily disperse spores from the numerous adult *M. pyrifera* plants growing on the wall to areas near the wall (area 1) than to area 2, where adult *M. pyrifera* were sparse.

Of the many possible mechanisms by which *S. muticum* may reduce *M. pyrifera* recruitment, shading is perhaps the most likely. The dense *S. muticum* bed (mean of 56.9 plants/m²) very effectively lowered light levels. Since dense kelp canopies can prevent the development of juvenile *M. pyrifera* (Anderson and North 1969), recruitment under the *S. muticum* canopy seems unlikely. Shading by *S. muticum* may occur at a critical

time in the *M. pyrifera* life cycle. In southern California the peak *M. pyrifera* recruitment occurs in early spring (Anderson and North 1969), after which recruitment is normally very low. The absence of *M. pyrifera* juveniles in the control areas approximately 4 months after the *S. muticum* die-back indicates that *S. muticum* did not simply cause the delay of *M. pyrifera* recruitment, but effectively prevented recruitment.

Sargassum muticum is an extremely successful introduced species, having spread through England and the west coast of the United States in a remarkably short time. Our observations suggest that there is intense competition between *M. pyrifera* and *S. muticum*. Both species utilize the same basic resource and both form canopies. Although *S. muticum* was present at Santa Catalina Island since 1971 (Setzer and Link 1971), it did not invade the east end of Bird Rock until the massive *M. pyrifera* die-off in 1976, suggesting that the

local distribution of *S. muticum* was restricted by competition with *M. pyrifera*. Once *S. muticum* became established at Bird Rock, *M. pyrifera* did not reinvade. Kelp forests are regularly exposed to natural as well as man-made disturbances (North and Pearse 1970, North 1971, 1976, Rosenthal *et al.* 1974). If our observations are generally true, it is possible that by exploiting these opportunities, *S. muticum* may have a substantial impact on *Macrocystis pyrifera* distribution in southern California.

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Outplanting large adult green abalone (*Haliotis fulgens*) as a strategy for population restoration

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Wild abalone populations are in decline around the globe. Given their high market value, abalone have been targeted for restoration in many areas where they were once abundant. Efforts to restore California green abalone (*Haliotis fulgens*) have had limited success for species recovery. This study aimed to use large (>14cm) adult green abalone as a strategy for restoration. Abalone of this size have few predators and are generally emergent, making them more visible during surveys. Sixty-nine large (average size 16.2 cm) farm raised abalone were outplanted in three batches (May, July and August) in Newport Beach, California, on natural reef structure at a depth of 8.4 m, monitored for 15 months, and then recaptured. Using multiple tagging devices and rigorous monitoring resulted in 40% survival at the end of the study, with 61% of the mortalities occurring within the first 30 days of outplanting, and 46% of the August outplants surviving to the end of the study period. Most of the trackable abalone movements, throughout the study, were confined to a 10 m radius of outplanting areas and 79% (22) of the surviving abalone stayed within 8 m of the outplant areas.

Key words: abalone, adult abalone, *Haliotis fulgens*, outplanting, restoration, restocking, size, stock enhancement

Abalone populations worldwide have been in decline for many decades (Campbell 2000). Over fishing, illegal harvest, disease and habitat degradation are thought to be the primary causes (Cook 2014). California once supported fisheries for five species of abalone (black, green, pink, red, white) and by 1998 all commercial and recreational fisheries were closed south of San Francisco bay. Rogers-Bennett et al. (2004) found that adult abalone densities in southern California were two orders of magnitude below the estimated minimal viable population of 2000 individuals/ha and at that point, abalone recruitment in southern California had declined 20-fold over the previous decade. Despite 20 years of closed fisheries, populations of all five of these abalone species have yet to rebound on coastal reefs in southern California indicating a need for restoration activities. McCormick et al. (1994) suggested that seeding areas with hatchery raised abalone may be the only means of increasing coastal abalone stocks on a time scale meaningful to fishery managers

The challenges facing abalone restoration include: captive spawning and rearing, protecting aggregated or outplanted animals from poaching, tracking reproduction, quantifying survival, and maximizing survival of captive-reared abalone in the wild (Henderson et al. 1988, Tegner and Butler 1989, Tegner 1992, Rogers-Bennett and Pearse 1998, Tegner 2000). Reseeding or outplanting projects have most often involved larvae and juveniles (0-100 mm) and have had mixed results around the globe with Japan and New Zealand reporting higher than 50% survival for some projects (Saito 1984, Schiel 1993, Kojima 1981). Results for reseeded or outplanting juveniles in southern California report much lower recovery rates ranging from 0-6% (Tegner and Butler 1985, McCormick et al. 1994, Davis 1995, Chick et al. 2013). Quantifying recovery rates is a challenge for comparisons of efforts across time, species and different geographic areas.

Green abalone, (*Haliotis fulgens; Philippi*), are native to southern California and range from Point Conception, California, USA, to Magdalena Bay, Baja California, Mexico, and include the offshore islands (Cox 1962). They were once part of a large recreational and commercial fishery, and have previously been a target for species recovery. The green abalone is listed as a federal Species of Concern (NOAA 2004) and based on historic landings, is estimated to be at less than 1% of its baseline density (Rogers-Bennett et al. 2002). The major threat to remaining populations is their low densities and the possibility of reduced reproduction resulting from the Allee effect (Allee 1931). Low densities of broadcast spawners can lead to poor fertilization and recruitment failure because of the distances between males and females (Babcock & Keesing 1999). Remnant populations are comprised primarily of solitary abalone, many of which may not be contributing to reproduction and are thus functionally sterile (Taniguchi et al. 2013). Results from a drift tube study by Tegner and Butler (1985) indicated that in the absence of local broodstock, a fishery closure alone would not be an effective management policy for the recovery of green abalone populations on the mainland in southern California.

There have been several attempts at restoration of green abalone beginning in the 1970s. Most attempts have involved outplanting small hatchery reared animals generally due to costs associated with raising this slow growing mollusk. Seeding or outplanting results are affected by many variables including condition of the abalone at release, size, planting method, season, as well as site specific conditions including habitat type, food availability, predation, and topography (Saito 1984, Schiel 1993, McCormick et al. 1994). Because of the cryptic and mobile nature of small abalone it is difficult to estimate survival in most studies (Breen 1992, Shepherd & Breen 1992). Juveniles are highly cryptic and are found during daylight hours beneath rocks or in the recesses and crevices; they move freely at night and seldom return to the same location as the preceding day (Leighton 2000). Outplanting activities in Baja California with approximately 20 mm (shell length) green and pink abalone have yielded recovery rates ranging up to 4.7% (Sercy-Bernal et al. 2013). In summary, abalone outplanting has many variables to consider and there has been no formula for "success" that works for all species in all locations.

Translocation of abalone involves aggregating wild animals into one location with the aim of increasing reproductive success. A recent trial involving the translocation of adult California green (*H. fulgens*), and pink (*H. corrugata*), abalone showed that green abalone were not a good candidate for this restoration technique because they exhibited site infidelity (Taniguchi et al. 2013). A previous trial of 4,453 translocated green abalone on the Palos Verdes Peninsula, California was inconclusive due to poaching of the aggregated animals in the second year of the project (Tegner 1992).

Natural mortality of juvenile abalone may vary with location, time, and generally declines with age (Tegner and Butler 1985, Prince et al. 1988, Shepherd and Daume 1996). Initial mortality rates for outplanted juvenile abalone species are quite high and the rates decrease as the abalone grow to larger sizes (Schiel 1993). Saito (1984) found that survival of outplanted abalone increased with seed size in the range of 10 to 50 mm. Outplanting large adults in high densities on isolated reefs seems to be more effective (Coates et al. 2013).

Studies conducted in the 40 years before this project noted issues with the following: tagging (tags falling off, not identifiable); tracking (outplanted animals were not surveyed with enough frequency, were too cryptic, or emigrated off study site); predation (the size of the outplanted animals were vulnerable to multiple predators); poaching; and mortalities from transport shock. With historically limited success in green abalone restoration utilizing juveniles, the aim of this study was to use large (>14 cm) adult abalone for outplanting as a possible restoration strategy and to quantify their survival. This project aimed to also address some of the previous noted issues by using multiple tags, surveying with greater frequency, minimal handling in transport, and removing sea star predators. The use of large animals may act as a model for other abalone species including the endangered white (*H. sorenseni*) and black abalone (*H. cracherodii*) as recommended by Davis et al. (1998). The results are compared with previous restoration studies to determine if larger (>14 cm) outplants yield higher survival rates. The premise is that, large abalone have fewer predators and they are more easily detected and tracked.

MATERIALS AND METHODS

Study site.—The green abalone outplant site was located in Crystal Cove State Park, Orange County, California, with coordinates 33° 34' 6.528" N, 33° 34' 6.528" W. The study site was chosen because it was familiar to the author, too far from shore for shore divers to reach, and was not a well-known recreational dive spot minimizing opportunities for poachers. Surveys were conducted to characterize the composition of the reef, describe the topography, and assess the predator population. Predators of large abalone (>14 cm) in Orange County include octopus (*Octopus sp.*), sea stars (*Pisaster sp.*), and the bat ray (*Myliobatis californica*). The surveys were conducted using two different methods. In one method, an observer conducted two 30 x 2 m band transect surveys and the other method included 30 random 1-m² quadrats along two 30-meter transects. Each surveyor collected information on reef composition (continuous reef, boulder, sand, or cobble on every meter), changes in rugosity (change in height of the reef at every meter), percent cover (sessile invertebrates, algal species), the presence of wild abalone, and presence/absence of predators.

The 450 m² reef was roughly rectangular and was divided into eight quadrants (approximately 9 x 6 m) using plastic clothesline stretched out across the reef and tied off to cinderblocks. Each quadrant was labeled with floating numbers to make the process of mapping the locations of abalone easier for volunteers. The large *Pisaster* stars were removed before outplanting and continuously removed during the project period. No octopus were removed from the reef but were present during the entire study, and two bat rays were observed near the reef, one before and one during the study.

Tagging.—Seventy adult abalone were purchased (\$38 each) from The Cultured Abalone, a commercial farm in Goleta, California. The average size of the abalone was 16.2 cm (max 17.9 cm, min 14.6 cm). These animals were used as broodstock on the farm and thought to be at least 10 years old. They were shipped in three batches to a holding facility in

San Pedro, California in moist foam and oxygen filled bags and held for up to thirteen days to tag, monitor, and reduce stress from transport. Upon arrival, the animals were measured, sexed, affixed with tags using Splash Zone marine epoxy or cyanoacrylate (Super Glue), and photographed. Of the 69 abalone tagged, 87% (60) were identified as female (Table 1). Since abalone are known for choosing crevices, ledges, and overhangs for their home scars multiple tags were used to make the identifiers visible from any angle. The tags identified which outplant batch the animal was from and had both a unique number identifier (Major Tag) and several auxiliary tags (Minor Tag). Each animal was given a “Major” tag with a number, a color coded zip tie, and up to four other “Minor” tags (Figure 1). The Major tags consisted of a 1.5 cm stainless steel disk with etched numbers; a 2.5 cm white plastic square with printed black numbers; or a 4 cm brass disk with printed black numbers. All of the abalone had a colored zip tie secured through the first or second respiratory pore. PIT (passive integrated transponder) tags were epoxied on the shells of 32 of the animals for the purposes locating the animals using a PIT tag reader. Minor tags consisted of one or more of the following: blue aluminum tree tags with etched numbers; colored plastic bottle caps; white plastic beads with black letters; red plastic key tags with white numbers; stainless steel washers; plastic chain links; and metallic painted plastic jewelry (shiny). No two animals had the same combination of tags. The white lettered beads were the only tag affixed with cyanoacrylate. Knowing that the abalone would be cryptic to the observing volunteer divers, the objects used for tagging were meant to help spot the animals and the combinations of tags helped to identify the animals in hard to see places.

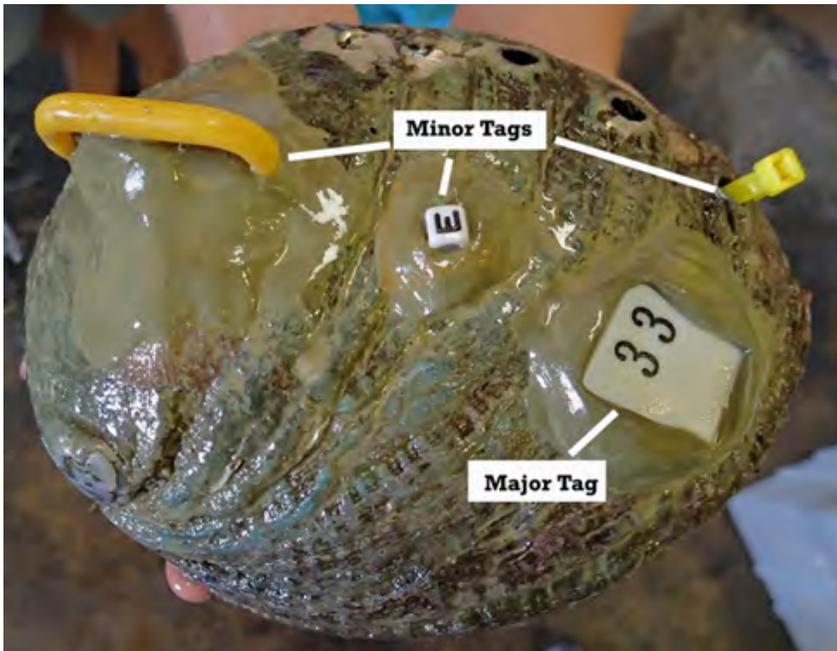


FIGURE 1.—Example of multiple tagging methods for green abalone outplants illustrating “Major” and “Minor” tags. Recorded as Major tag: #33, Minor tags “E”, yellow chain link, yellow zip tie, and PIT tag # (in the epoxy).

Following the tagging activity, the animals were placed in rectangular plastic milk crates and submerged in a recirculating seawater holding systems (18 °C) for up to 13 days. The top of the milk crate was covered with plastic mesh so the animals could not crawl out. There was one mortality while in the holding tanks presumably due to stress related to shipment.

Outplanting.—Sixty-nine green abalone were outplanted in three batches in May 2013, July 2013, and August 2013 (Table 1). The animals were monitored for survival for one year after the last outplanting (until August 2014). On the day of outplanting, the animals were checked for health and for any tag loss, the milk crates were put into large coolers with seawater from the holding tanks and transported to the outplant site by car and then by boat. They were in transport for approximately three hours. While on board the boat, fresh ocean water was exchanged with the water in the cooler by bucket. Divers descended to the reef with the milk crates. When on the bottom, the milk crates were turned on their side and four half-sized cinder blocks were zip-tied to each milk crate to weigh them down. The first and third outplant sites offered more ledges and overhangs while the second outplanting area was on the top of the reef just above the other two. All of the locations chosen to place the crates on were within 5 m of each other on the west end of the reef (Figure 2). In accordance with the outplanting permit, as many abalone as possible were recovered from the test site at the end of the study. All animals were measured at the beginning of the study and emergent animals were measured at the end of the study. Volunteers were asked to not share the outplanting location with anyone. Temperature loggers (Hobo) were deployed from 01 April 2013 to 25 March 2014.

Monitoring.—Monitoring began with the first outplanting in May 2013 and concluded one year after the last outplanting in August 2014, representing a 15-month study period. Rigorous monitoring was required to track the newly released animals as they were very mobile. In order to track this movement, the program utilized volunteers. In total, 28 volunteers were trained as abalone observers. Each dive was led by the Get Inspired project biologist and assisted by up to four other volunteer divers. During each dive, a diver was assigned a quadrant number within which to survey the reef for abalone. Every visible tag

TABLE 1.—Proportion, by sex, of green abalone that were outplanted in three batches and their survival in Crystal Cove State Park, Orange County, California. Average size 16.2 cm.

Outplanted			
	Batch 1 5/26/13	Batch 2 7/22/13	Batch 3 8/11/13
Females	17	21	22
Males	2	1	6
Total	19	22	28
Survival 8/11/14			
Females	7	7	8
Males	1	0	5
Total	8	7	13

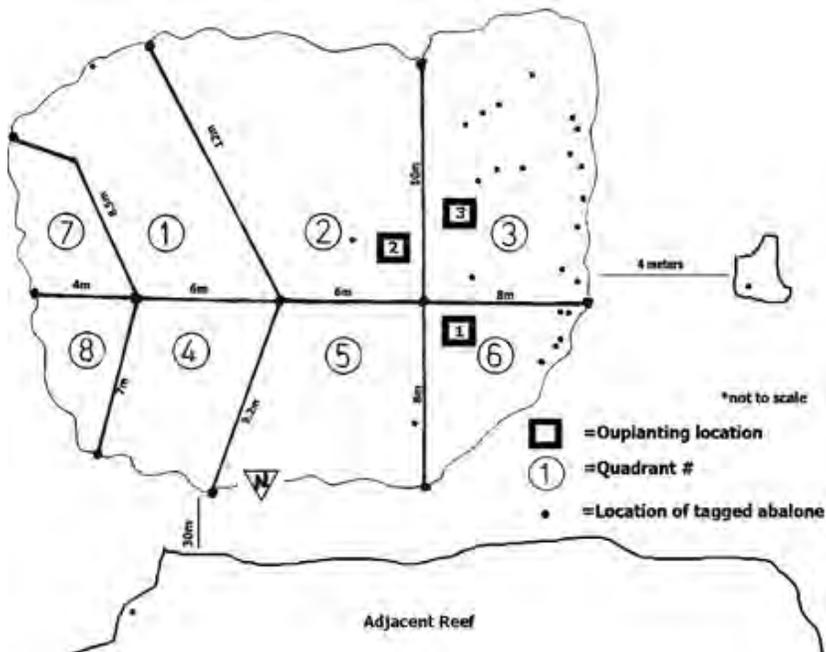


FIGURE 2.—Map of the relative locations of the surviving abalone created 11 August 2014.

on the animal was recorded and the shells and tags were cleaned with a toothbrush to reduce encrusting organisms. By recording every observable tag, even if a “Major” tag could not be seen, the combinations of other visible tags usually lead to the positive identification of a specific animal. If an abalone could not be positively identified, it was not counted that day. Empty shells and shell fragments were also collected for positive identification.

Over the 15-month (60 week) study period, 64 monitoring dives (approximately 45 min each) were made totaling 260 dive hours. Dives were conducted after each outplanting every 48-hours for approximately two weeks to track the immediate movements of the animals. Monitoring tapered off from every 48-hours to every four days, then once per week, then once every 10 days by the end of the study period. Telescoping mirrors and flashlights were used to look under ledges and in deeper crevices for abalone. A map of the location of each abalone was created/updated after each monitoring dive. An animated map was created, at the end of the project, to illustrate relative movements of the animals throughout the study period. Survival was calculated by finding and counting the actual live animals that were positively identified at the end of the study period.

RESULTS

Site Survey.—The study site is composed of continuous rocky reef approximately 450 m² in size and surrounded by sand. The reef is composed of bedrock and roughly rectangular with dimensions approximately 18 m wide by 25 m long, with the highest point being approximately 2 m from the sand that surrounds it. Changes in contour are minimal

on the top of the reef with rugosity being less than 1 m. The south and north ends of the reef are composed of ledges, the west end gently slopes down toward the sand, the east end of the reef is a wall that drops 2.5 m vertically to the sand. The reef was at a relatively uniform depth of 8.4 m on the top of the reef and it slopes on each side to a maximum depth of 11.5 m to the sand on the east end. Due to sea urchin removal activities during a giant kelp restoration project conducted on the reef by the author 10 years earlier, sea urchin densities were low with lots of crevice and ledge space available.

Both site survey methods provided similar results with mature giant kelp (*Macrocystis pyrifera*) covering 10% of the reef providing a 30% canopy, reaching the surface over the reef. Approximately 15% of the reef was covered with pink crustose coralline algae, and articulated coralline algae covered 10% of the reef. Subtidal algae (*Cystoseira osmundacea*) covered 5% of the reef surface and other low lying red and brown alga covered 15% of the reef. The remaining 45% of the reef was occupied by sessile invertebrates including tunicates, bryozoans, worms (*Serpulorbis* sp.), gorgonians, anemones, and sponges. There were no wild abalone observed on this reef before outplanting. The average temperatures on the reef during outplanting were as follows: May-18 °C, July-17.5 °C, August-15.8 °C

Tagging.—With continuous cleaning, the multiple tagging strategy worked well for the study period. Although the abalone routinely were wedged up and under rocks and ledges, the multi tag method allowed for identification of the animals from any angle. Only four of the major tags were lost due to poor epoxy application but the animals could still be identified by their minor tags. By the end of the project period, the brass tags (Major Tag) had tarnished making the numbers unreadable although we could still tell they were brass and coupled with the minor tags, each individual could still be identified. None of the zip ties or cyanoacrylate affixed tags were lost during the project period

Monitoring and movements.—The milk crates allowed for the abalone to attach to something that could easily be moved, placed in a cooler, and transported to the study site with minimal stress to the animal. Upon release, most of the animals immediately moved out of the crates and even within the period of the dive (approximately 45 minutes) they moved up to 2 m away. All of the abalone left the milk crates within 48-hours of outplanting. Some made their immediate homes inside the cinderblocks that weighted down the milk crates so after the first outplanting batch we covered the cinderblocks so the abalone would be forced out onto the reef. All the abalone were released on the west end of the reef and subsequently 96% of the animals stayed on the west side of the reef within a 10 m radius of their release site, either under ledges or oriented at the sand reef interface during the project period. The farthest distance moved by an abalone was 44 m and the shortest distance moved was <1 m, both of which survived until the end of the project (Figure 2).

The PIT tag reader was only used once and was not effective at locating abalone during that one use. An animated map was created from each survey by compiling location information allowing us to see the relative movements of the animals over the course of the study. This animated map is available from the author.

Survival.—Mortality was closely associated with outplanting events with 61% of mortalities (17) occurring within the first 30 days of being outplanted and 9% (6) mortalities occurring in the first week of outplanting. Being out and on top of the reef (emergent) was not the key factor in mortality because several animals survived through the entire project while in conspicuous places on top of the reef. No direct predation was observed, although we did remove a giant sea star (*Pisaster giganteus*) from the shell of a live abalone. Thirteen mortalities were observed with crushed shells (Figure 3) and the meat gone, with the shell



FIGURE 3.—Example of crushed shells which resulted in 13 mortalities, predator unknown.

fragments found in the same location that the live animal had been previously observed. The shell crushing predator was never observed.

During this 15-month study period, 28 animals (40%) survived (Table 1). We searched adjacent reefs and boulders off the study site. Two abalone were found on a boulder 4 meters away from the outplant reef. They migrated there independently over a two month period. Another abalone ventured across 10 m of sand, across 20 m of reef, then across another 4 m of sand to another adjacent reef. There were 13 animals or 19% of the original 69 that were missing and not accounted for at the end of the project. Some of these animals presumably could have survived. Of the 13 missing animals, seven went missing within 30 days of outplanting and were never seen again. Three of those abalone were missing from the first week of outplanting

After observing the habitat preferences of the first two batches of outplanted abalone, we chose the third outplanting site to match that of the first. It was 5 m away from the first on the edge of the west end of the reef with many overhangs and ledges. The last batch of abalone (28), outplanted in August, had 46% survival (Table 1). At the end of the 15-month period, eight abalone were retrieved in accordance with CDFW permits. The other 20 were not retrievable due to their positioning on the reef. The average growth of those eight surviving and retrieved abalone was 2.2 mm over the study period. Two of the 13 missing abalone were found dead two months after the end of the study period.

DISCUSSION

Based on findings from Tanaguchi et al. (2013), that green abalone expressed site infidelity when translocated; this survey site was specifically chosen because it was surrounded by sand. It was a disproven assumption that sand would act as a barrier and deter abalone movements. This finding presents a problem for future studies and may shed some light on previous studies where recapture rates were low. Green abalone will leave study sites even if it means crossing expanses of sand. It is possible more abalone emigrated from the survey site and these represent a proportion of the missing animals. Abalone movements and migrations are still poorly understood and continues to be a problem for abalone

outplanting/reseeding efforts. Current telemetry will add new knowledge to this question.

Juvenile abalone of all species may move tens of meters, but this tendency decreases with age (Cox 1962, Tutschulte 1976). Adult abalone generally have very limited movements (Shepherd 1973, Tutschulte 1976). Abalone have been known to move considerable distances which has made previous restocking projects challenging and often ineffective (Shepherd 1986, Ault & DeMartini 1987, Tegner & Butler 1989). The majority of the abalone that survived until the end of this project appeared to move very little during the project period, though this also made them easier for divers to find repeatedly. After each survey, a map of the relative locations of the abalone was created. From this, we noted that 22 (79%) of the surviving abalone were within an 8 m radius of the release sights at the end of the project (Figure 2). Many did not appear to move at all from these scars during the entire study. This may be an advantage of using large adult green abalone. In a telemetry study, Coates et al. (2013) mentions a “flight” response when pink abalone were translocated, this was thought to occur within the first 20 days after moving the animals. The reported 61% of the abalone mortalities from this study, occurred in the first 30 days and may have been due to this “flight” response in the initial phase after outplanting.

The fact that the abalone used for this study were farm raised has not been shown to be a factor in their ability to hide (Tegner and Butler 1985, Schiel and Weldon 1987). It appears that abalone have home scars and possibly home ranges for localized movements (Ault & DeMartini 1987, Tutschulte and Connel 1988). Some of the abalone in this study found their home scars right away while others seemed to “roam” throughout the study period. The challenge is to determine how long it takes for introduced/outplanted large emergent adult abalone to get acclimated to their outplanted reef so they “settle” in fast and find a home scar. Ideally, it would be most advantageous to be able to place abalone directly onto their preferred home scar location in hopes that they would stay there when outplanted.

There were at least 13 known abalone mortalities which involved crushed shells and there were many more shell fragments found that could not be identified. Given that these abalone were large with a shell thickness of at least 3 mm, the list of possible predators was small. Very large bat rays and humans are capable of such crushing forces. Giant seabass are capable of both “sucking” them off the reef and inflicting the force necessary to crush the shells (L. Allen, California State University Northridge, personal communication). Often the crushed shell would be found with all the pieces in the same spot that the live abalone was seen just 48 hours before. In October 2013, suspecting poaching as the possible cause of the crushing mortalities, floating signs were posted around the reef warning humans that they were under surveillance and that they were violating the law by taking or killing the animals. It should be noted that within 30 days of the signs being put up, the crushing mortalities stopped. This could be coincidence. It should be noted, that in January 2014 a mortality event (sea star wasting disease), which affected the west coast of North America, resulted in a die-off of all sea star species observed on the reef (Hewson et al. 2014). Sea stars, therefore, were not a predator of concern during much of this study.

Difficulties involved in quantifying the results of outplanting and reseeding efforts make it difficult to make comparisons between studies (McCormick et al. 1994). A summary of abalone outplanting projects around the world, their duration, and percent survival was compiled by Chick et al. (2013). In comparison with those studies, this study has notable survival rates for the project duration (>1 year) and species outplanted, and also used the largest size abalone. Of the studies conducted with larger red and green abalone (40-100

mm) in southern California, survival rates were only as high as 2.8% and the researchers claimed they found no evidence of size differential in survival (Tegner and Butler 1985, Tegner and Butler 1989, Davis 1995). Although survival may be quantified using several different methods, it is important to note that the survival rates reported for this study are actual, not estimates. Each animal counted as a survivor was physically observed.

The frequency with which the animals in this study were surveyed was an advantage for monitoring their survival and it may have been the key to the high recapture rates. We were able to observe their movements regularly (at most every 10 days). With the success of tracking and survival of the animals in this study, it is evident that the strategy of using larger animals for restocking green abalone is worthy of further study. The survival rate for this project is notable and far exceeds survival rates in other studies with green abalone. The animals used in this study were estimated to be at least 10 years old (ranging in size from 14.6 cm to 17.9 cm) by the farmer from whom they were purchased. The costs associated with raising them to this size may be great but there have been decades of attempts to restock. One expensive project may be worth 30 or more failed larval or juvenile outplanting attempts. Perhaps, outplants could be clustered to create reproductive "colonies". The animals used in this study seem to be the largest used in a California abalone restocking/outplanting study. We are currently spawning wild abalone to repeat this test in a future study in several different locations and may include animals 10 cm to 14 cm.

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Research Article

Distribution patterns of the non-native seaweeds *Sargassum horneri* (Turner) C. Agardh and *Undaria pinnatifida* (Harvey) Suringar on the San Diego and Pacific coast of North America

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Abstract

Here we report the occurrence of the two non-native brown macroalgal species *Sargassum horneri* (Turner) C. Agardh and *Undaria pinnatifida* (Harvey) Suringar in San Diego County and describe expansions in their ranges and new invasions on the California and Baja California coasts. Both species have exhibited characteristics of successful invaders: establishing in new areas, spreading locally, and persisting through multiple generations in areas that have been invaded. These species now occur primarily in harbors, but have also invaded open coast sites, suggesting that they can invade areas with relatively high wave action and with well-established native benthic communities. The rapid and uncontrolled spread of these species to date has serious implications for their expansion along the west coast of North America. The ecological and economic consequences of these invasions require further research.

Key words: invasive algae, macroalgae, California, Channel Islands, Cabrillo National Monument

Introduction

Marine algal invasions have become a pervasive problem. Diverse impacts include reductions in biodiversity and the abundance and performance of native species as well as changes in community structure and function (Walker and Kendrick 1998; Thresher 2000; Inderjit et al. 2006; Schaffelke et al. 2006; Valentine et al. 2007). Marine algal invasions can also threaten economically important species and industries such as aquaculture and tourism (Schaffelke et al. 2006). However, relatively few studies have comprehensively analyzed these invasions or addressed their effects (Nyberg and Wallentinus 2005; Inderjit et al. 2006; Schaffelke et al. 2006; Johnson and Chapman 2007; Schaffelke and Hewitt 2007; Valentine et al. 2007; Smith 2011). As a result, many gaps exist in our current knowledge of how specific non-native seaweeds affect indigenous ecosystems and the economies that depend on them.

Despite the fact that hundreds of species of non-native seaweeds have been documented around the world, research to date has largely focused

on a small fraction of these species and a limited number of invasion locations, or has simply documented occurrence without analyzing patterns of distribution or change over time (Inderjit et al. 2006; Johnson and Chapman 2007; Williams and Smith 2007). Consequently, very little is known about the natural history of non-native algal species in their invaded environments and their interactions with recipient environments, both important elements known to influence invasion success (Valentine et al. 2007). Studies that document species-and-region-specific patterns of establishment, spread, and persistence are a crucial first step in closing major gaps in our knowledge of the invasion process. Further, because invasions often proceed rapidly it is important to gain a better understanding of how new invaders spread in the early stages of establishment.

Southern California and the surrounding coastline have received multiple high-profile invasive algal species, but little information is available about the invasion dynamics of these taxa. A recent review by Miller et al. (2011) reports 27 non-native seaweed

species in California and 11 in Baja California, 9 of which are common to both areas. Most of these have been discovered in the last 30 years, and while the rate of introductions may not necessarily be increasing, climate change may increase the establishment of non-native species in Southern California and Baja California (Carlton 2000; Harley et al. 2006; Miller et al. 2011). This area has been invaded by some of the most high profile algal invaders in the world. *Caulerpa taxifolia* (M.Vahl) C.Agardh was first detected in two locations in Southern California in 2000 but was contained and successfully eradicated by 2006 (Jousson et al. 2000; Anderson 2005; Smith 2011). Other successful invaders include *Undaria pinnatifida* (Harvey) Suringar, first noted in 2000 (Silva et al. 2002), the globally invasive alga *Sargassum muticum* (Yendo) Fensholt, which was first noted in the 1970's and which has since become naturalized in this area (Norton 1981; Miller et al. 2007), and *Sargassum horneri* (Turner) C.Agardh, first noted in 2003 (Miller et al. 2007). Despite the long invasion history of this area, the dynamics and ecology of the non-native seaweeds in this region remain relatively unexplored.

Undaria pinnatifida is an aggressive invader worldwide, having colonized Argentina, New Zealand, Australia, Atlantic Europe, and the Mediterranean Sea (Silva et al. 2002; Nyberg and Wallentinus 2005). Its alarming rate of spread and ability to occupy and alter a variety of native systems have made this species one of only two algae on the International Union for the Conservation of Nature (IUCN) list of 100 most invasive species on the planet (Lowe et al. 2000). *Undaria pinnatifida* exhibits opportunistic life history traits that contribute to its successful establishment in new areas: a short, annual life span (Schaffelke et al. 2005; Miller and Engle 2009), high growth rate and fecundity, (Schaffelke et al. 2005; Valentine et al. 2007), and both a small and large dispersal shadow (Forrest et al. 2000). Serious negative ecosystem effects of this species - including reductions in native seaweed diversity- have been documented in shallow coastal communities elsewhere (Casas et al. 2004; Farrell and Fletcher 2006; Schaffelke and Hewitt 2007; Williams and Smith 2007). Because of the lack of knowledge of *U. pinnatifida* on the Pacific coast of North America and the potential for significant impacts of its further spread, we document the current distribution of this species in this region in the early stages of invasion.

In the early 20th century, *Sargassum muticum* was introduced to North America from northeast Asia and quickly spread throughout the west coast,

reaching southern California in the early 1970's (Miller et al. 2007). This species is a highly successful invader worldwide and is considered to be naturalized in intertidal and subtidal communities throughout southern California (Harries et al. 2007; Miller et al. 2007). Some of the ecological effects of this species, such as reduction of native algal abundance and inhibition of native kelp recruitment have been assessed in Washington, California, and Baja California (Norton 1977; Ambrose and Nelson 1982; Espinoza 1990; Aguilar-Rosas and Machado Galindo 1990; Britton-Simmons 2004). Yet despite its widespread presence in southern California, there have been few studies examining the effects of this naturalized species in this area (Deysner and Norton 1982; Miller et al. 2011) or its current distribution.

Sargassum horneri was first discovered in Long Beach Harbor in 2003 (Miller et al. 2007), the first instance of this species outside of its native range (Miller et al. 2007). *Sargassum horneri* is one of the most abundant members of the algal community in temperate areas of Japan and Korea (Choi et al. 2003; Pang et al. 2009). This alga is an ecosystem engineer in these areas, growing up to 5 m tall in dense forests that provide habitat and spawning grounds for a diverse assemblage of organisms (Choi et al. 2003; Choi et al. 2008). *Sargassum horneri* is known for its high reproductive capacity, ability to rapidly colonize new areas, and fast growth rate (3–5 m in 10 months) (Choi et al. 2003). Due to its life history characteristics and its rapid spread in the short time frame since its original introduction, *S. horneri* is recognized as having the potential to be highly invasive in Southern California, Baja California, and other areas along the west coast of North America (Nyberg and Wallentinus 2005; Miller et al. 2011). Despite the rapid invasion of *S. horneri*, little is known about its current distribution and ecological impacts in southern California and Baja California.

The goal of this study was to provide detailed information on the distribution of *S. horneri* and *U. pinnatifida* on the San Diego County coast, and to analyze patterns of establishment, spread, and persistence of these seaweeds along the California and Baja California coasts. Specifically, our first goal was to describe the distribution of these non-native algae in San Diego County. Second, we documented how the presence of these species has changed with regard to: the number of locations they have become established; spread of populations within invaded sites; and persistence of populations. Third, we compared

invasion locations to ascertain whether certain habitats appear to be more invasion prone than others. Finally we assess the occurrence of these species in San Diego County within the context of the invasion of the broader California and Baja California coastlines.

Methods

We used three approaches to describe the distribution, abundance, and invasion patterns of non-native macroalgae in San Diego and the broader region: broad-scale qualitative presence/absence surveys; smaller-scale quantitative benthic community surveys; and a synthesis of published and unpublished literature.

Site selection

Thirty-two sites (10s of m in extent) in eight locations (1–10 km apart, Figure 1) in San Diego County were assessed (n=1–7 sites per location, depending on availability of suitable habitat within each location). Surveys were initially conducted in January 2012 at Mariner’s Cove, Mission Bay, where the first population of *S. horneri* was discovered. Four additional sites with rip-rap substrate similar to Mariner’s Cove were surveyed between February and July 2013 (Supplementary material Table S1). In July 2013, permanent sites for qualitative and quantitative surveys were established. These sites were located between Oceanside Harbor and San Diego Bay. All sites were then surveyed during summer 2013 (23 July 2013 – 7 August 2013), winter 2013 (8 December 2013 – 20 December 2013), and summer 2014 (2 July 2014 – 1 August 2014).

Survey locations were grouped into three site types based on site characteristics: harbors (n=3), open coast jetties (n=2), and natural open coast locations (n=3). Harbor locations included San Diego’s three main harbors: Oceanside Harbor, Mission Bay, and San Diego Bay. The two open coast jetties, Ponto Jetty and Del Mar Rivermouth were located between Oceanside Harbor and Mission Bay. Open coast locations were La Jolla Cove in the Matlahuayl State Marine Reserve, Bird Rock in the South La Jolla State Marine Reserve, and the Cabrillo National Monument State Marine Reserve. Because it is an area of special ecological and management interest, Cabrillo National Monument was surveyed five times: fall 2013 (20 October 2013) and spring 2014 (2 April 2014) and the three survey rounds listed above.

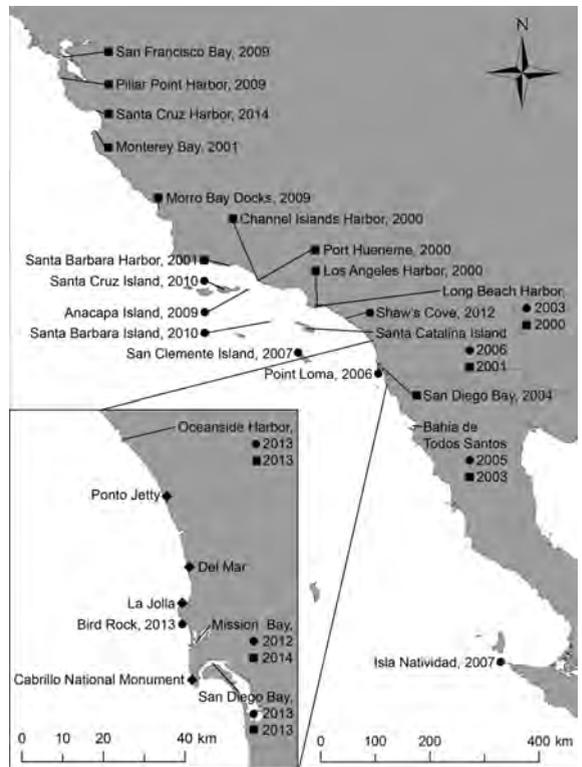


Figure 1. Map of locations where the non-native algae: *S. horneri* (circles) and *U. pinnatifida* (squares) have been documented; diamonds indicate locations where no non-native seaweeds were documented. The larger map presents results from our literature review and from herbarium collections while the inset map presents results from our survey of San Diego County. The year where each species was first documented is also shown for each location.

Within each of the eight locations, survey sites were established on hard bottom substrate suitable for the growth of macroalgae. Harbor and jetty sites were established on rip-rap rock that typically terminated in sand at maximum depths of 5 m. For open coast locations, survey sites were chosen from a habitat map generated in ArcGIS with LIDAR data. Fifteen stratified random coordinates in each open coast location were generated then ground-truthed for suitability (hard substrate, depths from 0–5m). From these, three points in each location were randomly selected as survey sites.

Qualitative surveys

Rapid qualitative surveys were conducted at 32 sites across all eight locations (Table S1) to note the establishment of populations at new sites and to describe how established populations were

Table 1. Summary information from presence-absence surveys with estimated peak abundance (# stipes / site) of non-native brown macroalgae at survey sites for all sampling rounds (Winter / Summer 2012–2014). Sites that were not sampled are shown with “ns”, white indicates absence of non-native macroalgae, light grey indicates *S. horneri* was found, dark grey indicates *U. pinnatifida* was found, and black indicates both species were found concurrently. Categorical abundances are shown as follows: absent (-); 1–10 stipes (+); 11–100 stipes (++); 101–1000 stipes (+++); >1000 stipes (++++).

Location	Site	Winter 2012	Spring 2013	Summer 2013	Winter 2013	Summer 2014	<i>S. horneri</i>	<i>U. pinnatifida</i>
Oceanside Harbor	Oceanside Harbor North	ns	ns				-	+
	Oceanside Harbor, Marker 6	ns					-	++
	Oceanside Harbor, Marker 4	ns					+	++
	Oceanside Harbor Docks	ns					-	++
Ponto Jetty	Ponto Jetty	ns					-	-
Del Mar	Del Mar Rivermouth	ns					-	-
	9th Street	ns	ns				-	-
	Flat Rock, Torrey Pines	ns	ns				-	-
La Jolla	Dike Rock, Scripps	ns	ns				-	-
	La Jolla Cove East	ns					-	-
	La Jolla Cove West	ns					-	-
	La Jolla Cove Central	ns					-	-
	Boomers Cove	ns	ns				-	-
	Casa Cove	ns	ns				-	-
	Marine Street	ns	ns				-	-
			ns	ns			-	-
Bird Rock	Bird Rock North	ns	ns				-	-
	Bird Rock Central	ns	ns				++++	-
	Bird Rock South	ns	ns				++++	-
Mission Bay	Mission Point						++++	-
	Hospitality Point	ns	ns				++++	-
	Vacation Island	ns	ns				++++	-
	Quivira Basin	ns	ns				++++	++
San Diego Bay	Harbor Island East	ns	ns				-	+
	Harbor Island Central	ns	ns				-	+
	Harbor Island West	ns					-	-
	Shelter Island North	ns					++++	-
	Shelter Island South	ns	ns				++++	-
	Marina Park, Seaport Village	ns	ns				-	-
	Coronado Ferry Terminal	ns	ns				-	++
Cabrillo Natl. Monument	North Cabrillo	ns	ns				-	-
	Central Cabrillo	ns	ns				-	-
	South Cabrillo	ns	ns				-	-
	# locations where non-native seaweeds found:	1	4	3	10	13	9	8

spreading in spatial extent through time for large swaths of coastline. At each site, we searched for *S. horneri* and *U. pinnatifida* at depths of 0–5 m along as much of the coastline as possible, using SCUBA in some sites to access deeper reefs. Hard bottom substrates in harbors, including harbor breakwalls and jetties, rip-rap, and along docks and dock pilings, were searched. At open coast sites (including jetties), hard bottom substrata

was searched, with a special focus on areas of low wave exposure.

Presence-absence and relative abundance (<10, 11–100, 101–1000, >1000 stipes per site) of *S. horneri* and *U. pinnatifida* were recorded. When either of these species was encountered, habitat characteristics (depth, substrate type, exposure to current and waves) and size and reproductive status of the algae also were recorded.

Quantitative surveys

Quantitative surveys were conducted at twenty sites across all eight locations (Table S1) to describe changes in non-native algal density through time and to determine if patterns of density and distribution existed with respect to benthic composition of survey locations. At each site, three 5 m transects were set 5 to 10 m apart, perpendicular to shore from 0–5 m depth. In five 1-m² quadrats placed on alternating sides of each transect line, brown macroalgal taxa (> 10 cm tall) were identified to species and the number of stipes was counted. In each quadrat, visual estimates of percentage of substrate covered were also made to the functional group level, which included all abiotic (bare rock, sand, shell), and biotic (articulated coralline algae, crustose coralline algae, fleshy crust, turf algae, brown, green and red fleshy macroalgae, seagrass, and sessile benthic invertebrates) components of the benthic community. Quadrats that contained substrate unsuitable for the growth of macroalgae (100% sand) were removed from the data set so that densities were reported per area of available hard bottom habitat.

Statistical analysis

Our hierarchical sampling scheme was designed to allow comparisons of non-native algal populations at the site, location, and site type (harbor versus open coast) level. To compare densities of native, non-native, and non-native naturalized brown algae (*S. muticum*) among sites, mean site-level stipe densities (# stipes / m²) were calculated for each site and sampling round. To compare non-native algal abundance between site types, a three-factor analysis of variance (ANOVA) was used with site type and sampling round as fixed effects and location as a random effect nested within site type. Jetties were not included in the comparison among site types due to the low number of jetty sites (n=2). To explore how algae may use space in different habitats, we plotted native versus non-native site-level mean stipe densities for each sampling round.

Benthic cover data from quantitative surveys were examined using principal components analysis (PCA). Scores along the first PC axis were used to examine if densities of native, non-native naturalized or non-native taxa were related to benthic composition across our data set. Statistical analyses were performed using SigmaPlot 13 (Systat Software Inc., San Jose, California, USA) and JMP 12 (SAS Institute Inc., Cary, North Carolina, USA).

Literature review and synthesis

To provide an updated regional distribution for both species, all published and unpublished accounts of *S. horneri* and *U. pinnatifida* on the Pacific coast of North America were gathered from ISI Web of Science and Google Scholar, the University of California Herbarium database (<https://webapps.cspace.berkeley.edu/ucjeps/publicsearch/publicsearch/>), and personal correspondence with researchers. Web of Science and Google Scholar were searched using the key words: Baja California, California, distribution, invasive algae, *Sargassum horneri*, and *Undaria pinnatifida*. Discovery dates, identifier, location, latitude/longitude, and any depth, habitat and density information were recorded.

Results

San Diego County distribution

Sargassum horneri was found at 28% of the thirty-two sites and *U. pinnatifida* was found at 25% of the sites (Table 1). In all cases, non-native algae were found at sites where they had not previously been documented. Overall, non-native algae occurred in 43.75% of San Diego sites surveyed, and occurred disproportionately in harbor sites, with 86.7% of harbor sites having non-natives present at some point during sampling. These two invaders were found at 13.3% of open coast sites and never found to occur in jetty sites. Both species occurred together at two of San Diego's three harbors, Oceanside Harbor, and Mission Bay. In general, native brown macroalgal species dominated at our survey sites, contributing $56.7 \pm 1.94\%$ (mean \pm SE) of all macroalgal stipes. The non-native naturalized alga (*S. muticum*) made up $29.1 \pm 1.74\%$, and non-native brown macroalgae made up $14.2 \pm 1.31\%$ of stipes. For the individual non-native macroalgal species, *S. horneri* contributed $12.4 \pm 1.26\%$, and *U. pinnatifida* made up $1.8 \pm 0.47\%$ of macroalgal stipes across all study sites.

Establishment of new populations in San Diego through time

The number of sites where *S. horneri* was found increased during our study from one to nine sites (Table 1). On 15 January 2012, *S. horneri* was discovered at a single site at Mission Point in Mission Bay. Spring 2013 surveys documented no new populations of *S. horneri*, though a second survey of Mission Point revealed a persistent, dense and localized population. All *S. horneri*

populations discovered during our survey effort persisted throughout the duration of the study. During our first comprehensive survey of thirty-two sites (summer 2013), *S. horneri* was found at two new sites, in Bird Rock South, an open coast site in Bird Rock, and on the south end of Shelter Island in San Diego Bay (Table 1). At Bird Rock, juvenile *S. horneri* thalli were found in the 3–5 m depth range on cobble coated in crustose coralline algae. This was the only open coast location to have *S. horneri* throughout our survey. At Shelter Island, *S. horneri* was found growing at depths of 1–5 m along the rip-rap breakwall on the south end of the island near the marina in an area of high boat traffic.

During the winter 2013 survey, *S. horneri* was found at five new sites (Table 1). The species appeared intermingled with native algae in a small patch near the mouth of Oceanside Harbor (Oceanside Harbor North). The previously localized population at Bird Rock South spread to the Bird Rock Central site. *Sargassum horneri* was also found at three new sites in Mission Bay: at Hospitality Point, in the boat marina at Quivira Basin, and on a rip-rap breakwall near the boat ramp at Vacation Island.

During our final comprehensive sampling round, summer 2014, *S. horneri* was found at one new site, Shelter Island North. While in past surveys the species was localized at Shelter Island South, during this final survey it was observed growing along the entire length of the harbor breakwall.

Overall, we found *U. pinnatifida* at eight sites in San Diego County, and the number of sites in which it was present increased through time (Table 1). *Undaria pinnatifida* was first found at three sites in Oceanside Harbor in spring 2013: near the mouth of the harbor on a rip-rap breakwall (Oceanside Harbor, Marker 6), deeper in the harbor on rocks surrounded by soft muddy substrate, (Oceanside Harbor, Marker 4), and attached to the underside of 10–15 docks within the marina (Oceanside Harbor Docks). *Undaria pinnatifida* was not found at any site during the summer 2013 survey, including the Oceanside Harbor sites. During the winter 2013 survey, *U. pinnatifida* was found at the eastern end of Harbor Island and at the Coronado Ferry Terminal. At Harbor Island we found a group of large isolated thalli (approx. 1–2 m length) on a rip-rap breakwall, a cement breakwall, and on pilings. In Coronado, *U. pinnatifida* was observed on the underside of the ferry landing docks. In summer 2014, *U. pinnatifida* reappeared in Oceanside Harbor at the same three sites it was previously

found and was found at Quivira Basin in Mission Bay and the central part of Harbor Island.

Spatial spread at sites through time

Within established sites, *S. horneri* consistently increased its spatial extent through time. At Mission Point in Mission Bay, this species was initially confined to a small section of protected rip-rap within Mariner's Cove growing on bare rock in an area sparsely populated by *S. muticum* and the native species *Dictyopteris undulata* Holmes and *Dictyota flabellata* (F.S. Collins) Setchell and N.L. Gardner. Further west on the harbor breakwall, where tidal current flows are much higher, and where native kelps (e.g. *Macrocystis pyrifera* (Linnaeus) C. Agardh, *Egregia menziesii* (Turner) Areschoug, and *Eisenia arborea* (Areschoug) occur in higher density than on the inner breakwall, no *S. horneri* was found. This population remained localized between January 2012 and spring 2013, but in summer 2013 the length of the breakwall occupied by *S. horneri* had expanded by roughly 0.33 km, with new recruits occurring in patches moving outward toward the mouth of Mission Bay. During winter 2013, these recruits developed into mature and fertile adult thalli, which then produced another cohort of recruits approximately 0.6 km further west on the breakwall in summer 2014.

A clear pattern of population expansion with each recruitment cycle was also observed at other sites. At Hospitality Point the population on the inner breakwall spread westward toward the mouth of the harbor with each recruitment cycle. At Shelter Island, the species spread from the south end of the island north, eventually reaching the north end by winter 2013. At Bird Rock, the population was discovered in a small patch at the center of the cove in summer of 2013, but eventually occupied the majority of the cove by winter 2013, again spreading with each recruitment event.

In contrast, *Undaria pinnatifida* occurred in low density populations that remained localized through time. At all sites where it was observed, densities were highest in spring to late summer, following the annual pattern of recruitment and development seen in native populations (Saito 1975) and previously observed in Santa Barbara Harbor (Thornber et al. 2004) and at Santa Catalina Island (Miller and Engle 2009). While other populations die off entirely in the late summer or early fall (Miller and Engle 2009), mature reproductive adults were observed in low densities year round in San Diego locations.

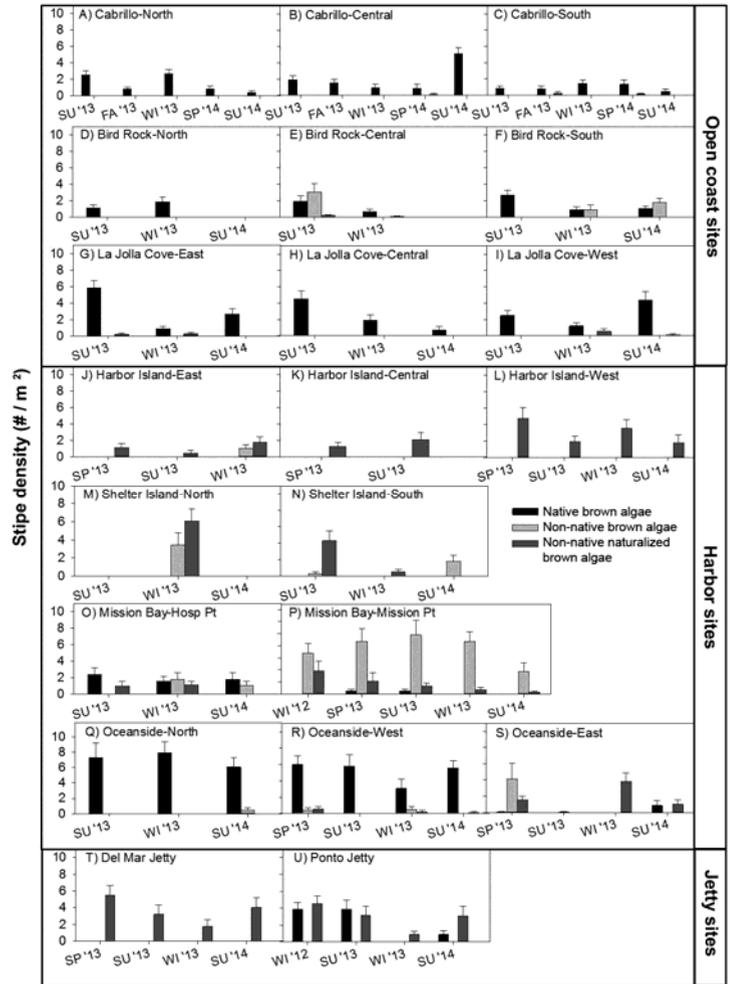


Figure 2. Mean density (\pm SE) of native (black), non-native (*S. horneri* and *U. pinnatifida*, light grey), and non-native naturalized (*S. muticum*, dark grey) brown macroalgae at quantitative survey sites in San Diego County during each survey round (Winter 2012, Spring 2013, Summer 2013, Fall 2013, Winter 2013, Spring 2014, Summer 2014 (abbreviated in figure)) and grouped by site type.

Changes in density through time

Despite an increase in both the number of sites where *S. horneri* and *U. pinnatifida* were present, and in the spatial extent of their populations, the density at each site did not increase for either species (Figure 2). At Bird Rock central and Bird Rock south (Figure 2E and F), the open coast sites where *S. horneri* was found, populations were patchy and densities were consistently low. Harbor sites had persistent but consistently low density populations of *S. horneri* (Shelter Island North and South, Figure 2M and N, and Hospitality Point, Figure 2O) and *U. pinnatifida* (Harbor Island East, Figure 2J). Finally, at Mission Point (Figure 2P) *S. horneri* densities were consistently higher than any other site, with the mean density ranging between 4.31 ± 1.54 stipes/m² (summer 2014) and 10.08 ± 1.53 stipes/m² (winter 2013).

Habitat type and benthic composition

Mean stipe densities of non-native species were significantly higher at harbors than at open coast sites (Table 2). Overall mean stipe densities (stipes / m² \pm SE) for the Summer 2013, Winter 2013, and Summer 2014 survey rounds were 1.03 ± 0.97 , 1.91 ± 1.10 and 1.11 ± 0.56 for harbors, and 0.0 ± 0.0 , 0.47 ± 0.35 , and 0.20 ± 0.20 for open coasts. There was significant variation in density of non-native species among sites within locations and among locations within site type. There were no differences in non-native species densities among sampling rounds, nor was there an interaction between site type and sampling round. Harbor sites exhibited either high native stipe densities or high non-native stipe densities, and no site had high densities of both concurrently (Figure 3). Open coast sites had low densities of non-native species and a range of densities of

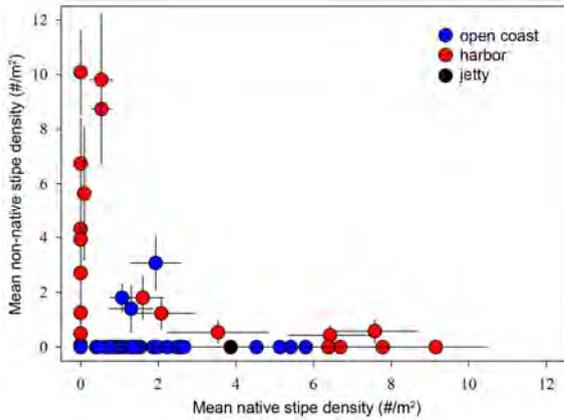


Figure 3. Mean (\pm SE) native vs. non-native algal stipe density (stipes/ m^2) for open coast sites (blue circles), harbor sites (red circles), and jetty sites (black circles).

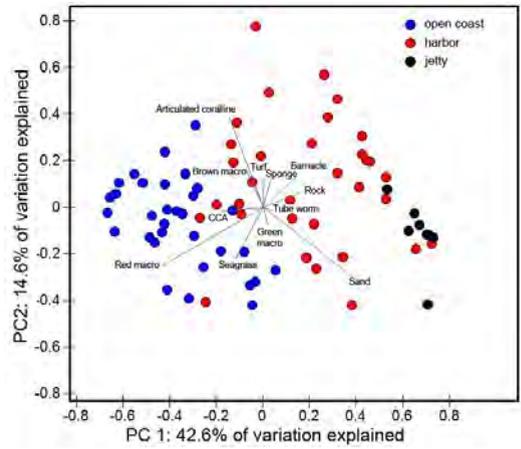


Figure 4. Principal component analysis (PCA) of major benthic groups from all sites with benthic cover survey data ($n = 32$ (open coast, blue), $n = 34$ (harbor, red), $n = 8$ (jetty, black)).

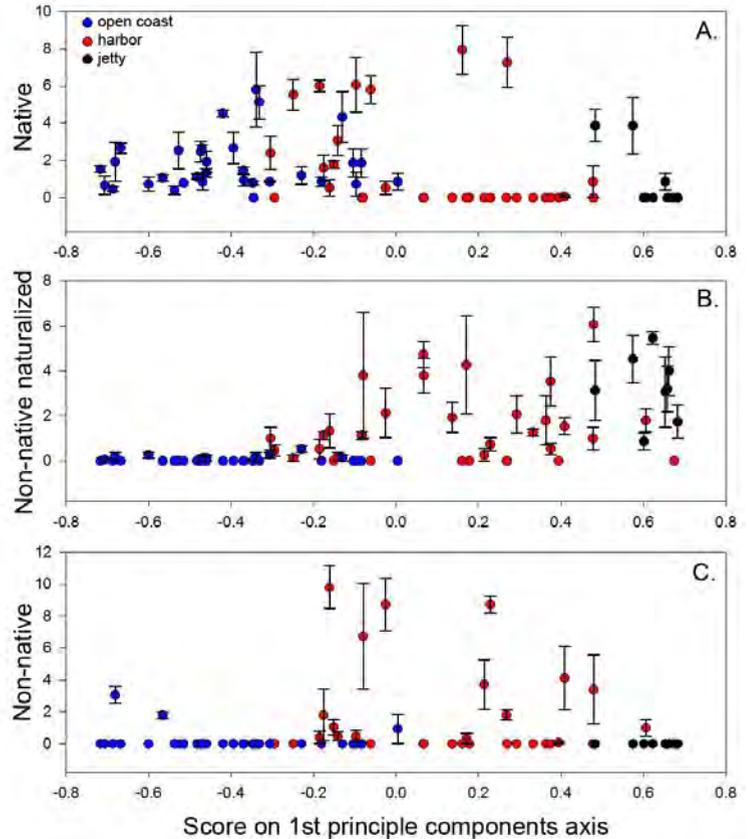


Figure 5. PC1 vs. native (A), non-native naturalized (B), and non-native (C) mean stipe densities at survey sites ($n = 32$ (open coast, blue), $n = 34$ (harbor, red), $n = 8$ (jetty, black)).

native algae. Jetty sites had intermediate densities of native species and lacked non-natives entirely. When examining benthic community composition across all sites surveyed in this study (Figure 4), the first two principal component axes (PC1 and

PC2) described over 50% of the variation in the data (with PC1 explaining 42.6% of variation). Sites within each site type clustered together along PC1 (left to right from open coast to jetty, Figure 4) suggesting that each site type was

Table 2. Results from three-way ANOVA testing for differences in mean non-native stipe densities across location nested within site type, site type (open coast versus harbor) and survey round.

Source	df	MS	SS	F	p
Site [Location, Site Type]	13	6.2331	81.0301	4.8074	0.0002
Site Type	1	26.5361	26.5361	20.4663	< 0.0001
Location [Site Type]	4	20.7358	82.9432	15.9947	< 0.0001
Site Type x Survey Round	2	0.4742	0.94833	0.3657	0.6968
Survey Round	2	1.8880	3.77605	1.4562	0.2491
Error	30	1.29658	38.8973		
Total	52				

Table 3. Summary of *S. horneri* and *U. pinnatifida* documentations on the Pacific Coast of North America from published sources and University of California Herbarium (Berkeley, California) database specimens.

Species	Year	Location	Source	Latitude	Longitude
<i>S. horneri</i>	2003	Long Beach Harbor	Miller 2007	33° 42.0' N	118° 14.0' W
	2005	Todos Santos Bay	Aguilar-Rosas 2007	31° 43.2' N	116° 40.2' W
	2006	Santa Catalina Island	Miller 2007	33° 24.3' N	118° 22.0' W
	2006	Point Loma	UC Herbarium	32° 41.2' N	117° 16.0' W
	2007	San Clemente Island	UC Herbarium	32° 58.7' N	118° 32.3' W
	2007	Isla Natividad	Riosmena-Rodriguez 2012	27° 27.8' N	115° 9.00' W
	2009	Anacapa Island	D. Kushner, US Natl. Park Service	34° 0.91' N	119° 22.5' W
	2010	Santa Cruz Island	D. Kushner, US Natl. Park Service	34° 2.61' N	119° 42.9' W
	2010	Santa Barbara Island	D. Kushner, US Natl. Park Service	32° 28.7' N	119° 24.2' W
	2012	Mission Bay	this study	32° 45.7' N	117° 14.8' W
	2012	Shaw's Cove	UC Herbarium	33° 32.6' N	117° 47.9' W
	2013	Oceanside Harbor	this study	33° 12.4' N	117° 23.6' W
	2013	San Diego Bay	this study	32° 42.4' N	117° 14.1' W
	2013	Bird Rock	this study	32° 48.9' N	117° 16.5' W
<i>U. pinnatifida</i>	2000	Los Angeles Harbor	Silva 2002	33° 42.9' N	118° 17.0' W
	2000	Long Beach Harbor	Silva 2002	33° 45.7' N	118° 12.0' W
	2000	Channel Islands Harbor	Silva 2002	34° 9.71' N	119° 13.4' W
	2000	Port Hueneme	Silva 2002	34° 9.17' N	119° 12.5' W
	2001	Santa Barbara Harbor	Silva 2002	34° 18.5' N	119° 41.4' W
	2001	Santa Catalina Island	Silva 2002	33° 24.2' N	118° 22.1' W
	2001	Monterey Bay	Silva 2002	36° 36.2' N	121° 53.3' W
	2003	Isla Todos Santos	Aguilar-Rosas 2004	31° 48.1' N	116° 47.3' W
	2004	San Diego Bay	Miller 2009	32° 42.5' N	111° 10.4' W
	2009	Morro Bay Docks	UC Herbarium	35° 22.2' N	120° 51.4' W
	2009	San Francisco Bay	Zabin 2009	37° 46.8' N	122° 23.1' W
	2009	Pillar Point Harbor	Zabin 2009	37° 30.1' N	122° 28.9' W
	2013	Oceanside Harbor	this study	33° 12.4' N	117° 23.6' W
	2014	Santa Cruz Harbor	H. Fulton- Bennett, Moss Landing Marine Laboratory	36° 57.8' N	122° 0.08' W
2014	Mission Bay	this study	32° 45.7' N	117° 14.8' W	

characterized by distinct benthic functional groups. The major loadings on PC1 were fleshy red macroalgae and crustose coralline algae in the direction of open coast sites and sand and bare rock in the direction of jetties. Sites within each classification spread along PC2, which had major loadings of articulated coralline algae, brown macroalgae, turf algae, sponges, and seagrass. This spread indicates that cover of these benthic functional groups was variable at sites within the three site types.

Densities of native, non-native naturalized, and non-native stipes were clearly grouped along PC1 based on site type (Figure 5). Native brown algal stipe densities (Figure 5A) were high in open coast sites characterized by native fleshy red macroalgae and articulated coralline algae and low in harbor and jetty sites. Harbor and jetty sites had consistently high stipe densities of non-native naturalized *S. muticum* (Figure 5B). Non-native macroalgae (Figure 5C) were found almost exclusively in harbor sites that were

characterized by turf algae, articulated coralline algae, sponges, and barnacles.

Regional abundance and current distribution

In the relatively short invasion history of *S. horneri* and *U. pinnatifida* on the Pacific coast of North America, each species has spread rapidly to occupy a range of different habitats in multiple biogeographic regions with different environmental conditions (Figure 1, Table 3). Both species have been documented on man-made and natural substrates in protected harbors, open mainland coasts, and on offshore islands.

Since its discovery in Long Beach Harbor in 2003, *S. horneri* has shown a general southward spread, remaining in the southern California Bight and expanding southward down the coast of Baja California, Mexico (Figure 1, Table 3). In this time, it has expanded its range approximately 200 km north and 750 km south, from Santa Barbara, California to Isla Natividad, Central Baja California, Mexico.

While the first documentation of *Sargassum horneri* was in a harbor, this species has been found in few harbor locations since then. In 2010, *S. horneri* was found growing in the Port of Ensenada. In this study we report the occurrence of *S. horneri* in Oceanside Harbor, Mission Bay, and San Diego Bay. *Sargassum horneri*'s greatest invasion success has been on offshore islands along the coasts of California and Baja California. *Sargassum horneri* was first discovered at Santa Catalina Island in April 2006, and it was reported at San Clemente Island in May 2007. It has since spread to Anacapa and Santa Cruz Islands. In the south, the species has been found from the Coronado Islands in 2015 (N. Kaplanis, pers. obs.), to Isla Natividad, a small island off of the central Pacific coast of Baja California, Mexico in 2007. *S. horneri* has also successfully invaded open coast mainland locations in California and Baja California, Mexico. In 2005, *Sargassum horneri* was first reported as drift wrack at La Jolla, Baja California and growing at Rancho Packard in Todos Santos Bay, Ensenada B.C. In 2006, the first population in San Diego County was discovered at New Hope Rock, Point Loma. It has since been found in isolated populations along the southern California coast in Santa Barbara (D. Reed, University of California Santa Barbara, Santa Barbara, CA, pers. comm.), Laguna Beach, and Crystal Cove, Orange County. Our study adds a total of nine sites to the list of locations where this species is now present.

Undaria pinnatifida has primarily spread northward since its discovery in Los Angeles Harbor in March 2000, growing almost exclusively on man-made structures in protected harbor locations (Figure 1, Table 3). *Undaria pinnatifida* was reported in rapid succession at harbors throughout the California coast: Port Hueneme and Santa Barbara Harbor in November 2000 and April 2001, respectively; and as far north as Monterey Harbor by August 2001. In 2004 it was first found in San Diego Bay, and by 2009 the species was also found at Morro Bay Harbor, at Pillar Point Harbor in Half Moon Bay, and marinas in San Francisco Bay. *Undaria pinnatifida* was then found on floating structures in two marinas within the Bahía Todos Santos in April of 2012, and most recently was discovered in Santa Cruz Harbor in June of 2014 (H. Fulton-Bennett, Moss Landing Marine Lab, Moss Landing, CA, pers. comm.). In two instances *U. pinnatifida* has been found on natural substrates in island locations. In June of 2001, a deep water population was found in Button Shell Cove, an open-coast location on Santa Catalina Island. This documentation represents the first and only instance of *U. pinnatifida* occurring on a natural reef on the open coast in California. In September 2003, *U. pinnatifida* was found at Isla Todos Santos, the first documentation of this species on the Pacific coast of Mexico. *Undaria pinnatifida* has not yet been documented growing on a natural reef on the mainland Pacific coast of California. In Baja California though, populations have been observed in a natural reef setting at Punta Banda, Bahía Todos Santos (Aguilar-Rosas 2014). Here we document eight new sites from San Diego harbors where *U. pinnatifida* is present.

Discussion

We investigated the presence, establishment, spread and persistence of *S. horneri* and *U. pinnatifida* in San Diego County. Both species are well established, occurring throughout the county in multiple locations characterized by distinct habitats. Further, both species appear to be spreading locally within a short time frame to an increasing number of sites where they are found. Finally, the persistence of both species at invasion locations through multiple generations indicates that these species are established.

Sargassum horneri has proven to be a successful invader in San Diego, rapidly colonizing new areas, forming dense thickets and spreading quickly within invasion sites. This suggests that

it is competitively equal to, or dominant over, native macroalgal species when conditions are right. The life history characteristics of this species may explain its success as an invasive species. Like in its native range, in invasion locations *S. horneri* grows very rapidly between November and July, reaching full size (3–5 m in length) and reproductive maturity in nine to ten months (L. Marks, University of California Santa Barbara, Santa Barbara CA, pers. comm.; N. Kaplanis pers. obs.). *Sargassum horneri* is an annual species (Gao and Hua 1997) and is capable of persisting through multiple generations because it is monoecious and extremely fecund (Miller and Engle 2007). Once released, *S. horneri* eggs have the potential to be fertilized for up to 48 hours, a window of viability much longer than related species (Pang et al. 2009). In San Diego, mature senescent thalli bearing reproductive conceptacles have been observed in spring of 2014 and 2015 throughout the county as beach wrack and as drift (N. Kaplanis, pers. obs.). Whether these drifting thalli are capable of releasing viable embryos is unknown, but it appears likely that *S. horneri* is capable of local dispersal even without a human transport vector.

Undaria pinnatifida has remained a relatively inconspicuous invader in San Diego. Its spread has been slow, has been mostly confined to man-made substrates, and no obvious ecological effects of its colonization have yet been observed. However, this survey provides only a snapshot of *U. pinnatifida* in a relatively early stage of a potential invasion. More detailed studies that investigate the interactions of this invader with the native benthic community are needed to better understand and track the progress of this invasion along the Pacific coast of North America.

Both non-native species investigated here were found almost exclusively in harbors in San Diego County. In these harbors, densities of non-native macroalgae are high when densities of native macroalgae are low and vice-versa. This pattern may result from occupation by the non-native species of an open niche that is not suitable for the growth of native macroalgae, or may be due to competitive displacement by the invaders. The disproportionate presence of these species in harbors may be a result of these locations being initial points of introduction, suggesting boats as a vector for long distance transport. Once present in harbors, the invaders may remain restricted to these habitats or they may spread into adjacent open coast sites. Whether the rocky reefs of our study area are more resistant to invasion than

harbors, or whether they have simply not been exposed to propagules of the non-native species remains to be determined. However, it appears that several offshore islands in southern California and in Baja California are highly susceptible to invasion. Whether these new open coast invasions are the result of El Niño associated conditions that have negatively impacted kelp communities, potentially opening space for invader colonization, is yet to be determined. More long-term monitoring in conjunction with experimental manipulations are needed to better understand the dynamics and potential impacts of these invaders along the Pacific coast of North America.

The results of our surveys also provide valuable insight into the distribution of the naturalized invader *S. muticum*, which was abundant at nearly every survey site. Unlike *S. horneri* and *U. pinnatifida*, *S. muticum* was abundant on open-coast jetties year round. *Sargassum muticum* was also abundant in low energy environments throughout San Diego's harbors, as well as high energy wave-swept intertidal and subtidal areas along the open coast. Further, *S. muticum* was found both in areas devoid of other macroalgae and intermingled with native macroalgal species. While *S. muticum* was ubiquitous, it was never found in dense canopy-forming stands, as it is observed in its native range (Deysher and Norton 1982) and was observed during its initial invasion of San Diego in the 1970's (P. Dayton, Scripps Institution of Oceanography, San Diego, CA, pers. comm., Ambrose and Nelson 1982). At present, it appears as though *S. muticum* has become naturalized in San Diego but little is known about how this species interacts with native benthic communities or the new invaders over time. Continued monitoring is needed to better understand the invasion ecology of these three non-native species.

Comparing patterns of invasion of these macroalgal species along the San Diego County coast to the broader coastal region provides important context to understanding patterns of spread. In San Diego County, *S. horneri* grows in large meadows in the local harbors. These harbor populations are similar to the extensive populations now observed on the leeward side of Santa Catalina Island, though their spatial extent is more confined by limited availability of suitable hard substrate. On the open coast of San Diego, *S. horneri* remains contained in small localized populations with small spatial coverage and lower densities. In the wave and current exposed areas along the west and southern coasts of Santa

Catalina, the Northern Channel Islands, and the southern California mainland, *S. horneri* has also not yet been observed to form large or persistent meadows. The mechanisms driving these patterns of establishment remain unclear but may be tied to wave and current exposure.

Despite *U. pinnatifida*'s reputation as an aggressive invader, the colonization pattern for San Diego, as with the rest of the Pacific coast of North America, has shown that *U. pinnatifida* is largely restricted to man-made structures in harbors. This is strikingly different from other invasion locations such as Australia and New Zealand, where widespread invasion on the open coast has prompted aggressive removal and control programs (Lonhart and Bunzel 2009). Instances where *U. pinnatifida* has invaded natural substrates on the open coast of California and Baja California remain rare despite fears that these observed populations are the beginning of a widespread and devastating invasion. The pattern of colonization along the Pacific coast of North America may be a result of a limited temperature tolerance (Aguilar-Rosas et al 2004; Miller and Engle 2009), an inability to become established in areas of high wave exposure (Miller and Engle 2009), or an inability to compete with native macroalgae for settlement space on the benthos – but these mechanisms have yet to be explored.

The spread of *S. horneri* and *U. pinnatifida* along the Pacific Coast of North America in the past two decades has been swift and reveals that these two species are capable of becoming invasive in a range of habitats within this region. Colonization of areas far from their native ranges indicates that these species are capable of utilizing a human-mediated transport vector. Distribution patterns suggest hull fouling of large commercial vessels as a likely vector for initial introduction and fouling of smaller recreational vessels as a vector for secondary spread. Further, their capability to spread locally from these initial points of introduction may also suggest secondary spread through sexual and asexual propagation. These two species have also proven to be highly versatile. While *S. horneri* has remained confined to the southern California Bight and the Baja California Coast, it has successfully colonized a wide range of habitat types in this region. *Undaria pinnatifida* has also proven capable of invading a variety of habitat types, and has expanded its range from Baja California to northern California, spanning across multiple distinct biogeographic provinces. Finally, the persistence of both species since their initial

introductions indicates they are also able to withstand competition and with native algal species and grazing pressure from native herbivores.

The ecological and economic impacts of these seaweed invaders have yet to be explored. In its native range, *S. horneri* is known to influence a variety of different coastal environmental parameters including dissolved oxygen concentration, water flow, pH, and light conditions (Komatsu et al. 2007). It is also known to play an important ecological role in offshore waters, forming large, dense, drifting mats (Komatsu et al. 2007). In its native range, this species is an important biofilter that removes inorganic nutrients from mainland effluent discharges (Pang et al. 2009). The impacts of the large invasive stands and drifting mats of this species on the Pacific coast of North America on coastal environmental conditions and nutrient distributions remains unknown. Few studies have assessed the impacts of *U. pinnatifida* on native communities in other areas, and ecological effects of *U. pinnatifida* on native species have been variable based on invasion location. Further, little is known about how this species may affect the rocky reef communities of the Pacific coast of North America if it spreads further (Lonhart and Bunzel 2009).

The coastal communities of the Californias are currently undergoing invasion by multiple non-native macroalgae. The majority of these species have appeared in the past 30 years, and species such as *S. horneri* and *U. pinnatifida* are still in the early stages of the invasion process, providing the opportunity to gain insight into the early stages of algal invasions. Further, environmental shifts associated with climate change, including increases in the frequency and intensity of ENSO events, may be making the California and Baja California coasts more susceptible to invasion by non-native algal species through creating more space and reducing natural resistance (Miller et al. 2011). While the current distributions of these species may be confined by latitudinal temperature barriers, with the North Equatorial Current possibly confining the spread of *U. pinnatifida* south, and the California Current possibly confining the spread of *S. horneri* north, temperature shifts associated with climate change could potentially alter these barriers and allow for further spread of these species. Identifying the underlying mechanisms that facilitate or inhibit further spread is the next logical step in advancing our knowledge of the invasion ecology of these species.

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The following supplementary material is available for this article:

Table S1. Summary of surveys in San Diego County.

This material is available as part of online article from:

http://www.aquaticinvasions.net/2016/Supplements/AI_2016_Kaplanis_etal_Supplement.xls

Rapid Communication

Range expansion of a non-native, invasive macroalga *Sargassum horneri* (Turner) C. Agardh, 1820 in the eastern Pacific

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Abstract

Sargassum horneri (Turner) C. Agardh, 1820 is a fast growing brown alga native to shallow reefs of eastern Asia. It has spread aggressively throughout southern California, USA, and Baja California, México since it was discovered in the eastern Pacific in 2003 and poses a major threat to the sustainability of native marine ecosystems in this region. Here we present a chronology of the rapid geographic expansion of *S. horneri* in the eastern Pacific and discuss factors that potentially influence its spread.

Key words: introduced species, invasion, distribution, seaweed, *Sargassum filicinum*, southern California, Baja California

Introduction

Introductions of marine non-native species continue worldwide and are expected to increase with the expansion of global trade. The spread and ecological effects of newly-established non-native species can vary; some proliferate and compete vigorously in their introduced range and are considered “invasive” (Miller et al. 2011). Introduced marine macroalgae are no exception, although detailed records of the geographic expansion of introduced marine macroalgae are rare (Lyons and Scheibling 2009) despite there being at least 277 introduced seaweed species

globally (Williams and Smith 2007). Documenting the spread of these species can be challenging given the logistical difficulties associated with sampling in subtidal habitats where they occur (e.g. time- and depth-limitations when using scuba and the expense of accessing remote sites). Yet such studies are valuable for not only documenting their distributions but also providing insight into the mechanisms influencing the spread of non-native species.

Here we present the chronology of the geographic expansion of the non-native macroalga *Sargassum horneri* (Turner) C. Agardh, 1820 (Fucales) along the southern region of the Pacific



Figure 1. *Sargassum horneri* morphology and life cycle. (A) Recruit, (B) Mature thallus with reproductive receptacles indicated by arrow, (C) Thick canopy on a shallow reef. Photo credits: Jessie Alstatt (A), Dan Richards (B), Tom Boyd (C).

coast of North America, where it has spread rapidly since it was first detected in Long Beach Harbor, California, USA, in 2003 (Miller et al. 2007). We also discuss potential factors influencing the spread of this species and the implications of its invasion to native ecosystems.

Study area

The study area encompassed the shallow coastal waters of the eastern Pacific Ocean from northern California, USA, to the southern tip of Baja California, México. Much of this coast is actively monitored by government and academic researchers and citizen scientists, and is therefore an ideal region in which to document the spatio-temporal dynamics of the spread of an invasive macroalga.

Study species

Miller et al. (2007) initially identified the introduced population of *Sargassum* discovered in Long Beach, California as *S. filicinum* (Harvey, 1860). This annual brown alga is monoecious, with ellipsoidal pneumatocysts, and has a narrow geographic range on the coast of western Japan (Yoshida et al. 1983; Tseng et al. 1985) and southern Korea (Lee and Yoo 1992). On the basis of molecular population studies, Uwai et al. (2009) merged *S. filicinum* with *S. horneri*, a dioecious species with spherical pneumatocysts that is widespread in the warmer waters of eastern Asia (Tseng et al. 1985). Therefore, we refer to the eastern Pacific population as *S. horneri*.

The morphology of *S. horneri* changes throughout its annual, diplonic, life cycle. Embryos develop into small plants with lateral fern-like branches anchored by a common holdfast (Figure 1A). Plants give rise to a single erect frond up to several meters in length that bears numerous vegetative blades buoyed by many small gas bladders (Yoshida 1983). Eventually, the frond ceases vertical growth and develops hundreds of reproductive receptacles (Figure 1B). Fertilization occurs when sperm penetrate an egg inside a receptacle positioned on the surface of a receptacle. The resulting embryo is released and settles to the bottom. After embryos are shed the frond senesces and the entire thallus dies, completing the life cycle. Sexual reproduction is the only known means of propagation.

Miller et al. (2007) recorded the presence of *S. horneri* in the eastern Pacific in southern California shortly after it was first discovered in 2003. Rapid communication, coupled with the species' conspicuous morphology and widely distributed

information on its identification, facilitated the subsequent monitoring of *S. horneri* by many researchers in California and Baja California.

Methods

We compiled records of *S. horneri* from herbaria, publications, government and academic groups and trained citizen scientists monitoring subtidal and intertidal reefs in California and Baja California (Supplementary material Table S1). Its presence or absence was recorded during ecological surveys by observers trained to identify the species. Because this region is extensively and regularly monitored by many trained observers, the spread of *S. horneri* can be described with high spatial and temporal resolution. Using these data, we present a timeline of *S. horneri* spread in southern California and Baja California.

Results

Since 2003 when *S. horneri* was first detected in Long Beach Harbor, it has spread north and south along the mainland coast and westward across several nearshore islands (Supplemental material Table S2, Figure 2). The geographic expansion of *S. horneri* is characterized by isolated introductions to new islands and locations on the mainland widely separated from existing populations, followed by the steady colonization of surrounding areas.

In 2005, just two years after *S. horneri* was first detected in Long Beach, it was found drifting on the surface 260 km south in Todos Santos Bay, Baja California, México. One year later it was confirmed to be growing on natural reefs there, and along the coast of San Diego and the leeward side of Santa Catalina Island, California. Since then it has progressively spread north in southern California and south in Baja California. By 2007, *S. horneri* had spread to Isla Natividad in Baja California, 500 km south of the nearest known population. In 2013, the northern range of *S. horneri* reached Santa Barbara, California, 186 km northwest of Long Beach. The known northern and southern limits of the range of established populations have not changed since 2013, though additional populations were recorded within the previously established range. However, individual thalli were found floating at the surface west of the current range at Santa Rosa and San Miguel Islands in 2012 and 2015 respectively, and at multiple islands near the southern end of its range in 2015.

The abundance and persistence of the recorded populations varied. Many reports consisted of only a few individuals or groups of individuals in small patches, often at sites where *S. horneri* had not previously been recorded. Patchy distribution continued in subsequent years at many sites, and occasionally *S. horneri* was recorded at a site but not found there again. However, in some areas, such as Santa Catalina and Anacapa Islands, *S. horneri* spread profusely and was persistent, covering large portions of reefs with adult densities $> 100 \text{ m}^{-2}$ and recruit densities $> 1000 \text{ m}^{-2}$ for multiple years (e.g. Figure 1C).

Discussion

Invasive traits

Sargassum horneri has several life history characteristics that make it well adapted for colonizing distant areas and rapidly populating an area once it is established. Thalli are buoyed by gas bladders and are capable of self-fertilization, making the establishment of new populations from long-range dispersal of a single floating thallus possible. Indeed, floating *S. horneri* thalli have been observed frequently off southern California and Baja California and are estimated to remain afloat for several weeks before decomposing (Yatsuya 2008). Local population growth can occur quickly because *S. horneri* is a fast-growing ($4.46\% \text{ day}^{-1}$ adult blade weight maximum relative growth rate; Choi et al. 2008) and highly fecund alga (up to 50% of the biomass of a mature individual is composed of reproductive tissue; L. Marks unpublished data). Furthermore, the patchy distribution and reoccurrence of dense aggregations of *S. horneri* in successive years (Figure 1C) may be explained by the heavy embryos of *S. horneri* which, like other furoid algae, are thought to have limited capacity for dispersal.

Dispersal vectors

The distribution and rapid spread of *S. horneri* is likely influenced by both natural and human-mediated dispersal. Reproductive thalli can become dislodged naturally if severed from their holdfast by grazers or strong wave action and carried to new sites on ocean currents. Divers may also dislodge thalli accidentally or intentionally, inadvertently contributing to its dispersal by either freeing them to float away or transporting them elsewhere. Boaters can dislodge thalli when setting and retrieving anchors lying in *S. horneri*-populated areas. *Sargassum horneri* is also adept

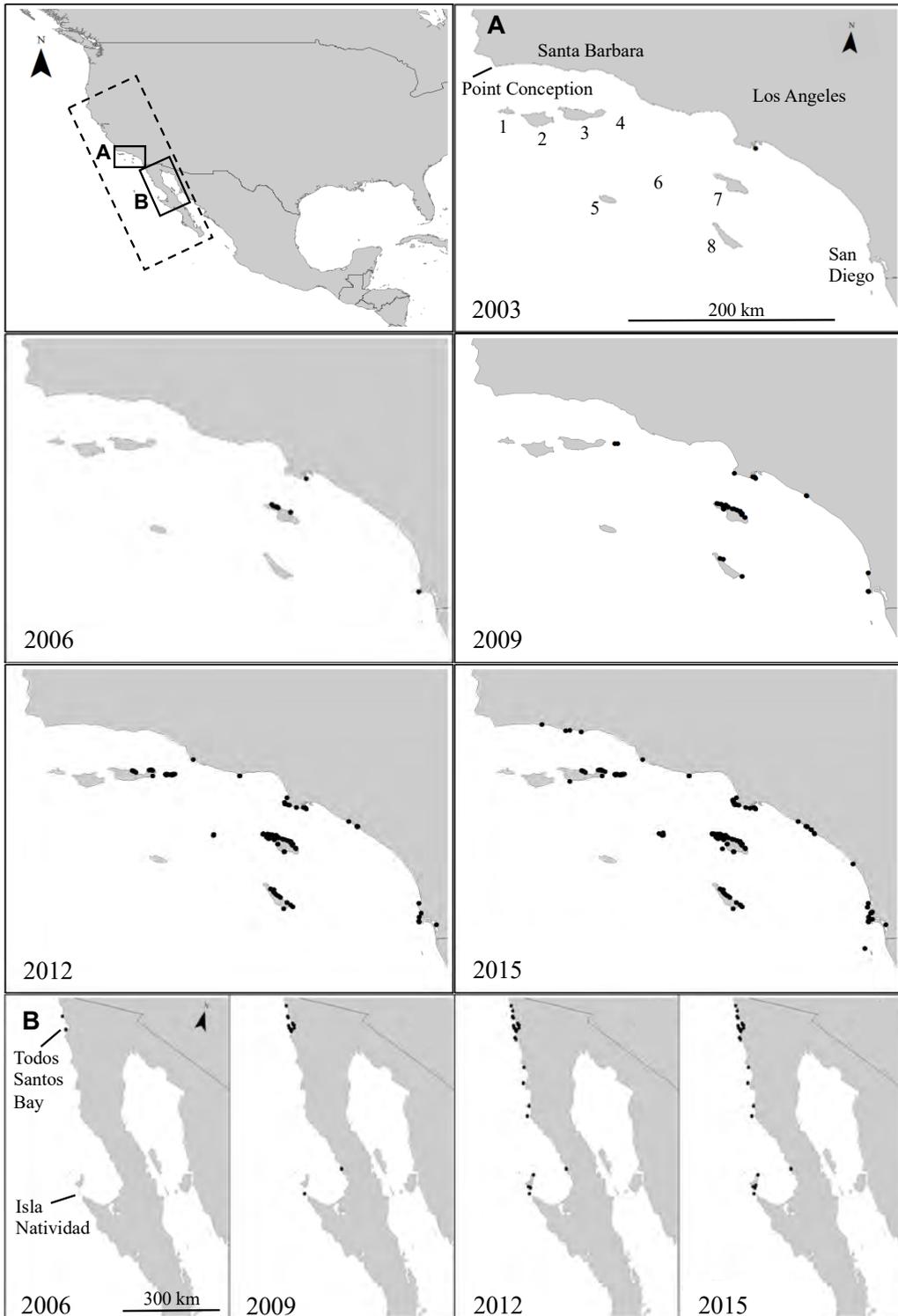


Figure 2. Chronology of geographic expansion of *S. horneri* in the southern California, USA (A) and Baja California, México (B) regions. Each dot identifies a location where *S. horneri* was found attached to the substrate at least once. Each map includes observations from all previous years to display the distribution of *S. horneri* at each interval. The California Channel Islands are identified by number as follows: 1. San Miguel, 2. Santa Rosa, 3. Santa Cruz, 4. Anacapa, 5. San Nicolas, 6. Santa Barbara, 7. Santa Catalina, 8. San Clemente. The entire study area is outlined by the dashed line. Maps were created by P. Carlson.

at growing on a variety of both natural and man-made surfaces, so vessels fouled with *S. horneri* may be an effective means of transporting it to new locations as suggested by Miller and Engle (2009). The current distribution of this seaweed includes many sites that are frequently visited by boaters and divers, such as harbors or anchorages, supporting the idea that *S. horneri* is being transported regionally by recreational and commercial vessels.

Potential for further spread

Sargassum horneri has expanded significantly further south along the eastern Pacific coast than north, spanning 6.18 and 0.76 degrees latitude from the initial detection site in Long Beach Harbor, respectively. The thermal tolerances of *S. horneri* may play an important role in determining range limits in the eastern Pacific. Sea surface temperatures in its native range in western Japan and southern Korea average between about 18–22° C (Chu et al. 1998). Baja California water temperatures typically range between 14–22° C on the Pacific coast (Zaytsev et al. 2003), so the continued expansion of *S. horneri* southward along this peninsula is likely. Warmer average temperatures in the Gulf of California and mainland México will likely prevent expansion beyond the peninsula. Ocean temperatures north of Point Conception rarely exceed 18°C, which may prevent *S. horneri* from spreading further north under present ocean climate conditions. However, predictions for a warmer ocean in the future may serve to increase the northward expansion of *S. horneri* in the eastern Pacific.

Implications of S. horneri invasion

Sargassum horneri can be locally very abundant and highly persistent. Therefore, its continued expansion in the eastern Pacific may pose a major threat to the sustainability of native marine ecosystems. Its high growth rates and long, floating thalli may provide a competitive advantage over other macrophytes. In addition, it appears to be avoided by most herbivores (Navarro 2009; Vogt 2010), possibly due to high concentrations of phenolic compounds that have been shown to deter grazing in other furoid algae (Steinberg 1985). Mesoinvertebrates that use macroalgae as biogenic habitat and the fish that depend on these invertebrates may also be affected by the *S. horneri* invasion. Research investigating the interactions between *S. horneri* and ecologically important species is critically needed to understand

how its invasion may be altering the structure and functioning of existing ecosystems of the eastern Pacific.

Continued monitoring of *S. horneri* distribution is essential to identify environmental factors influencing its spread and prioritize management actions. Researchers and citizen scientists can contribute to this effort by reporting observations of *S. horneri* occurrence to an online database and map designed to help track its spread (Marine Invasive Species Tracking website 2015).

Conclusion

The range of *S. horneri* has expanded rapidly in the eastern Pacific since it was first detected in 2003. Its expansion to the south has been more extensive and occurred more quickly than to the north, suggesting that it may be better suited to warmer southern waters. The prevalence of *S. horneri* at popular boating and diving destinations suggests that its spread is the result of multiple introductions. The life history of this species allows distant areas to be colonized by a single individual, which facilitates its spread. The high abundance and persistence of *S. horneri* in novel areas has heightened the awareness of its invasion potential and raised concerns about its possible adverse effects on existing ecosystems. Future research aimed at determining the environmental factors affecting its spread and the ecological and economic consequences of *S. horneri* invasion will provide much needed insight into the cost and need for human intervention in controlling its invasion.

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Supplementary material

The following supplementary material is available for this article:

Table S1. Sources for records of *Sargassum horneri* occurrence.

Table S2. Records of the occurrence of *Sargassum horneri* in California, USA, and Baja California, México, since 2003, the year it was first discovered in the eastern Pacific.

This material is available online for download from Long Term Ecological Research Network Data Portal (see Marks et al. 2015, <http://dx.doi.org/10.6073/pasta/63012c4e436214239ebed11ee57cbe03>)

Research Article

Assessment of control methods for the invasive seaweed *Sargassum horneri* in California, USA

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Editor's note:

This study was first presented at the 9th International Conference on Marine Bioinvasions held in Sydney, Australia, January 19–21, 2016 (<http://www.marinebioinvasions.info/previous-conferences>). Since their inception in 1999, ICMB series have provided a venue for the exchange of information on various aspects of biological invasions in marine ecosystems, including ecological research, education, management and policies tackling marine bioinvasions.

Abstract

Determining the feasibility of controlling marine invasive algae through removal is critical to developing a strategy to manage their spread and impact. To inform control strategies, we investigated the efficacy and efficiency of removing an invasive seaweed, *Sargassum horneri*, from rocky reefs in southern California, USA. We tested the efficacy of removal as a means of reducing colonization and survivorship by clearing *S. horneri* from 60 m² circular plots. We also examined whether *S. horneri* is able to regenerate from remnant holdfasts with severed stipes to determine whether efforts to control *S. horneri* require the complete removal of entire individuals. The experimental removal of *S. horneri* in early winter, just prior to the onset of reproduction, reduced recruitment in the next generation by an average of 54% and reduced survivorship to adulthood by an average of 25%. However, adult densities one year after clearing averaged 83% higher in removal plots and 115% higher in control plots. We attribute these higher densities to anomalously warm water associated with the 2015–16 El Niño that reduced native canopy-forming algae and enhanced the recruitment and survival of *S. horneri*. We did not find any evidence to suggest that *S. horneri* has the capacity to regenerate, indicating that its control via removal does not require the tedious task of ensuring the removal of all living tissue. We developed efficiency metrics for manual removal with and without the aid of an underwater suction device and found the method with maximum efficiency (biomass removed worker⁻¹ hr⁻¹) varied based on the number of divers and surface support workers. Our findings suggest that controlling *S. horneri* via removal will be most effective if done over areas much larger than 60 m² and during cool-water years that favor native algae. Such efforts should be targeted in places such as novel introduction sites or recently invaded areas of special biological or cultural significance where *S. horneri* has not yet become widely established.

Key words: introduced species, management, marine, macroalgae, rocky reef, *Sargassum filicinum***Introduction**

Invasive species are one of the greatest agents of human-induced change to ecosystems worldwide (Pejchar and Mooney 2009). Coastal marine systems are especially vulnerable to introductions of nonindigenous species via trans-oceanic shipping, aquaculture

and the aquarium trade, which have greatly extended the distribution of many marine species outside of their native ranges (Carlton 1989). Marine invasions have steadily increased over the past two centuries (Ruiz et al. 2000) and are expected to continue to rise as global trade expands. Costs associated with the impact and management of invasive species are high, totalling over \$1 billion annually in the USA

(Pimentel et al. 2000), while resources available for management are limited. Therefore, agencies tasked with controlling invasions must be efficient in their management strategies. Exploration of techniques aimed at controlling the spread and impact of marine invasive species and identification of species-specific traits that increase the efficacy of control are urgently needed.

A seaweed recently introduced to southern California, USA, presented an opportunity to test the efficacy of removal in controlling invasive algae on rocky reefs. *Sargassum horneri* (Turner) C. Agardh, 1820 (Fucales) is a large, annual brown alga native to shallow reefs of eastern Asia. It was first discovered in the eastern Pacific in Long Beach Harbor in 2003 and identified as *S. filicinum* Harvey, 1860 (Miller et al. 2007), now considered a synonym of *S. horneri* (Uwai et al. 2009). The species has spread aggressively across 700 km from Santa Barbara in southern California to Isla Natividad in Baja California, Mexico (Marks et al. 2015). It occurs primarily at offshore islands though it has also been found along the mainland and in coastal embayments. In southern California we have observed *S. horneri* growing in the intertidal down to 33 m depth, with its highest densities occurring between 5–15 m. In places where *S. horneri* is established, juveniles can attain high cover with upwards of 1,000 individuals m⁻² during the summer and fall, and these grow to form thick canopies in the winter with dense stands of over 100 adults m⁻² (author's unpublished data). While definitive evidence of ecological impacts on rocky reef systems from *S. horneri* invasion is not yet available (but see Cruz-Trejo et al. 2015), the detrimental effects on native assemblages caused by other invasive seaweeds (e.g., de Villèle and Verlaque 1995; Levin et al. 2002; Casas et al. 2004; Britton-Simmons 2004) suggest management of *S. horneri* is worth exploring (Anderson 2007; Schaffelke and Hewitt 2007; Forrest and Hopkins 2013).

Several life history characteristics of *S. horneri* make it potentially suitable for control by removal. First, it is a large and conspicuous alga consisting of a single main axis with multiple lateral branches that reaches up to several meters high (Yoshida 1983). The annual thallus is anchored by a small holdfast that gives rise to a stipe buoyed by many small gas bladders (Marks et al. 2015). The conspicuous adult thalli allow for efficient identification and removal by divers using SCUBA. Second, *S. horneri* propagates via sexual reproduction. Fertilization occurs in winter on the surface of reproductive structures born on the lateral branches of a mature thallus where embryos are developed and shed (author's unpubli-

shed data). Senescence of the thallus ensues after embryos are shed, completing the annual life cycle. *Sargassum* embryos tend to sink quickly (Gaylord et al. 2002) and the vast majority likely settle within a few meters of the parent thallus (Deyscher and Norton 1982; Stiger and Payri 1999; Kendrick and Walker 1995). Clearing thalli in relatively small areas on the order of tens of square meters may therefore reduce colonization resulting from local dispersal. However, because colonization over longer distances is thought to occur via reproductively mature thalli that are dislodged and set adrift (Yatsuya 2008), any thalli removed must not be released. Asexual reproduction in *S. horneri* via fragmentation or regeneration from remnant tissue has not been studied, although it is known to occur in other fucoid species (McCook and Chapman 1992; Fletcher and Fletcher 1975). Information on the capacity of *S. horneri* to propagate asexually is needed to develop an effective management strategy for controlling its spread.

A new tool that has been developed to help control algal invasions is an underwater suction device. This type of device has been used on coral reefs in Oahu, Hawaii, to reduce densities of invasive algae (Conklin and Smith 2005), and a similar device was recently developed to aid in controlling seaweed invasions on rocky reefs in California. The device has been used to transport *S. horneri* removed from the ocean floor by divers to a platform at the sea surface, where the material can be collected for disposal on land (Meux 2013). However, the effectiveness of this approach in controlling *S. horneri* on temperate rocky reefs and how the efficiency of this method compares to non-mechanical techniques require further investigation.

To inform efforts to manage the spread and impact of *S. horneri*, we removed it from experimental areas and followed colonization and survivorship for one year to address three questions. First, how effective is local removal in controlling populations of *S. horneri*? Second, what is the capacity of the species to regenerate from remnant holdfasts? Third, how much effort is required to remove established populations with and without the aid of an underwater suction device?

Methods

This study was performed on the leeward side of Santa Catalina Island, California, USA on two nearby reefs (Howland Landing: 33.465°N; 118.522°W and Lion Head: 33.453°N; 118.502°W) at 6–8 m depth (Figure 1). We chose these locations because they are representative of the topography of reefs in the area, and have dense populations of *S. horneri*.

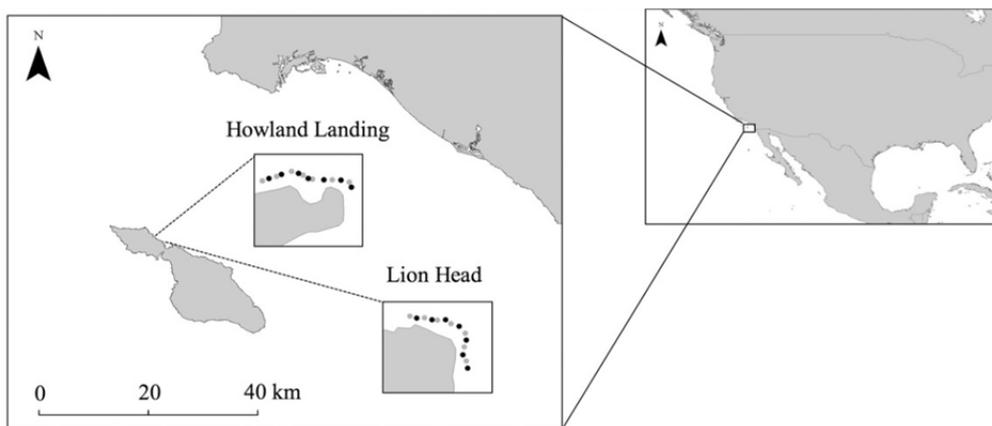


Figure 1. Map of Santa Catalina Island, located 27 km off the coast of southern California, USA. The insets show the distribution of 28 experimental plots spread across two sites: Howland Landing and Lion Head. Dark circles represent removal plots, and grey circles represent control plots.

Removal experiment

To evaluate the effectiveness of *S. horneri* extraction in reducing local populations, we performed a removal experiment and monitored colonization and survivorship of the next generation. We established twenty-eight 60 m² circular plots in areas where *S. horneri* was abundant and assigned plots alternately to either a removal or non-removal (i.e., control) treatment (Figure 1). Fourteen plots were located 15–20 m apart at each of the two study sites.

We extracted *S. horneri* from removal plots in the winter (February 2015) when individuals were at their largest size and lowest densities, but before the vast majority (i.e., 99%) of them became fertile so as to minimize the source of *S. horneri* propagules within the removal plots. Immediately prior to removal we counted the number of *S. horneri* adults (defined as > 5 cm tall) in sixteen 0.25 m² quadrats plot⁻¹ that were placed within each plot at 0, 1, 2 and 3 m from the edge along two perpendicular diameters. To prevent mature thalli from drifting away and starting distant populations, we captured all material removed and transported it to boats anchored at the surface. On deck, workers immediately transferred material into heavy-duty trash bags. We later emptied these bags at an upland location where we left the algae to decompose.

We removed all *S. horneri* from the substrate manually and employed one of two methods to transport it to the surface: mesh bags and lines, or an underwater suction device. The bag and line method involved divers placing *S. horneri* into weighed bags (Figure 2A). Once filled, buoyant bags were released from their weights and attached to lines hanging off

the side of the boat (Figure 2B) and a worker at the surface hauled them onboard. The suction device consisted of a mechanical water pump (Subaru PTX201D Robin Pump) with 7.6 cm-diameter input and output hoses that is operated on the deck of the boat (Figure 2C). Divers fed material into the hose at depth and it was transported to the surface by the movement of a diaphragm (Figure 2D). Regardless of the method used, most individuals were completely removed from the substrate (Figure 2E). However, the holdfasts of some individuals remained after their stipes were severed.

Removal plots were resampled immediately after clearing to confirm all thalli had been removed and to quantify the density of remnant holdfasts. In September 2015, we measured colonization by counting the number of juveniles (defined as < 5 cm tall) in all plots. In February 2016, one year after experimental removals, we counted the number of adults in each plot to assess the effects of removal on population density. Because *S. horneri* grows on rock and the percent cover of rock was consistently high but slightly variable (mean ± SE = 97.9 ± 0.19%) we adjusted estimates of density within each quadrat by the percent cover of rock in that quadrat. Hence *S. horneri* is reported as number m⁻² of rock rather than number m⁻² of sea floor.

We tested the effects of removal on colonization (i.e., juvenile density in September 2015) and population density (i.e., adult density in February 2016) in separate two-way hierarchical ANOVAs with treatment (removal versus control) as a fixed factor and site (Howland Landing versus Lion Head) as a random factor and plots nested within sites. We considered plots independent replicates of treatment effects in cases when the random effect of site was not significant.

Fate of individuals with severed stipes

To determine whether severing a *S. horneri* stipe near its base while leaving the holdfast intact is sufficient to prevent it from regenerating, we followed the fate of individuals after cutting their stipes in March 2015. We attached identifying markers to the reef adjacent to 80 holdfasts and revisited the marked individuals monthly for four months to record whether they remained attached to the substrate and, if so, whether they regenerated new tissue. We also collected observations of the remnant holdfasts in the plots we cleared. Although we were not able to follow these holdfasts individually, we looked for perennating *S. horneri* holdfasts when resampling the plots.

Efficiency of removal

We evaluated the efficiency of removal with and without the aid of the suction device (Figure 2) by quantifying the effort required for each method for a given quantity of *S. horneri* biomass. We did this by recording the removal method being used (i.e., suction device or bags and lines), time spent collecting, number of workers (i.e., scuba divers and surface support person) and amount of biomass removed for each dive. To estimate the biomass removed, we collected the algae into bags as soon as it was brought to the surface and weighed it to the nearest 0.5 kg using a hanging scale. In addition, we measured the rate of transport to the surface using the suction device across a range of stipe lengths to determine if size affected performance. We fed 30 pieces of several stipe lengths that are often naturally observed (30 cm, 60 cm, 100 cm and 150 cm) into the hose and recorded the time it took to bring them up to the surface.

Results

Removal experiment

The density of adult *S. horneri* prior to experimental removal in February 2015 was similar in removal and control plots ($F_{1,1} = 0.98$, $p = 0.504$) averaging 46.4 and 50.4 individuals m^{-2} , respectively (Figure 3A). Adult density differed significantly between the two sites ($F_{1,420} = 26.95$, $p < 0.001$) with density ~55% higher at Howland Landing. Quadrat sampling and visual surveys of entire plots verified that experimental clearing resulted in the removal of virtually all visible thalli in removal plots, but some holdfasts with severed stipes remained. The density of remnant holdfasts immediately after clearing was 46.1% of the initial adult population (mean \pm SE = $22.3 \pm 2.9 m^{-2}$).

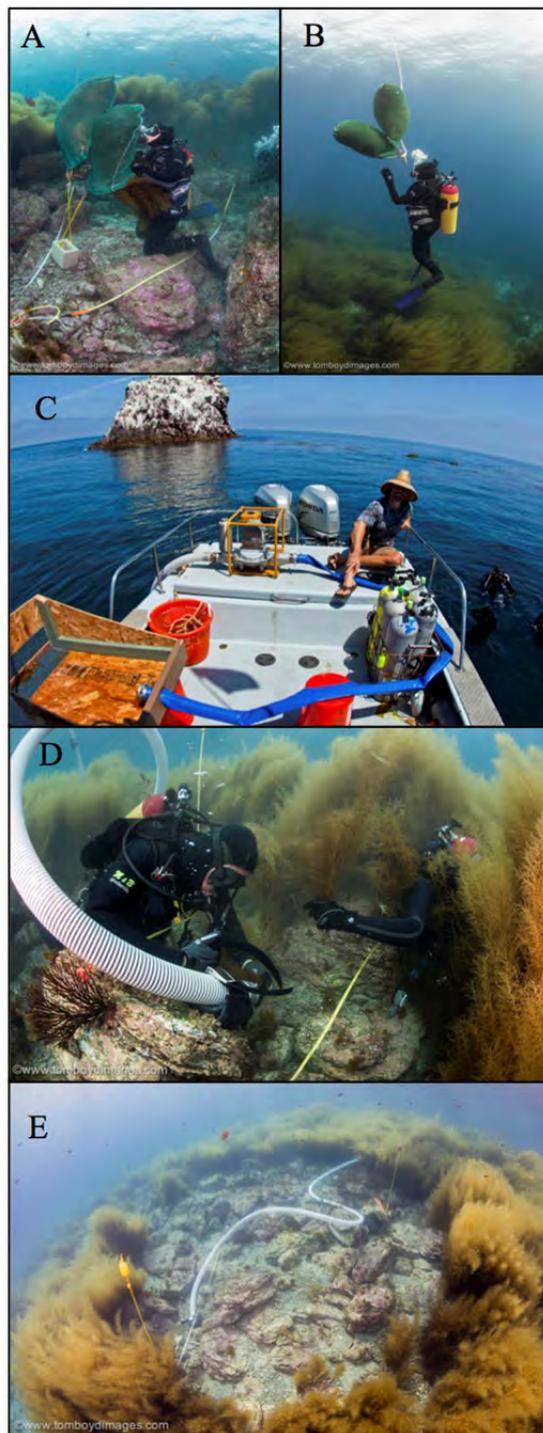


Figure 2. Two methods used to transport *Sargassum horneri* to the surface. Using the bag and line method, a diver fills bags anchored by a cinderblock (A), then clips bags to a line hanging from a boat anchored overhead (B). Using the suction device method, two divers work together to feed *S. horneri* into the hose (C), and a person at the surface collects the material from a sorting table after inspecting it for bycatch (D). After clearing using both methods, plots were left barren of *S. horneri* (E). Photo credits: Tom Boyd (A-B, D-E), Adam Obaza (C).

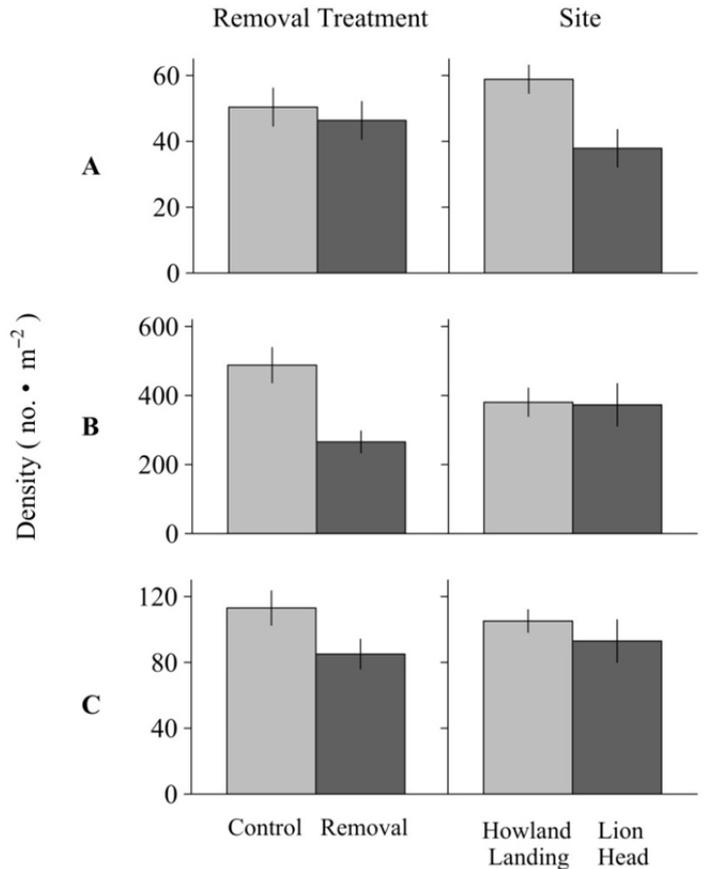


Figure 3. Results of removal experiment showing the average density \pm SE of *Sargassum horneri* (A) adults prior to their removal, (B) juveniles ~220 days after removal, and (C) adults ~366 days after removal (N = 14 plots).

Similarly high densities of recently colonized juveniles were observed in all plots in September 2015, ~7 months after clearing (Figure 3B; $F_{1,420} = 0.08$, $p = 0.775$). Removal had a significant effect on subsequent colonization ($F_{1,26} = 12.95$, $p = 0.001$) as juvenile density was 54% lower in removal plots compared to control plots. The effect of removing *S. horneri* on colonization by juveniles was similar at both sites (treatment \times site: $F_{1,1} = 0.236$, $p = 0.125$).

The reduced densities in removal versus control plots persisted but became less pronounced over time as juveniles grew into adults (Figure 3C). By February 2016, one year after clearing, adult densities averaged 25% lower in removal plots compared to control plots. However, overall adult densities were 83% higher in removal plots and 115% higher in control plots compared to February 2015 prior to removal (Figure 3A versus 3C).

Fate of individuals with severed stipes

Significant tag loss resulted in reduced and unequal sample sizes for estimating survivorship on the different sampling dates, which compromised our ability

to quantitatively evaluate the regenerative capacity of individuals with severed stipes. Nonetheless, the data that we collected indicate that *S. horneri* has little or no capacity for regenerating from remnant holdfasts as none of the individuals with severed stipes that remained tagged generated new tissue. Fifty-six of the 80 tags remained after 31 days and remnants of holdfasts were found for only 20 of these 56 individuals. Remnants of 10 of 14, 4 of 9 and 0 of 8 holdfasts remained after 54, 85 and 113 days, respectively (Figure 4). Furthermore, when we sampled the removal experiment in September 2015, we found no remnant holdfasts, which suggests they had all senesced and disappeared within seven months.

Efficiency of removal

The efficiency of removing *S. horneri* varied by the method used to transport it to the surface and the number of workers. Three workers using the bag and line transport method yielded the slowest average removal rate of $29 \text{ kg worker}^{-1} \text{ hr}^{-1}$, while the suction device method with three workers (two divers and one surface support person) yielded an average of 38 kg

worker⁻¹ hr⁻¹ (Figure 5). Limits on the amount of material that can be fed into the hose at any given time resulted in two divers being the optimum number to maximize the transport of algae to the surface. By contrast, the manual transportation method using bags and lines allowed for more divers to work efficiently in the same area. While the overall rate of removal using bags and lines increased with the number of workers, the maximum per capita efficiency was about 45 kg worker⁻¹ hr⁻¹ (Figure 5). The rate of transport using the suction device was highest at intermediate stipe lengths (~60 cm; Figure 6).

Discussion

Our results show that the experimental removal of *S. horneri* reduced the local population in the next generation by ~25% relative to control plots. However, despite this reduction, removing *S. horneri* did not lead to a decline in population density relative to the previous year as adult densities in both the removal and control plots were substantially greater in 2016 than in 2015 prior to removal. These results highlight some of the challenges associated with efforts to reduce established populations of *S. horneri* via removal. Moreover, they suggest that measurable success using removal techniques as a means of controlling *S. horneri* will likely require that removals be done over much larger areas to ensure an adequate reduction in propagule supply, which will be costly. The effect of removing *S. horneri* on its abundance in subsequent generations (as measured by the difference in *S. horneri* density between control and removal plots in the year following removal) was most apparent during the fall when the majority of individuals were juveniles, and became less pronounced in the winter when most were adults. The order of magnitude higher densities that we observed for juveniles compared to adults is consistent with self-thinning induced by intra-specific competition, which is common in large brown algae (Schiel and Choat 1980; Schiel 1985; Dean et al. 1989; Reed 1990). The dampened effect of removal between the juvenile and adult phases suggests removal accelerated the self-thinning process.

The increased density of *S. horneri* that we observed in our removal and control plots may have been due to the unusually warm water resulting from the 2015–16 El Niño. The native canopy-forming kelps *Macrocystis pyrifera* and *Eisenia arboria* commonly found on shallow reefs of Santa Catalina Island thrive in cool, nutrient-rich water. These species largely disappeared from the leeward side of the island during our study while *S. horneri* flourished, as did other species with warm water affinities (e.g.,

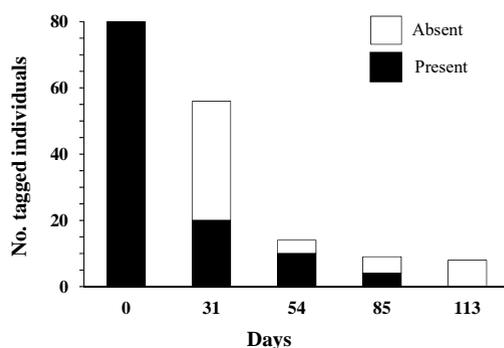


Figure 4. Survivorship of *Sargassum horneri* with severed stipes. Solid bars represent the number of thalli with remnant tissue remaining. Open bars represent the number of tags relocated where holdfasts had senesced. Combined, the bars represent the total number of tags found, and the number of individuals upon which survivorship was based for each sampling period.

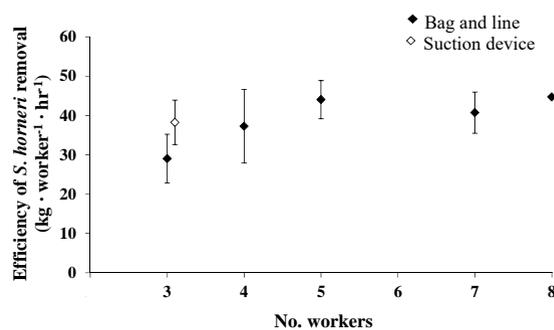


Figure 5. *Sargassum horneri* average removal rate (kg wet biomass worker⁻¹ hr⁻¹) ± SE reported for each removal method. Replication varies by the number of dives with each given number of workers using each method. N = 15 dives with 3 workers using the suction device, and N = 6, 4, 6, 6 and 1 dives with 3, 4, 5, 6, and 8 workers using the bag and line removal method, respectively.

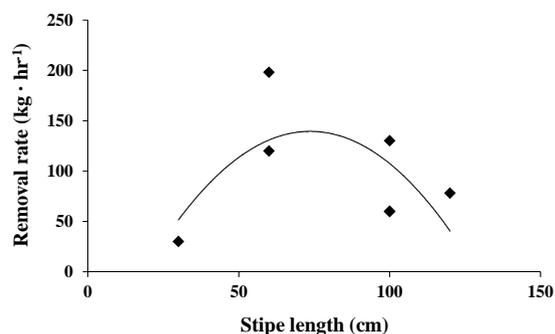


Figure 6. The rate (kg wet biomass hr⁻¹) at which stipes of *Sargassum horneri* were transported by workers using the suction device as a function of stipe length.

Zonaria farlowii, *Dictyota* spp. and *Dictyopteris undulata*). Evidence from the nonindigenous congener *Sargassum muticum*, which became abundant at Santa Catalina Island for several years following the El Niño of 1976 (Coyer 1979), suggests that *Sargassum* spp. with warm-water affinities decline once cooler waters return and large, perennial native kelps become re-established (Ambrose and Nelson 1982) Whether *S. horneri* declines over time remains to be seen, but if the warming observed in 2015–16 is a preview of future conditions, then tropicalization of an algal assemblage that favors *S. horneri* may be the norm.

The efficacy of removing invasive algae could be strengthened by selecting conditions under which native species can exert biotic control on the remaining population, or even by enhancing these controls. Researchers in Hawaii attributed their success in controlling invasive *Eucheuma* spp. and *Kappaphycus* spp. on patch reefs to introducing urchins after performing removals (Conklin and Smith 2005). Once divers reduced the algae below a critical threshold, the herbivores were able to prevent it from growing back. While this is an effective strategy on coral reefs where indiscriminant grazing is acceptable, introducing generalist herbivores is not a viable strategy to control invasive algae on temperate rocky reefs, which are often dominated by a diversity of macroalgae.

An alternative strategy to enhance biological resistance to the regrowth of invasive algae on rocky reefs is to perform removals under conditions favoring the colonization of native species of macroalgae and sessile invertebrates that compete for space and/or light. Resource competition is recognized as an important mechanism structuring communities (MacArthur 1970; Levine and D'Antonio 1999; Tilman 2004), and competition for space and light plays a key role in organizing the benthic community on rocky reefs (Miller and Etter 2008; Arkema et al. 2009). The invasion of a community is thought to be inversely related to species richness due to the enhanced ability of resident species to preempt resources (Elton 1958), and manipulative field experiments have shown that decreasing native diversity increases limited resources and the abundance and survivorship of non-native species in subtidal benthic communities. For example, Stachowicz et al. (2002) found that experimentally increasing sessile invertebrate species richness decreased both the availability of space, the limiting resource in this system, and the abundance of non-indigenous ascidians by buffering against temporal fluctuations in the cover of individual native species. Furthermore, multiple resources might be limiting the success of a non-native species throughout its life

cycle, and higher functional diversity may allow a community to preempt multiple resources more effectively. A native algal community with crustose and turfing algae preempting space and understory and canopy-forming algae preempting light sequentially suppressed the recruitment and survivorship of the nonindigenous seaweed *Sargassum muticum* (Britton-Simmons 2006). The preemption of limited resources by native species of algae and invertebrates in areas where *S. horneri* has been removed could likewise limit *S. horneri*'s ability to re-establish.

Another important factor to consider when controlling invasive algae through removal is the mechanisms by which it recolonizes cleared areas. Many species of invasive algae have the ability to regenerate from miniscule amounts of tissue (e.g., Fletcher and Fletcher 1975; McCook and Chapman 1992) and this characteristic presents a challenge when considering control via removal (Smith 2015). We found no evidence that *S. horneri* has the capacity to regenerate from remnant holdfasts. This suggests that severing stipes, which is far less time consuming than carefully scraping all tissue from the reef, would be an effective and efficient means of reducing *S. horneri* abundance.

Whether an underwater suction device, such as the one tested in this study, would be the preferred method for invasive algae control depends on staff and budget limitations. The bag and lines method is optimal when many workers (i.e., > two divers and one surface support worker) are available. It also requires minimal training and material costs, and so may be preferred with constrictive budgets. A suction device minimizes surface support effort, particularly associated with lifting heavy bags, and offers increased efficiency with a limited number of workers (< 3 divers). Drawbacks of using a suction device include increased start-up costs, logistical challenges associated with equipment transportation and maintenance, and limitations on working depths. In addition, significant time can be spent troubleshooting, such as identifying appropriately sized pieces of algae to reduce the frequency of clogs. However, removal efficiency is likely to improve as operators become more familiar with the device and alter equipment to better suit the target species. Workers in Hawaii designed several models using different kinds of pumps until they identified the optimal configuration for their target species (Conklin personal communication). Therefore, long-term efficiency gains may make a suction device preferable if an extended control effort is expected.

Eradicating problematic species from their novel habitats is most likely to be successful if attempted before they become widely established (Myers et al.

2000; Bax et al. 2003; Hulme 2006). *Caulerpa taxifolia*, a green alga native to the Indo-Pacific region, was introduced in two protected embayments in southern California in 2000 and a rapid response effort successfully eradicated this species (Anderson 2005). The appearance of *S. horneri* off the open coast of North America is the first record of this species outside of its native range in Asia (Marks et al. 2015). While the aggressive spread of *S. horneri* throughout southern California and Baja California, Mexico makes total eradication in this region highly unlikely, *S. horneri* has the potential to spread to other temperate reefs around the globe. Knowledge about the life history and effective methods for controlling *S. horneri* abundance will prepare resource managers in other regions to eradicate new populations before they become widely established. Our study is one of the first on targeted control of an invasive species on the open coast of California. Development of a removal protocol along with awareness generated by this work will better prepare resource managers and the general public for future invasions of *S. horneri* in other regions.

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Supplemental material

The following supplementary material is available for this article:

Table S1. Recruitment and survivorship of *Sargassum horneri* following removal

Table S2. Survivorship of *Sargassum horneri* with severed stipes

Table S3. Per capita removal rate of *Sargassum horneri*

Table S4. Transport rate of *Sargassum horneri* using suction device

This material is available online for download from the Long Term Ecological Research Network Data Portal, <http://dx.doi.org/10.6073/pasta/a812d149f4d6e9cd5662d4c44eaedd2>

Article

Niche Complementarity and Resistance to Grazing Promote the Invasion Success of *Sargassum horneri* in North America

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Abstract: Invasive species are a growing threat to conservation in marine ecosystems, yet we lack a predictive understanding of ecological factors that influence the invasiveness of exotic marine species. We used surveys and manipulative experiments to investigate how an exotic seaweed, *Sargassum horneri*, interacts with native macroalgae and herbivores off the coast of California. We asked whether the invasion (i.e., the process by which an exotic species exhibits rapid population growth and spread in the novel environment) of *S. horneri* is influenced by three mechanisms known to affect the invasion of exotic plants on land: competition, niche complementarity and herbivory. We found that the removal of *S. horneri* over 3.5 years from experimental plots had little effect on the biomass or taxonomic richness of the native algal community. Differences between removal treatments were apparent only in spring at the end of the experiment when *S. horneri* biomass was substantially higher than in previous sampling periods. Surveys across a depth range of 0–30 m revealed inverse patterns in the biomass of *S. horneri* and native subcanopy-forming macroalgae, with *S. horneri* peaking at intermediate depths (5–20 m) while the aggregated biomass of native species was greatest at shallow (<5 m) and deeper (>20 m) depths. The biomass of *S. horneri* and native algae also displayed different seasonal trends, and removal of *S. horneri* from experimental plots indicated the seasonality of native algae was largely unaffected by fluctuations in *S. horneri*. Results from grazing assays and surveys showed that native herbivores favor native kelp over *Sargassum* as a food source, suggesting that reduced palatability may help promote the invasion of *S. horneri*. The complementary life histories of *S. horneri* and native algae suggest that competition between them is generally weak, and that niche complementarity and resistance to grazing are more important in promoting the invasion success of *S. horneri*.

Keywords: introduced species; biological invasion; macroalgae; canopy shading; competition; herbivory; *Sargassum filicinum*

1. Introduction

Marine ecosystems are increasingly threatened by invasive species as global trade expands and human-mediated introductions via commercial shipping occur at escalating rates [1–5]. Developing a predictive understanding of factors influencing the success of marine invasive species has clear implications for managing their spread and impacts. Yet relative to terrestrial systems, little is known about the ecological processes that influence marine invasions [6,7]. In terrestrial ecosystems, once an introduced species becomes established, biotic interactions with native species can play a major role in limiting population growth, spread and ecological impacts [8–11]. These interactions can either

promote or inhibit “invasion”, here defined as the process by which an exotic species exhibits rapid population growth and spread in the novel environment [12].

In terrestrial and freshwater plants, biotic interactions such as competition with natives and herbivory can affect invasion success [9,13]. Competition for limited resources among native and invasive species is expected to be most intense when they have similar life histories and resource requirements [14–16]; invasion is promoted when exotic plants employ resource acquisition strategies superior to native competitors, reducing their abundance or diversity [17]. Invasion success can also be promoted when exotic species have functional traits or resource requirements that differ from the native biota, which allows them to take advantage of underutilized resources in space and time [16,18–21]. Such niche complementarity can facilitate invasions by allowing exotics to avoid interacting with natives that have superior competitive abilities. Like native plants, native consumers can promote or hinder invasion depending on their dietary preference. For example, herbivores that prefer exotic plants to natives can inhibit invasion, while those that prefer native plants can facilitate invasion by reducing the strength of competition between exotic and native plants [10,22,23]. Studies aimed at determining the mechanisms affecting the invasiveness of exotic marine macrophytes are needed to derive meaningful generalizations about the role of biotic interactions in influencing the invasibility of a wide range of ecosystems.

The Asian brown alga *Sargassum horneri* (Turner) C. Agardh, 1820 (Fucales) was first detected in California in 2003 [24] and has since spread throughout southern California and Baja California, Mexico [25]. Several life history characteristics of *S. horneri* are typical of “weedy” invaders with r-selected traits including broad habitat requirements and high fecundity with >60% of its biomass dedicated to reproductive tissue at its peak fertility [26,27]. It has highly localized propagule dispersal, as well as the ability to disperse long distances via the dislodgement and drifting of buoyant fertile adults [27]. The biomass of *S. horneri* is strongly seasonal: juveniles prevalent in the summer exhibit rapid growth to several meters in height during the winter, and reproduction and biomass peak in the spring [27]. *S. horneri* has the potential to compete with native algae by reducing the amount of light reaching algae growing beneath its canopy. Throughout the invaded range, *S. horneri* has become a dominant macroalga in some areas, but remains rare in others [7,25]. However, it is unclear whether this dominance results from competitive displacement of native species or opportunistic occupation of an underutilized niche.

During its reproductive phase, *S. horneri* can form dense canopies that shade the bottom, and canopy shading by invasive algae has been shown to cause the decline or exclusion of native seaweeds [28–30]. However, it has been hypothesized that the invasion of *S. horneri* is suppressed in areas dominated by native algae [7], suggesting that niche complementarity rather than competitive superiority accounts for its rapid spread in North America. Detailed information on patterns of distribution of *S. horneri* and native algae across space (e.g., depths) and through time (e.g., seasons) can provide valuable insight into the relative importance of competition versus niche complementarity in accounting for the invasion success of *S. horneri*.

The effects of herbivores in structuring temperate marine communities are well documented [31–33], but less is known about their potential role in influencing invasions. Exotic seaweeds with traits that deter herbivory (i.e., structural or chemical defenses) can gain an advantage over native competitors in areas with high grazing pressure. Such may be the case for *S. horneri* as it is in the order Fucales, which is known for having high levels of phenolic compounds that deter grazing [34–36]. Thus, preferential consumption of less defended native algae such as laminarian kelps [36,37] could facilitate the spread of *S. horneri* by weakening competition with other more palatable native algae.

The purpose of this study was to determine the degree to which competition, niche complementarity and herbivory account for patterns of abundance of *S. horneri* in an area where it has become established. To do this, we documented patterns of co-occurrence between *S. horneri* and native algae spatially across a depth gradient and temporally over multiple years in experimental plots with *S. horneri* removed or left intact to evaluate niche complementarity and competition as mechanisms contributing to the

invasiveness of *S. horneri*. If invasion by *S. horneri* results from its ability to outcompete native algae, then we expected the biomass and taxonomic richness of native algae to increase in areas where we experimentally removed *S. horneri*. Alternatively, if the invasion success of *S. horneri* relies on its ability to occupy underutilized resources, then we expected to see little change in the native algal assemblage in response to *S. horneri* removal. We also performed a field experiment involving the major herbivores to examine their grazing preferences for *S. horneri* versus other algae. Using a combination of feeding assays and distributional surveys, we tested the hypothesis that herbivores facilitate *S. horneri* by preferentially consuming native algae.

2. Materials and Methods

2.1. Study System

Field experiments and surveys were conducted on rocky reefs on the leeward side of Santa Catalina Island, located 35 km offshore of Los Angeles, CA, USA. Study reefs consisted of bedrock, boulders and cobble distributed along a moderate slope that transitioned to sand at depths of about 30 m. The reefs were dominated by native macroalgae and the invasive *Sargassum horneri*. Native macroalgae included the canopy-forming giant kelp *Macrocystis pyrifera*, subcanopy-forming species of kelp (e.g., *Eisenia arborea* and *Agarum fimbriatum*) and furoid algae (e.g., *Sargassum palmeri*, *Stephanocystis neglecta* and *Halidrys diocia*), and understory-forming foliose and calcified algae. Sessile invertebrates occupied only about 3% of the reef surface. *S. horneri* has become one of the most common macrophytes on shallow reefs at Santa Catalina Island since its introduction in 2006.

The primary grazers at Santa Catalina Island include sea urchins and herbivorous snails. *Centrostephanus coronatus*, the most abundant species of urchin, takes refuge in crevices and forages within <1 m from its shelter during the night before returning to the same location before sunrise [38]. This behavior leads to the formation of urchin “halos” where they commonly graze down algae within small home ranges.

2.2. Competition

To test the effects of *Sargassum horneri* on the abundance and taxonomic richness of native algae, we compared the native algal assemblages in experimental plots from which *S. horneri* was continually removed (hereafter referred to as S−) with those in unmanipulated control plots with *S. horneri* left intact (S+) over 3.5 years. We also measured the reduction in the amount of light permeating through its canopy as a potential mechanism of competition. This experiment was conducted at Isthmus Reef (33.4476° N, 118.4898° W) at 6 m depth, within the range where *S. horneri* is most abundant. Twenty-four 1 m² plots separated by a distance of at least 2 m were established on areas of reef comprised of >90% rock and with a high density (i.e., at least 30 individuals) of *S. horneri*. *S. horneri* was removed from 12 randomly assigned plots (S−) beginning in spring 2014 and every 6 to 12 weeks thereafter until summer 2017. S− plots had a 30 cm wide buffer zone around the perimeter where *S. horneri* was removed to minimize potential edge effects such as shading by individuals outside of the plot. Removal entailed divers using knives to pry all *S. horneri* holdfasts off the substrate, minimizing disturbance to the other biota within the plot as much as possible. Since competitive interactions may vary with time and among seasons, we sampled the algal communities in all S+ and S− plots just prior to the initial removal of *S. horneri* in spring 2014 and quarterly thereafter (i.e., summer, autumn, winter and spring) over three consecutive growing seasons (2014–2015, 2015–2016 and 2016–2017).

Algae were identified to the lowest taxonomic level possible, which in most cases was species (Table S1), and measurements of all understory and subcanopy-forming algae were taken in order to estimate the damp biomass of algae in each plot. The abundance of low-lying understory algae was measured as percent cover using a uniform point contact (UPC) method that involved recording the presence and identity of all algae intersecting 49 points distributed in a grid within each 1 m² plot. Percent cover was determined as the fraction of points a taxon intersected × 100. Although multiple

organisms may intersect a single point if they overlay one another, a taxon was only recorded once at a given point even if it intersected that point multiple times. Using this technique, the percent cover of all taxa combined in a plot can exceed 100%, but the percent cover of any individual species or morphological group cannot. This sampling resolution was sufficient to detect species covering at least 2% of the area in a quadrat. If a species was present in the plot but not recorded at one of the 49 points, then it was assigned a percent cover value of 0.5%. Since percent cover does not necessarily scale with biomass for larger subcanopy-forming algae, we recorded the density and the average size of these taxa. Damp biomass was estimated from density and size data of subcanopy algae and percent cover data of understory algae using taxon-specific relationships obtained from the literature [27,39–41] or developed specifically for this project (Table S2).

All but two species of algae recorded in the study plots were native to the region; the non-native *Sargassum muticum* and *Codium fragile* occurred in low abundance. Both of these species and *S. horneri* were excluded from analyses to test specifically for the effects of *S. horneri* on the native algal assemblages [42]. The surface canopy-forming giant kelp, *Macrocystis pyrifera*, was present at the beginning of the experiment, but it declined quickly during a warming trend and disappeared by December 2014 for the duration of the study. Consequently, its presence did not factor into our analyses.

The effects of *S. horneri* removal on the taxonomic richness and aggregate biomass of native algae were evaluated using linear mixed effects models [43]. Taxonomic richness was calculated as the number of unique native algal taxa within each plot, and aggregate biomass was calculated as the summed damp biomass of all native algae within each plot. Since we hypothesized that treatment effects may differ among seasons and develop over time, we included season, treatment (S+ or S−) and days since the start of the experiment (elapsed time) as main effects in the model. To account for variation associated with resampling individual plots, we included plot and the summed damp biomass of native algae within each plot at the start of the experiment prior to the first removal of *S. horneri* as random effects. Full models with the main effects in question (i.e., season, removal treatment, elapsed time and the interactions of time–removal treatment and season–removal treatment) were compared against null or full models without the effects in question using likelihood ratio tests with chi-square test statistics to select the best fit based on the Akaike Information Criterion (AIC). Model assumptions of normality and homoscedasticity were validated through visual inspection of the residuals, and biomass data were square-root transformed to meet model assumptions. To identify which time periods contributed to the time-by-removal treatment interaction, we used Tukey’s Honest Significant Difference (HSD) post hoc analysis to compare the means of S+ and S− treatments for each sampling period.

Differences in the composition of the algal community between S+ and S− plots were tested using non-metric multi-dimensional scaling (nMDS) and analysis of similarities (ANOSIM). We compared the mean biomass of each taxon in S+ and S− plots in spring and summer 2017, during and after the sampling period when *S. horneri* removal had a significant effect. We used an unrestricted permutation of raw data (999 permutations) on Bray–Curtis similarity matrices with square-root transformation applied. A similarity percentage (SIMPER) analysis was used to determine the taxa that contributed most to dissimilarity between S+ and S− plots.

To determine the amount of shading caused by the *S. horneri* canopy we calculated the percent transmission of photosynthetically active radiation (PAR, 400–700 nm) during the spring sampling periods in S− and S+ plots. Light was measured using a handheld spherical quantum sensor (LI-COR Model LI-192) oriented vertically in the center of each plot 30 cm above the bottom. Ten readings of Photosynthetic Photon Flux Density (PPFD in $\mu\text{mol m}^{-2} \text{s}^{-1}$) were taken in each plot and averaged. Percent transmission was calculated from the average of 10 PPFD readings taken at the surface before and after the dive as:

$$\% \text{ transmission PAR} = \left[1 - \frac{\text{PAR}_{scf} - \text{PAR}_{plot}}{\text{PAR}_{scf}} \right] \times 100$$

We assessed how percent transmission of PAR was affected by *S. horneri* canopy biomass in S+ plots during spring using linear regression. We also tested the hypothesis that the removal of *S. horneri* increases PAR reaching the bottom compared to unmanipulated plots during spring following the initial removal of *S. horneri* using a repeated-measures ANOVA with removal treatment as a fixed factor, and plot and year as random factors. We used one-tailed *t*-tests to determine how the years differed from each other with respect to light transmission because we had an a priori expectation that light would be lower in S+ plots than S− plots. Percent transmission light data were arcsin-transformed prior to analyses to meet the assumptions of ANOVA.

2.3. Complementarity

We examined seasonal patterns of biomass of *Sargassum horneri* and native algae in the experimental plots described above to test their degree of temporal complementarity. Comparisons of native algae and *S. horneri* in S+ plots were used to determine whether the seasonality in biomass differed between the two, while comparisons of native algae in S+ and S− plots were used to determine whether seasonal fluctuations in biomass of native algae occurred independent of *S. horneri* abundance.

We examined the degree of spatial complementarity between *S. horneri* and native algae by comparing their biomass across the depth range within which most species of brown algae at Santa Catalina Island occur (0–30 m). Scuba divers counted the number of recruit (defined as <5 cm tall) and adult (defined as >5 cm tall) *S. horneri* and native species of subcanopy-forming macroalgae within 1 m² quadrats placed every 5 m along transects at four sites that ran perpendicular to shore from the intertidal to 30 m depth or where the reef transitioned to sand, whichever came first. Density data were converted to units of damp biomass using the method described above (see 2.2 Competition). Since these algae grow only on hard bottom substrate, we visually estimated the percent cover of rock within each quadrat and standardized density estimates to m^{−2} hard bottom. We performed these surveys in April of 2016, the time of year when the biomass of *S. horneri* reaches its peak [27]. Although smaller native understory species may also compete with *S. horneri*, limits on bottom time prevented us from sampling them.

Measured depths were adjusted relative to the Mean Lower Low Water (MLLW) and quadrats were binned into depth intervals of 5 m. Between one and three quadrats were sampled within each depth interval at each site, depending on the grade of the reef. The aggregate biomass of native algae within a quadrat was calculated as the sum of the biomass of the juvenile and adult stages of all native species measured. A two-way ANOVA was used to test whether the biomass of *S. horneri* and the aggregate biomass of native algae varied by depth interval and taxa.

2.4. Herbivory

We performed grazing assays and surveys of benthic algae within and adjacent to urchin halos to assess whether the palatability of *S. horneri* differed from that of other algae. In September 2016, replicate arrays consisting of *Sargassum horneri*, its native and introduced congeners *S. palmeri* and *S. muticum* and the native kelps *Macrocystis pyrifera* and *Eisenia arborea* were deployed at Isthmus Reef for periods of 48 h. Arrays were either exposed to grazing by urchins and snails or placed inside cages nearby that were designed to exclude these grazers. Cages were constructed from 1 cm-gauge plastic mesh and were cylindrical in shape (1 m in height and 0.5 m in diameter) with mesh covering the top. Cages were open at the bottom and a 1 m-wide weighted skirt secured them to the reef and prevented grazers >1 cm from entering. All urchins and snails were removed from the cages at the beginning of each assay.

During each of the four deployments, 15 arrays containing one sample of each of the five target species of algae were placed in urchin halos while another 15 were placed inside cages. Urchin halos were defined as sections of the reef adjacent to a small ledge where >10 urchins were found and grazing activity was apparent from a lack of algae growing within a 30 cm radius. Some herbivorous snails were also present in the halos, including *Tegula eiseni*, *Tegula aureotincta*, *Megastrea undosa* and *Norrisia norrisii*. Cages were left in the same location for the duration of the experiment, but we selected

unique halos for each deployment so that herbivores would be naïve to the arrays. In the day preceding each deployment, we collected and weighed similarly sized blades or thalli of the five target species. Damp weights were quantified prior to deployment and immediately after collection by spin-drying samples for 10 s before weighing them. Three repeat measurements of each sample were taken by re-hydrating the sample and repeating the drying and weighing process. The average of three replicate measurements for each sample was used to optimize our ability to detect small changes in tissue loss.

Herbivore preference was assessed by comparing algal weights measured before and after each deployment in the exposed versus caged arrays. We calculated the percent of biomass lost as:

$$\% \Delta = \left[\frac{G_{\text{final}} - G_{\text{initial}}}{G_{\text{initial}}} \right] \times 100$$

where G_{initial} and G_{final} represent the mean of the three replicate weights measured for each sample before and after deployment respectively. For each deployment, exposed and caged arrays were randomly paired and the biomass of each species of algae lost due to grazing was calculated as the difference in the change in biomass between paired arrays. One-way ANOVA was used to evaluate whether the biomass lost due to grazing differed by species, and post hoc contrasts were tested for significance with a Tukey HSD test to determine which species were preferentially consumed. Model assumptions of normality and homoscedasticity were validated through visual inspection of the residuals.

To provide a more time-integrated assessment of the feeding preferences of grazers, we tested whether the relative abundance of *S. horneri* differed from that of native algae in heavily grazed areas during the final deployment. We did this by measuring the percent cover of all subcanopy and understory algae in 1 m² quadrats placed adjacent to the 15 urchin halos and at 15 nearby reference locations with high algal cover. Percent cover was assessed using the uniform point contact sampling method described above (see 2.2 Competition). We standardized estimates of cover for individual algal taxa to the total cover of subcanopy and understory algae within each quadrat to compare the relative algal composition adjacent to and away from halos. We ignored encrusting algae and unoccupied space in order to focus on the differences between the foliose algal species that are likely to be consumed by the grazers. Algae were identified to the lowest taxonomic level possible, and were analyzed in the following groups: *S. horneri*, *S. palmeri* and other native algae (Table S3). We used a two-way ANOVA to test whether the cover of these taxonomic groups differed adjacent to and away from urchin halos, and Tukey HSD post hoc contrasts were used to determine how the taxonomic groups differed from one another. Standardized percent cover data were arcsin-transformed prior to analyses to meet the assumptions of ANOVA.

2.5. Software Used for Statistical Analysis

All univariate statistical models and tests were completed using RStudio (version 1.1.414) for R Statistical Computing Package [44]. Linear mixed models were fit using the lme4 package [45], and post hoc comparisons were performed using the multcomp library [46]. All multivariate analyses were conducted using PRIMER v7.0 [47] and PERMANOVA+ for PRIMER [48].

3. Results

3.1. Competition

The aggregated biomass and taxonomic richness of native algae varied significantly by season (Table 1). Biomass peaked during summer and autumn, declined by winter and remained low into spring (Figure 1a), while richness also peaked in summer and declined slightly through spring (Figure 1b). The effects of experimentally removing *Sargassum horneri* on the biomass and species richness of native algae were dependent on season (see season × removal interactions in Table 1).

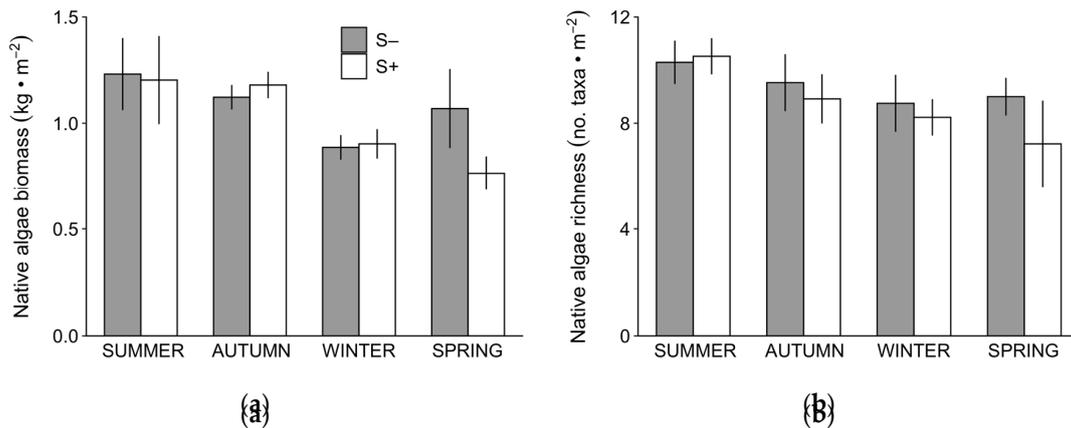


Figure 1. Mean (\pm SE) biomass (a) and taxonomic richness (b) of all native algae measured in *Sargassum horneri* removal (S-, grey bars) and non-removal (S+, white bars) plots. N = 4 years for 3 summer, and 3 years for autumn, winter and spring.

Although there was a significant interaction between season and removal for both biomass and species richness, post hoc tests revealed no particular season as driving the difference ($p < 0.05$ for all comparisons). Closer examination of the data revealed that the effects of *S. horneri* removal varied dramatically with days since the start of the experiment (Figure 2) as post hoc testing showed a significant difference in algal biomass between treatments in spring 2017 only, approximately 1100 days since the start of the experiment (Tukey's HSD, $p = 0.02$ indicated by * in Figure 2a; all other periods $p > 0.05$). This difference was driven by a bloom in native algae in S- plots that coincided with a dramatic increase in the biomass of *S. horneri* in S+ plots (Figure 2a). The biomass of native algae in S- and S+ plots began to converge again by summer 2017 when *S. horneri* biomass declined. The taxonomic richness of native algae decreased over the course of the study (Figure 2b), independent of the removal of *S. horneri* (Table 1b).

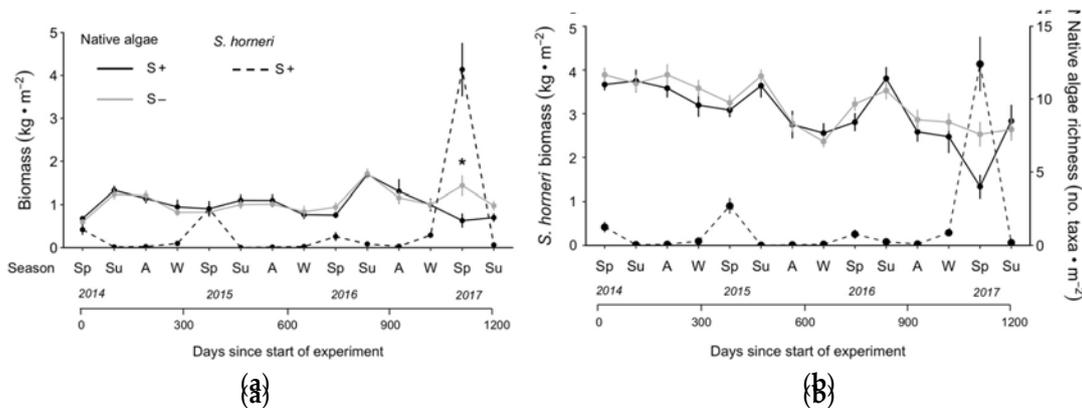


Figure 2. Mean (\pm SE) biomass (a) and taxonomic richness (b) of native algae in *Sargassum horneri* removal (S-, grey solid line) and non-removal (S+, black solid line) plots, overlaid by biomass of *S. horneri* in non-removal plots (S-, dashed line). Asterisk indicates sampling period where multiple comparisons tests (Tukey's HSD, $p < 0.05$) indicated a significant difference between treatments. N = 12 plots per sampling period. First data points in each series are from the pre-removal census.

Table 1. Results from likelihood ratio tests and model selection for determining the influence of the percent of photosynthetically active radiation (PAR) reaching the bottom in spring was inversely related to the biomass of *S. horneri* in S- plots when examined across all four years (Figure 3a; $R^2 = 0.88$). Explanatory variables included were: Days since the start of the experiment (Days); *S. horneri* removal treatment (Removal); and season of the sampling period (Season). Individual variables were tested against the null model and interactions were tested against additive models with the same ANOVA, $F_{1,22} = 25.2, p < 0.0001$. Post hoc tests revealed that *S. horneri* removal significantly increased PAR in each year: 2015: $t = 3.00, df = 22, p = 0.003$; 2016: $t = 1.78, df = 22, p < 0.04$; 2017: $t = 7.84, df = 22, p < 0.001$, especially in 2017 when the biomass of *S. horneri* in S+ plots was greatest.

parameters. Models were ranked according to Akaike Information Criterion (AIC) selection with lower AIC values indicating a better fit of the data. Significance was based on chi-square test statistics.

Diversity 2019, 11, x. Statistically significant *p*-values are in bold text.

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Variables	Model	df	AIC	χ^2	Chi df	<i>p</i> (< χ^2)
a. Biomass of native algae¹	Null	4	2140.4			
Individual parameters	Days	5	2106.3	87.83	1	<0.001
	Removal	5	2139.9	0.52	1	0.472
	Season	7	2141.9	48.51	3	<0.001
Interactions	Days + Removal	6	2106.3	40.22	3	<0.001
	Days + Season	7	2106.3	40.22	3	<0.001
	Days + Removal + Season	8	2107.7			
	Season × Removal	11	2104.7	9.02	3	0.029
b. Richness of native algae¹	Null	4	1489.7			
Individual parameters	Days	5	1403.8	87.83	1	<0.001
	Removal	5	1491.6	0.04	1	0.842
	Season	7	1447.1	48.54	3	<0.001
Interactions	Days + Removal	6	1405.8	0.72	1	0.397
	Days + Season	7	1407.1	11.15	1	<0.001
	Days × Removal	7	1432.2			
	Season + Removal	8	2107.7			
	Season + Season	8	1449.1			
	Season × Removal	11	1446.6	8.54	3	0.036

¹ data square-root transformed. The percent of photosynthetically active radiation (PAR) reaching the bottom in spring was inversely related to the biomass of *S. horneri* in S+ plots, when examined across all four years (Figure 3a; $R^2 = 0.33$, $F_{1,46} = 24.03$, $p < 0.001$). This reduction in light can be attributed to the development of the *S. horneri* canopy, as evidenced by the significant effect of *S. horneri* removal on PAR (Figure 3b; ANOVA, $F_{1,22} = 25.2$, $p < 0.0001$). Post-hoc tests revealed that *S. horneri* removal significantly increased PAR in each year (2015: $t = 3.00$, $df = 22$, $p = 0.003$; 2016: $t = 1.78$, $df = 22$, $p = 0.04$; 2017: $t = 7.84$, $df = 22$, $p < 0.001$), especially in 2017 when the biomass of *S. horneri* in S+ plots was greatest.

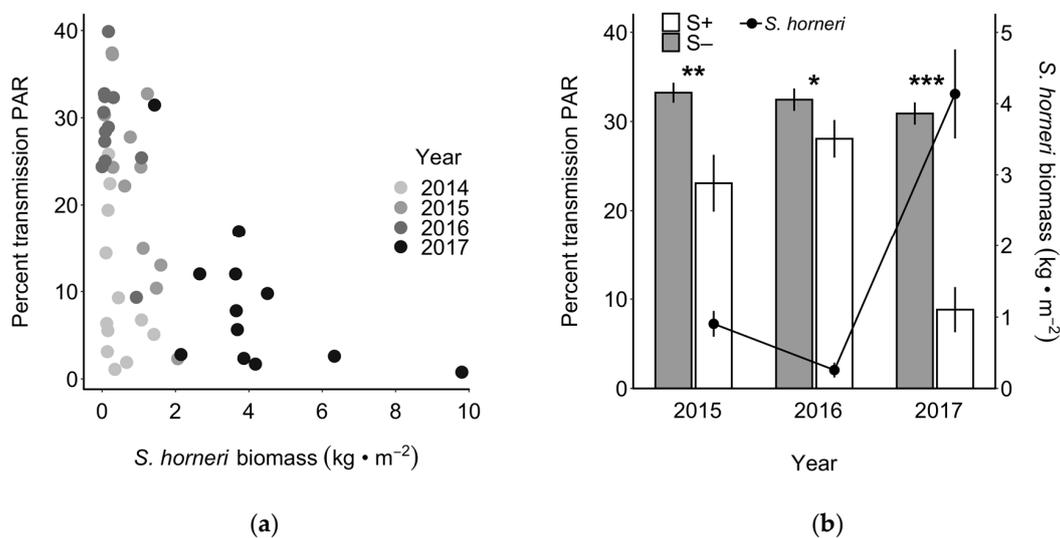


Figure 3. Percent transmission of PAR (i.e., percent of surface light reaching the bottom) of photosynthetically active radiation (PAR; 400–700 nm) related to *Sargassum horneri* biomass during spring. (a) Percent transmission of PAR related to *S. horneri* biomass in non-removal (S+) plots only, with each sampling year indicated by a different shade. (b) Effect of *S. horneri* removal on PAR. The left y-axis shows percent transmission of PAR (mean ± SE) in *S. horneri*-removal (S-; grey bars) and non-removal (S+; white bars) plots, and the right y-axis shows damp biomass of *S. horneri* (± SE) in non-removal (S+) plots when light measurements were taken. Asterisks indicate sample dates where *t*-tests indicated significant differences between treatments (*, **, ***: $p < 0.05$, 0.01, and 0.001, respectively).

year indicated by a different shade. (b) Effect of *S. horneri* removal on PAR. The left y-axis shows percent transmission of PAR (mean \pm SE) in *S. horneri*-removal (S-; grey bars) and non-removal (S+; white bars) plots, and the right y-axis shows damp biomass of *S. horneri* (\pm SE) in non-removal (S+) plots when light measurements were taken. Asterisks indicate sample dates where *t*-tests indicated significant differences between treatments (*, **, ***: $p < 0.05, 0.01, \text{ and } 0.001$, respectively).

Since *S. horneri* manipulation had no significant effect on the total biomass of native algae until spring 2017, however, its manipulative effect on the total biomass of native algae until spring 2017 is greater than in 2015. Analysis of community structure using a third set of plots and data collected during spring 2017 in the 2017 spring (Figure 4a) and PERMANOVA analysis included the native algae and diatoms (Fig. 4b) revealed PERMANOVA = Pseudo-SIMPER analysis (Table 2) revealed that nearly fifty percent ($F_{1,22} = 2.02$) of the dissimilarity between S- (Table 2) revealed that nearly fifty percent of the dissimilarity between S- (*Sargassum palmeri* and *Zonaria farlowii*) and three species in spring summer (*Sargassum palmeri* and *Zonaria farlowii*) and three species in summer (*Z. farlowii*, *S. palmeri* and *Colpomenia sinuosa*).

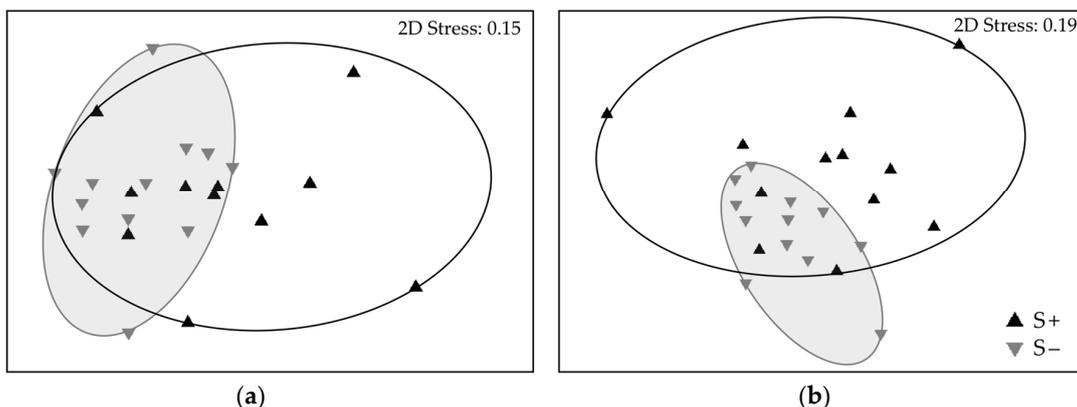


Figure 4. Non-metric multidimensional scaling (MDS) plots showing benthic algal assemblage structure in triplicate plots where *Sargassum horneri* was removed (S-) and in non-removal plots (S+) sampled in 2017. Data are presented by season as (a) spring and (b) summer. The 2D plots analysis used a thin-plate spline with square root transformation and Bray-Curtis similarity similarity index. Two-dimensional (2D) stress values in the plots indicate the degree of mismatch between the predicted values from the regression of the similarity matrix and the distances between samples.

Table 2. Composition of the native algae present in spring and summer 2017 in unmanipulated (S+) plots and those where *Sargassum horneri* was removed (S-) Data are damp biomass (mean \pm SE g m⁻²) and the percent contribution of individual taxa to the top 70% of the dissimilarity between S+ and S- treatments in SIMPER analysis.

Taxonomic Group	Taxon	Spring			Summer		
		S+	S-	%	S+	S-	%
Subcanopy	<i>Stephanocystis neglecta</i>	25.9 \pm 15.3	27.2 \pm 11.3	7.6	27.5 \pm 10.3	19.0 \pm 8.1	5.9
Subcanopy algae	<i>Stephanocystis neglecta</i>	25.9 \pm 15.3	27.2 \pm 11.3	7.6	27.5 \pm 10.3	19.0 \pm 8.1	5.9
	<i>Sargassum palmeri</i>	415.1 \pm 154.1	911.3 \pm 232.5	29.0	172.3 \pm 45.8	262.6 \pm 67.6	14.3
Understorey algae	Articulated coralline spp.	0 \pm 0	1.6 \pm 1.6	.	0.4 \pm 0.4	1.6 \pm 1.6	.
	<i>Aspalagopsis taeniformis</i>	0 \pm 0	0 \pm 0	.	0.3 \pm 0.3	1.1 \pm 1.1	.
Brown blades spp.	Brown blades spp.	0 \pm 0	6.6 \pm 4.3	.	0 \pm 0	0 \pm 0	.
	<i>Cladophora gemminia</i>	0 \pm 0	0 \pm 0	.	0 \pm 0	0 \pm 0	.
Colpomenia sinuosa	<i>Colpomenia sinuosa</i>	0 \pm 0	10.6 \pm 8.4	.	148.3 \pm 72.3	121.7 \pm 5.4	12.3
	<i>Chondria californica</i>	0 \pm 0	0.3 \pm 0.3	.	1.1 \pm 0.5	0.9 \pm 0.4	.
Coraline chlorellids	<i>Coraline chlorellids</i>	20.2 \pm 10.6	30.5 \pm 13.0	7.2	13.6 \pm 7.4	12.3 \pm 7.0	.
	<i>Dictyopteris undulata</i>	3.9 \pm 2.2	23.5 \pm 6.1	6.6	14.8 \pm 6.7	45.9 \pm 12.4	7.2
Dictyota spp.	<i>Dictyota</i> spp.	1.1 \pm 0.7	20.4 \pm 7.7	.	24.0 \pm 11.2	24.0 \pm 11.6	.
	Filamentous brown spp.	0 \pm 0	0 \pm 0	.	24.0 \pm 11.2	24.0 \pm 11.6	.
Filamentous green spp.	Filamentous green spp.	0 \pm 0	0.2 \pm 0.2	.	0.2 \pm 0.2	0.5 \pm 0.3	.
	Filamentous red spp.	1.1 \pm 1.0	1.2 \pm 0.2	.	0.2 \pm 0.2	0.3 \pm 0.3	.
Giant cell spp.	Giant cell spp.	0 \pm 0	0.3 \pm 0.3	.	0 \pm 0	2.4 \pm 1.9	.
	<i>Idicostopsis</i> spp.	0 \pm 0	0.5 \pm 0.5	.	0.7 \pm 0.2	0.7 \pm 0.4	.
Halimnion spp.	<i>Halimnion gracile</i>	20.6 \pm 11.5	20.4 \pm 7.9	.	37.8 \pm 21.3	28.7 \pm 13.4	6.6
	<i>Hydroclathrus clathratus</i>	0 \pm 0	0 \pm 0	.	8.5 \pm 4.2	5.3 \pm 4.3	.
Laurencia pacifica	<i>Laurencia pacifica</i>	0.6 \pm 0.6	3.6 \pm 1.1	.	7.7 \pm 3.0	11.3 \pm 5.2	.
	<i>Lithothrix asperillum</i>	19.8 \pm 10.0	21.0 \pm 11.5	6.5	33.7 \pm 20.9	25.1 \pm 15.0	6.3
Plocamium cartilagineum	<i>Plocamium cartilagineum</i>	2.7 \pm 2.3	2.2 \pm 1.5	.	17.3 \pm 10.5	2.4 \pm 1.4	.
	<i>Pterocladia capillacea</i>	4.6 \pm 4.6	5.3 \pm 4.3	.	8.5 \pm 4.2	6.4 \pm 4.3	.
Rhodymenia californica	<i>Rhodymenia californica</i>	0 \pm 0	1.9 \pm 1.3	.	1 \pm 0.5	0.6 \pm 0.4	.
	<i>Scytosiphon lomentaria</i>	0 \pm 0	0 \pm 0	.	1 \pm 0.7	0 \pm 0	.
Zonaria farlowii	<i>Zonaria farlowii</i>	171.9 \pm 34.9	347.6 \pm 72.2	16.2	182.7 \pm 47.1	513.2 \pm 84.7	17.7
	Cumulative % contribution to dissimilarity	-	-	73.1	-	-	70.3

<i>Rhodymenia californica</i>	0 ± 0	1.9 ± 1.3	.	1 ± 0.5	0.6 ± 0.4	.
<i>Scytosiphon lomentaria</i>	0 ± 0	0 ± 0	.	1 ± 0.7	0 ± 0	.
<i>Zonaria farlowii</i>	171.9 ± 34.9	347.6 ± 72.2	16.2	182.7 ± 47.1	513.2 ± 84.7	17.7
Cumulative % contribution to dissimilarity	-	-	73.1	-	-	70.3

Sargassum horneri displayed a different seasonal pattern in biomass compared to the aggregated biomass of native algae. There was strong seasonality in the biomass of *S. horneri* in S+ plots, remaining low and displaying a different seasonal pattern, slightly increasing in winter and slightly decreasing in spring (Figure 5). By contrast, the aggregated biomass of native algae fluctuated much less throughout the year, with highest mean values recorded in summer and biomass declining through winter. In S- plots, the biomass of native algae continued to decrease into spring, while in S+ plots, an increase in the biomass of native algae occurred, which was driven primarily by the native congener *S. palmeri* in spring 2017.

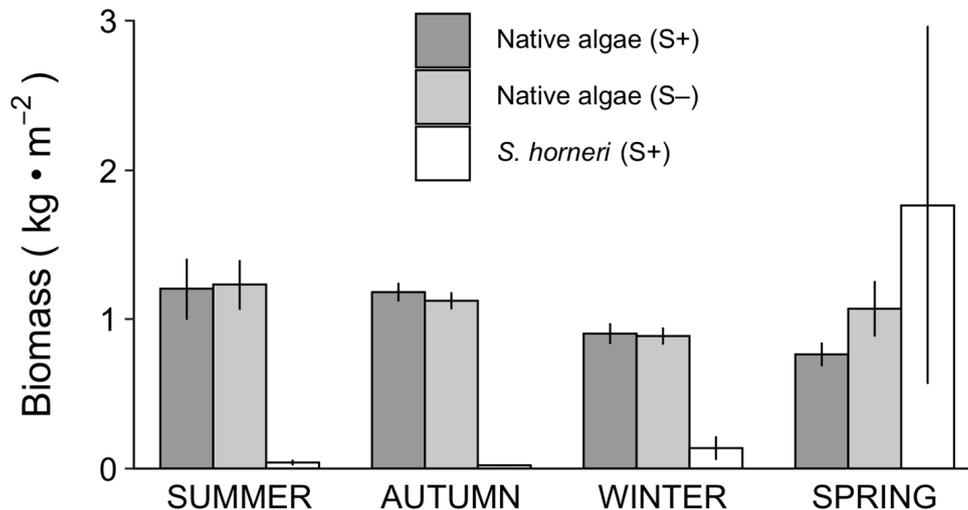


Figure 5. Seasonal mean (\pm SE) damp biomass of native algae (all species combined) and *Sargassum horneri* in *S. horneri*-removal (S-) and non-removal (S+) plots. $N = 4$ years for summer, and 3 years for autumn, winter and spring. **Figure 5.** Seasonal mean (\pm SE) damp biomass of native algae (all species combined) and *Sargassum horneri* in *S. horneri*-removal (S-) and non-removal (S+) plots. $N = 4$ years for summer, and 3 years for autumn, winter and spring.

Results of the depth surveys were consistent with the hypothesis that spatial complementarity with native algae facilitates the invasiveness of *S. horneri*. Two-way ANOVA revealed that the effect of depth on biomass differed for *S. horneri* and native algae ($F_{3,11} = 11.78, p < 0.0001$ for depth \times taxa interaction), and the two were inversely related (Figure 6a). Two-way ANOVA revealed that the effect of depth on biomass differed for *S. horneri* and native algae ($F_{3,11} = 11.78, p < 0.0001$ for depth \times taxa interaction), and the two were inversely related (Figure 6a). *S. horneri* was present from the intertidal to the deepest depth (Sample 9). Biomass of focal species (both as 5 depths of 5–20 m, with only biomass of *S. horneri* and native algae) peaked at the shallowest depth (Figure 6b). The peak at shallow depths of native algae varied with depth (Table S4). Biomass of diatom species (mainly *Sclerodactylus fimbriatus*, *Halidryx dioica* and *Sargassum palmeri*) as well as the native kelp *Eisenia arborea* peaked at shallow depths, while *E. arborea* also occurred at deeper depths in addition to another native kelp, *Agarum fimbriatum*.

3.3. Herbivory
The effects of grazing on the biomass of algae remaining after 48 h assays differed significantly among the five species of algae tested (Figure 7a; ANOVA, $F_4 = 35.146, p < 0.001$). Approximately five times more biomass of *Macrocystis pyrifera* and four times more biomass of *Eisenia arborea* was lost due to grazing compared to the three species of *Sargassum*.

Surveys revealed that the taxonomic composition of algae varied between areas adjacent to and away from urchin halos (Figure 7b; Table S3). There was a significant interaction between taxonomic group and proximity on the relative percent cover (ANOVA, $F_{2,1} = 12.97, p < 0.0001$). Post hoc tests revealed that the cover of *S. horneri* was approximately two times greater near the halos ($p = 0.01$). By contrast, the proximity to halos had no effect on the cover of *S. palmeri* ($p = 0.98$), while that of other native algae taxa near halos was about one third of the level away from halos ($p = 0.001$).

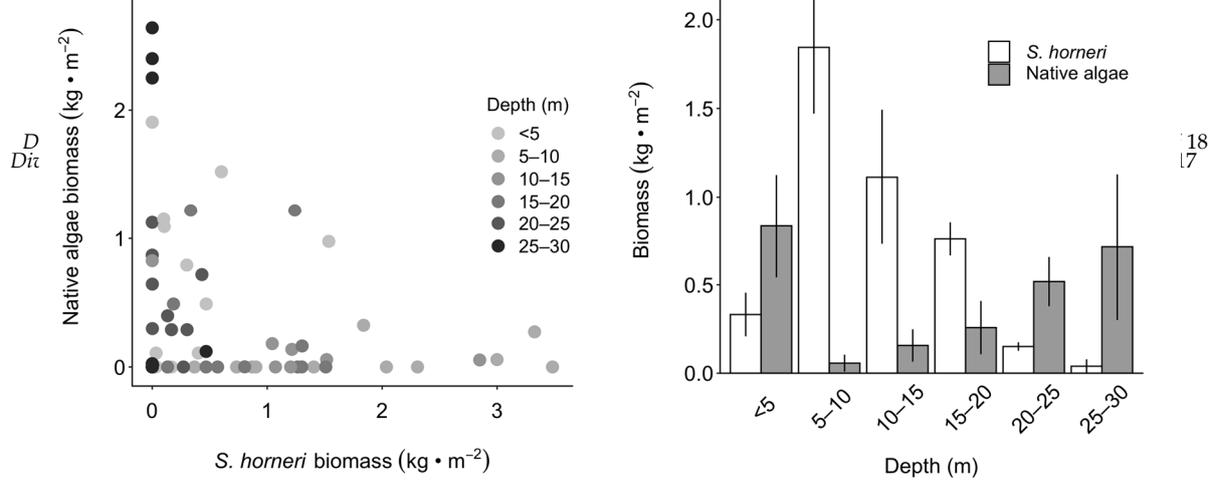


Figure 6. Spatial co-occurrence of *Sargassum horneri* and native algae. Data are damp biomass of *S. horneri* and aggregated damp biomass of all native algae measured within 1 m² quadrats sampled along transects running perpendicular to shore. (a) Points represent individual quadrats, and are shaded by depth bin. $N = 64$ quadrats sampled across 4 sites. (b) Bars represent mean (± SE) biomass of *S. horneri* (white) and the native algae (grey) by 5 m depth bins. The mean and SE of individual species is provided in Table S4. $N = 4$ sites per depth bin except 25–30 m where $N = 3$ sites.

3.3. Herbivory

Figure 6. Spatial co-occurrence of *Sargassum horneri* and native algae. Data are damp biomass of *S. horneri* and aggregated damp biomass of all native algae measured within 1 m² quadrats sampled along transects running perpendicular to shore. (a) Points represent individual quadrats, and are shaded by depth bin. $N = 64$ quadrats sampled across 4 sites. (b) Bars represent mean (± SE) biomass of *S. horneri* (white) and the native algae (grey) by 5 m depth bins. The mean and SE of individual species is provided in Table S4. $N = 4$ sites per depth bin except 25–30 m where $N = 3$ sites.

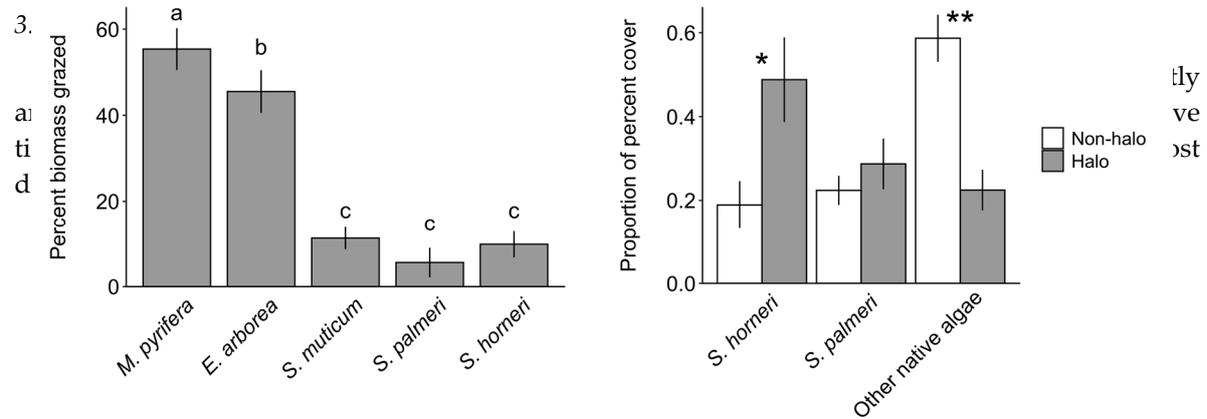


Figure 7. Evidence for consumer avoidance of *Sargassum horneri*. (a) The difference in percent change in biomass (mean ± SE) in randomly paired samples of algae deployed in urchin halos and away from halos inside mesh cages over 48 h periods. Lower case letters differentiate statistically significant differences between species (Tukey's HSD, $p < 0.005$). $N = 57$ paired arrays from four deployment events. (b) Composition of algae adjacent to urchin halos (grey bars) and in nearby reference areas (white bars). Data are the mean proportion of the total percent cover of algae measured in quadrats at 5 m. $N = 15$ quadrats sampled per treatment. Asterisks indicate a significant difference between treatments (*, **: $p < 0.05, 0.01$, respectively).

4. Discussion

The ability of invasive plants to periodically utilize a limited resource has been well documented (e.g. 49, 50) and is the primary mechanism (0.95) that has been attributed to the depressed biomass of *Sargassum* in the adjacent waters off Washington and in USA (30) (reference (31) here). Data are the mean proportion of the total percent cover of algae measured in 1 m² quadrats. $N = 15$

a similar potential to displace native algae as a result of shading caused by the high canopy biomass it achieves during the spring [27]. However, we found little evidence that competitive superiority explains the high invasiveness of *S. horneri* in California as its sustained removal had a minimal effect on the biomass and composition of native algae over a 3.5-year period. Taxonomic richness of the native flora declined over the course of this study but was unresponsive to *S. horneri* removal. The total biomass of native algae was also unaffected by *S. horneri* manipulation until 2017, when it increased sharply in plots where *S. horneri* had been removed. The increase was driven primarily by a perennial congener, *S. palmeri*. This bloom of *S. palmeri* coincided with a large increase in the ambient biomass of *S. horneri* in spring 2017, which dramatically reduced the amount of light reaching the bottom in non-removal plots. Studies of aquatic plants and animals, marsh grasses and marine macroalgae have shown that impacts scale with the abundance of an invader (e.g., [51–54]). In this study, *S. horneri* had no detectable effects until it reached extremely high abundance, at which point only modest impacts to the native algal community occurred, driven primarily by a single closely related species.

The strength of competition between introduced and native species can vary spatially and temporally, depending on fluctuations in biomass driven by species' life histories or environmental factors [55]. The seasonal phenology of the macroalgal community suggested that *S. horneri*'s peak biomass was generally complementary to that of most of the native macroalgae, whose biomass tended to be highest in summer. This pattern was consistent regardless of the presence of *S. horneri* (i.e., in removal and non-removal plots) except during spring 2017 when *S. horneri* was extremely abundant, suggesting it was not a consequence of *S. horneri*, but rather a natural cycle. This conclusion is substantiated by similar estimates of seasonal biomass of native algae at Santa Catalina Island and elsewhere in southern California prior to invasion by *S. horneri* [39,56]. Since the giant kelp, *M. pyrifera*, was absent from our survey and experimental sites throughout nearly the entire course of this study, it did not factor into our analyses. However, like the other native algae we observed, the biomass of *M. pyrifera* in southern California often peaks in the summer and autumn and drops during winter and spring due to wave-induced disturbance to the canopy [57]. Hence, the success of *S. horneri* may be attributed in part to the decreased abundance of native algae during its period of peak growth and reproduction.

The depth distribution of *S. horneri* relative to that of native subcanopy algae could reflect the strength of their competitive interactions or physiological preferences for different parts of the environment. We found that *S. horneri* displayed spatial complementarity with other subcanopy algae as it was most abundant at intermediate depths (5–20 m), while native algae were most abundant at shallower (<5 m) and deeper (>20 m) depths. That the depth distributions of native subcanopy algae observed in our surveys were similar to those reported by others at Santa Catalina Island prior to the arrival of *S. horneri* [58–61] suggests that their lower abundance at intermediate depths was not due to competition with *S. horneri*.

The reasons for the peak in *S. horneri* abundance at intermediate depths in our study are unknown. However, the distribution of *S. horneri* in other regions indicates great versatility in light requirements, and opportunistic growth in situations where competition is minimal. For example, in its native range in Japan, *S. horneri* grows from the intertidal to 20 m [62] but is most common on shallow reefs from the low intertidal to 4 m [63]. In Baja California, Mexico, near the southern extent of its invaded range, *S. horneri* has been reported to occur from the intertidal [64,65] to at least 8 m depth [66]. Perhaps robust subcanopy-forming macroalgal communities at Santa Catalina Island deter *S. horneri* at very deep (>20 m) and very shallow (<5 m) depths, while increased space and light available at intermediate depths allow *S. horneri* to thrive with minimal competition. Such appears to be the case for the annual Asian kelp, *Undaria pinnatifida*, whose invasion success in the United Kingdom has been attributed in part to its broad depth range as well as its niche dissimilarities with native algae as the abundances of *U. pinnatifida* and native algae were inversely correlated along a depth gradient [67].

Our findings revealed that *S. horneri* has the greatest biomass at depths where, and times when, the abundance of native macroalgae is lowest. The consistent phenology of *S. horneri* in its native and invaded range [27] and of most native algae in the presence or absence of *S. horneri* suggest

that niche complementarity between them occurs throughout the year. Recent work by Sullaway and Edwards [68] at nearby sites at Santa Catalina Island supports this idea, showing that *S. horneri* increased rather than decreased levels of community production and respiration in this system. They concluded that *S. horneri* takes advantage of environmental conditions that disturb native algae and thrives as a consequence of disturbance, rather than causing an ecosystem shift due to its ability to outcompete the native flora [69]. Consistent with this idea is the observation by Caselle et al. [7] that *S. horneri* abundance at nearby Anacapa Island was significantly lower in older, well-established marine protected areas (MPAs) where the abundance of native algae was high relative to newly established MPAs. These authors argued that the differences in *S. horneri* abundance between new and old MPAs reflect stronger competition between native algae and *S. horneri* in the older MPAs where native algae flourish. Thus, niche complementarity may allow *S. horneri* to achieve high abundance only in places where competition from native algae is not strong.

Herbivores can influence the invasion success of freshwater and marine macrophytes directly through consumption of the invader, or they can mediate interspecific competition through preferential consumption of native species [23,36,37,70]. These preferences may arise from morphological differences or chemical defenses. For example, algae in the order Fucales (which includes the genus *Sargassum*) typically have high levels of phenolic compounds that are known to deter grazing [37]. Our results are consistent with this hypothesis, demonstrating that grazers consumed the native kelps *M. pyrifera* and *E. arborea* while avoiding *S. horneri* and its congeners *S. palmeri* and *S. muticum*. Our results also support the hypothesis posed by Caselle et al. [7] that urchins avoid *S. horneri* and preferentially consume native algae in areas where they co-occur, thereby reducing the potential for competition between them.

The composition of the benthic algal community reflected the grazer preferences we observed. *Centrostephanus centrotus*, the most abundant species of sea urchin in our study, is known to display strong feeding preferences, decreasing the abundance of favored species dramatically before switching to less-preferred species [38]. We found that native foliose algae were reduced and *S. horneri* was more dominant adjacent to urchin halos compared to nearby reference areas. Interestingly, we found no biomass response to grazing by its perennial congener *S. palmeri*, which is native to southern California. Thus while grazers avoided both species of *Sargassum* in favor of native foliose algae, only *S. horneri* responded to a lack of herbivory with increased abundance. It may be that *S. horneri* is able to colonize space created on the reef more readily than *S. palmeri* due to its annual life history and high fecundity. Traits related to rapid growth and high fecundity, as well as deterrence to herbivory, are often associated with invasive plants [71]. However, defenses often come at a fitness cost [72] and shorter lived, r-selected plants are not typically heavily defended [73]. Yet *S. horneri* is a species with r-selected traits that allow it to rapidly colonize available space, and it is also a member of an order of algae that typically displays high levels of chemical defense. These traits undoubtedly contribute to the ability of *S. horneri* to proliferate in places where interactions with native species are weak.

5. Conclusions

We found that the high propensity of *S. horneri* to invade southern California reefs results largely from its ability to occupy resources underutilized by native species in space and time and to resist grazing relative to native algae. Its annual life history, high fecundity and capacity for widespread dispersal further enhance its ability to colonize novel habitats. The complementary phenology of *S. horneri* and native algae suggest competition between them is generally weak, which is consistent with the results of our 3.5-year manipulative experiment. Our findings indicate the greatest potential for competitive interactions between *S. horneri* and native algae is at intermediate depths during spring when *S. horneri* peaks in biomass. Future work testing the effects of *S. horneri* on native algae should focus on this depth range and season. Collectively, our results highlight the importance of considering exotic marine species in the context of the invasibility of native assemblages when assessing their invasiveness and developing management strategies for controlling their spread.

Supplementary Materials: The following are available online at <http://www.mdpi.com/1424-2818/12/2/54/s1>, Table S1: List of native subcanopy and understory algal taxa recorded in the *Sargassum horneri* removal experiment, Table S2: Coefficients (a = intercept, b = slope), r^2 , and p values or SE for formulas to convert size-specific density, or percent cover to damp biomass (g), Table S3: Proportional percent cover of algae adjacent to and away from sea urchin halos, Table S4: Depth distribution of *Sargassum horneri* and native subcanopy algae. The data presented in this manuscript are available online for download from the Long Term Ecological Research Network Data Portal, doi:10.6073/pasta/2c2237bb3cee86e7c6d9488e8ce2795d [74].

Author Contributions: Conceptualization, D.C.R., L.M.M. and S.J.H.; methodology, D.C.R., L.M.M. and S.J.H.; software, L.M.M.; validation, L.M.M.; formal analysis, L.M.M.; investigation, D.C.R. and L.M.M. resources, D.C.R.; data curation, L.M.M.; writing—original draft preparation, L.M.M.; writing—review and editing, D.C.R., L.M.M. and S.J.H.; visualization, L.M.M.; supervision, D.C.R. and S.J.H.; project administration, L.M.M.; funding acquisition, D.C.R., L.M.M. and S.J.H. All authors have read and agreed to the published version of the manuscript.

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Region Nine Kelp Consortium Annual Meeting August 18, 2020

Size of Kelp Beds in 2019 Orange & San Diego Counties



OUTLINE OF PRESENTATION

1. Background Information
2. Survey Methods
3. Region Nine Survey Results
 - Status in 2018
 - Regional Overview for 2019
 - Descriptions of Individual Beds
4. Factors Affecting Kelp Beds
5. Conclusions
6. Preview of 2020

BACKGROUND INFORMATION



Region Nine Kelp Surveys

- ▶ Annual surveys each year >50 years (1967 to 2019)
- ▶ Methods developed by Dr. Wheeler North, Caltech (Pasadena)
- ▶ Region Nine Kelp Survey Consortium formed in 1982 (San Diego RWQCB and several ocean dischargers)
- ▶ Program funded by NPDES permit requirements for major dischargers

Central Region Kelp Surveys

- ▶ Sporadic surveys >50 years (five from 1967 to 1998, annually 1999 to 2019, except 2001)
- ▶ Central Region Kelp Survey Consortium formed in 2003 (Los Angeles RWQCB and several ocean dischargers)
- ▶ Used Region Nine model - program funded by NPDES permit requirements for major dischargers

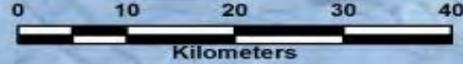
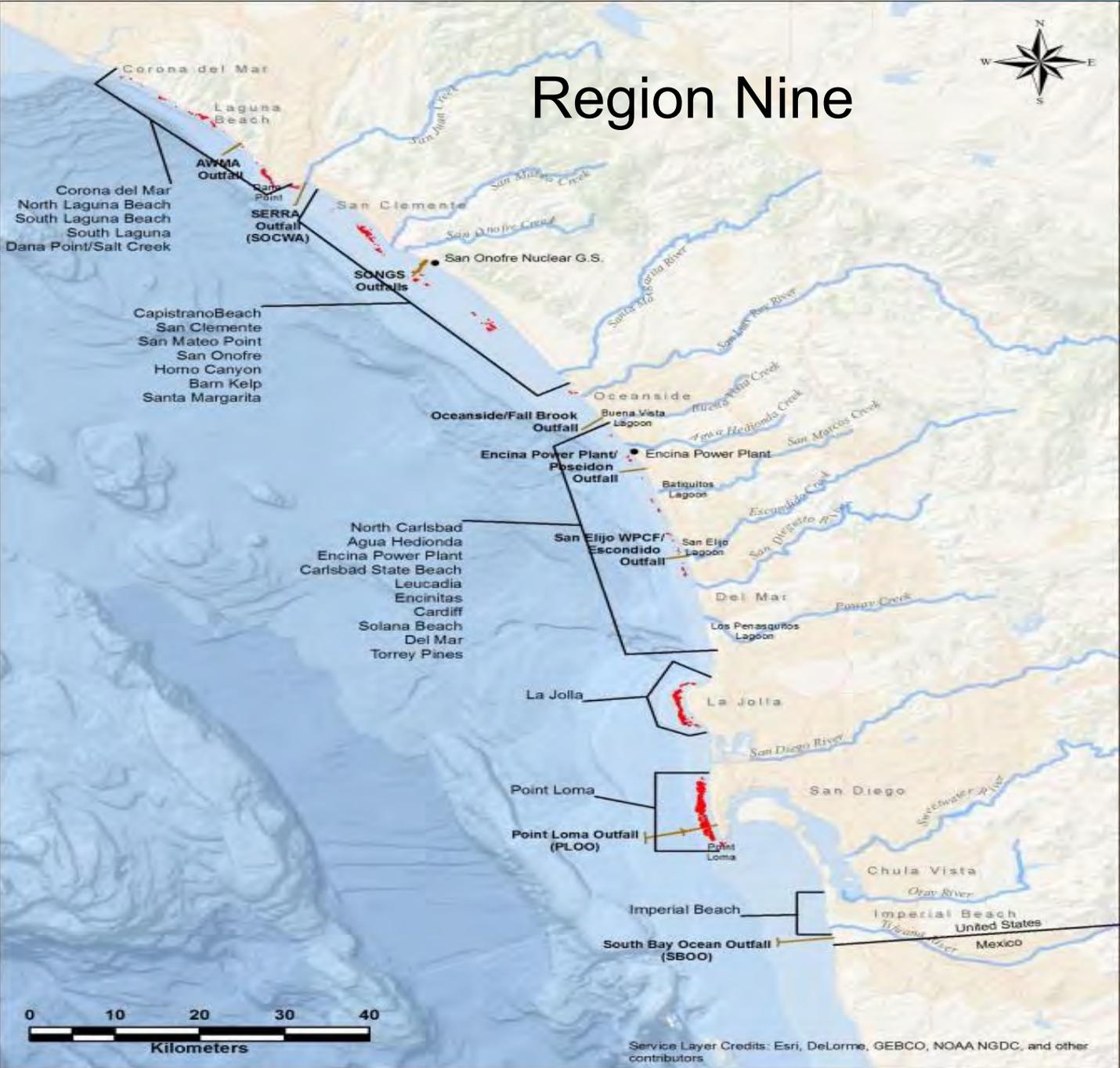
SoCal Kelp Consortia Web Site

- ▶ <https://www.mbcaquatic.com/service/socal-kelp-consortium>
- ▶ Annual reports: 2010 to 2018
- ▶ List of consortium members for Region Nine and Central Region
- ▶ Meeting information
- ▶ Status of kelp in 2018

SURVEY METHODS



Region Nine



Service Layer Credits: Esri, DeLorme, GEBCO, NOAA NGDC, and other contributors

Kelp Overflights



- Ecoscan
(Santa Cruz)
- Cessna 182
- 30-mm lens
- Color IR film

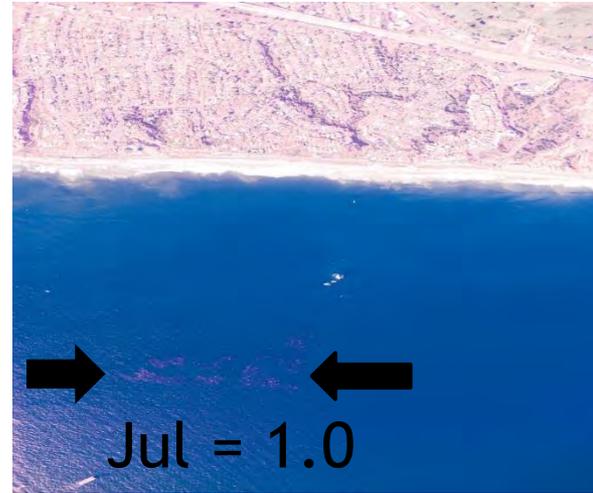
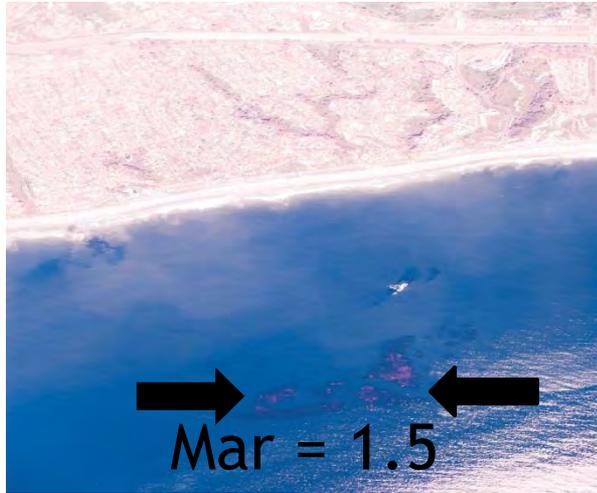
@200 photos per survey

- 10-14,000 feet
- Wind <10 knots
- Swell <1.5 m
- Tides <1 foot range
- Sun angle >30°

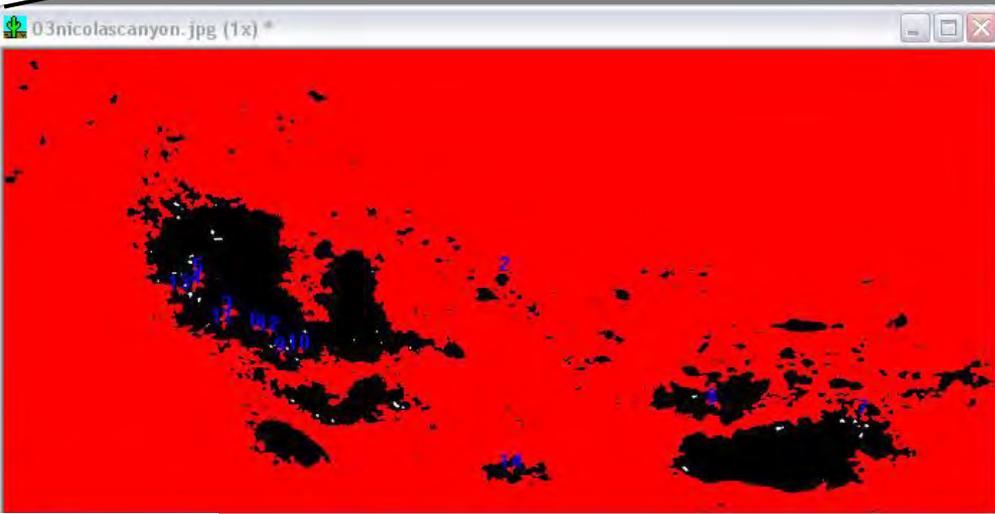
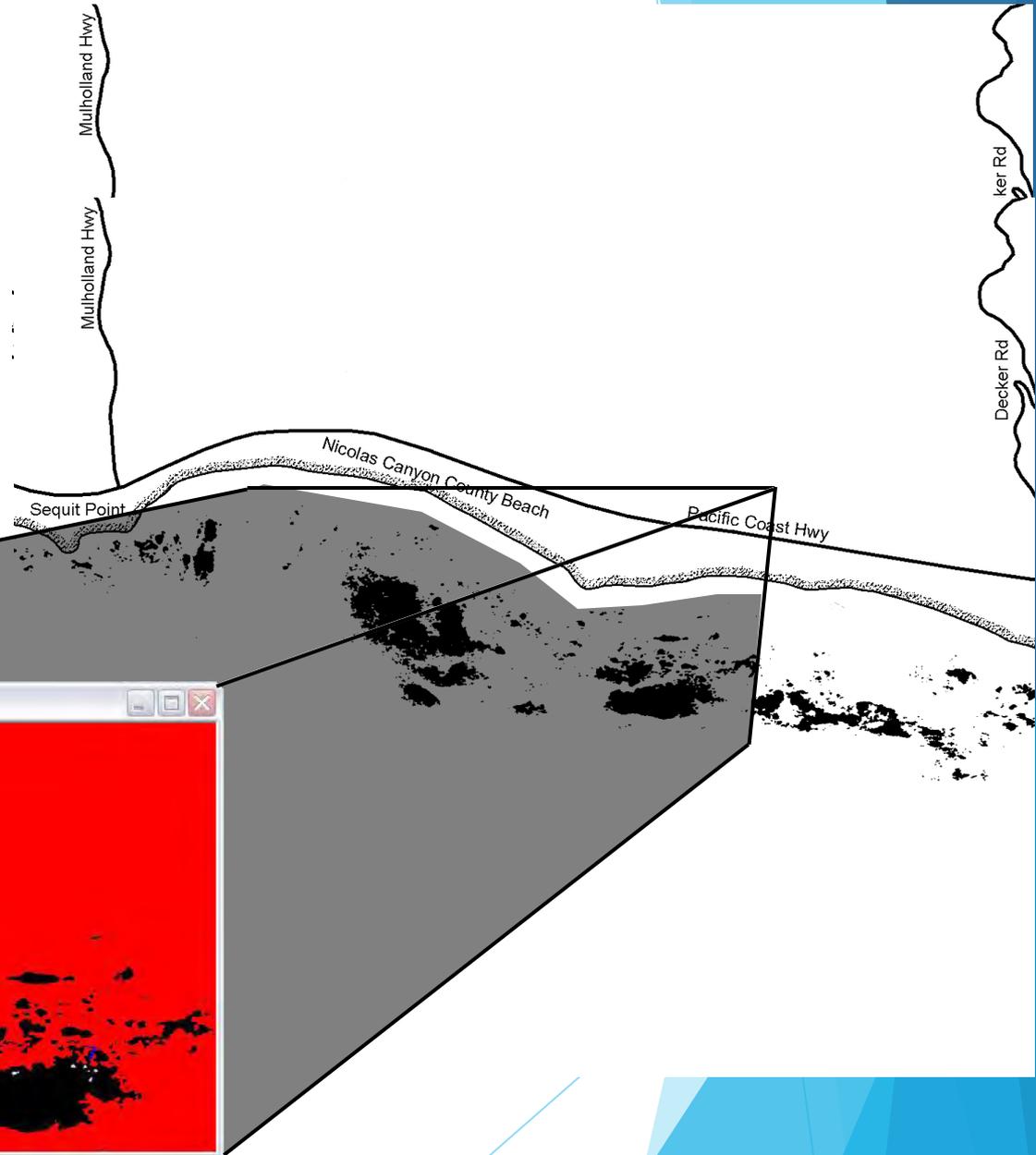
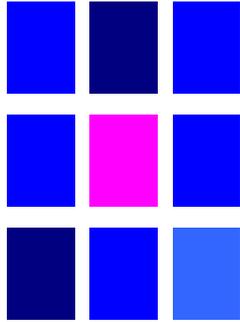
2019 Kelp Overflights

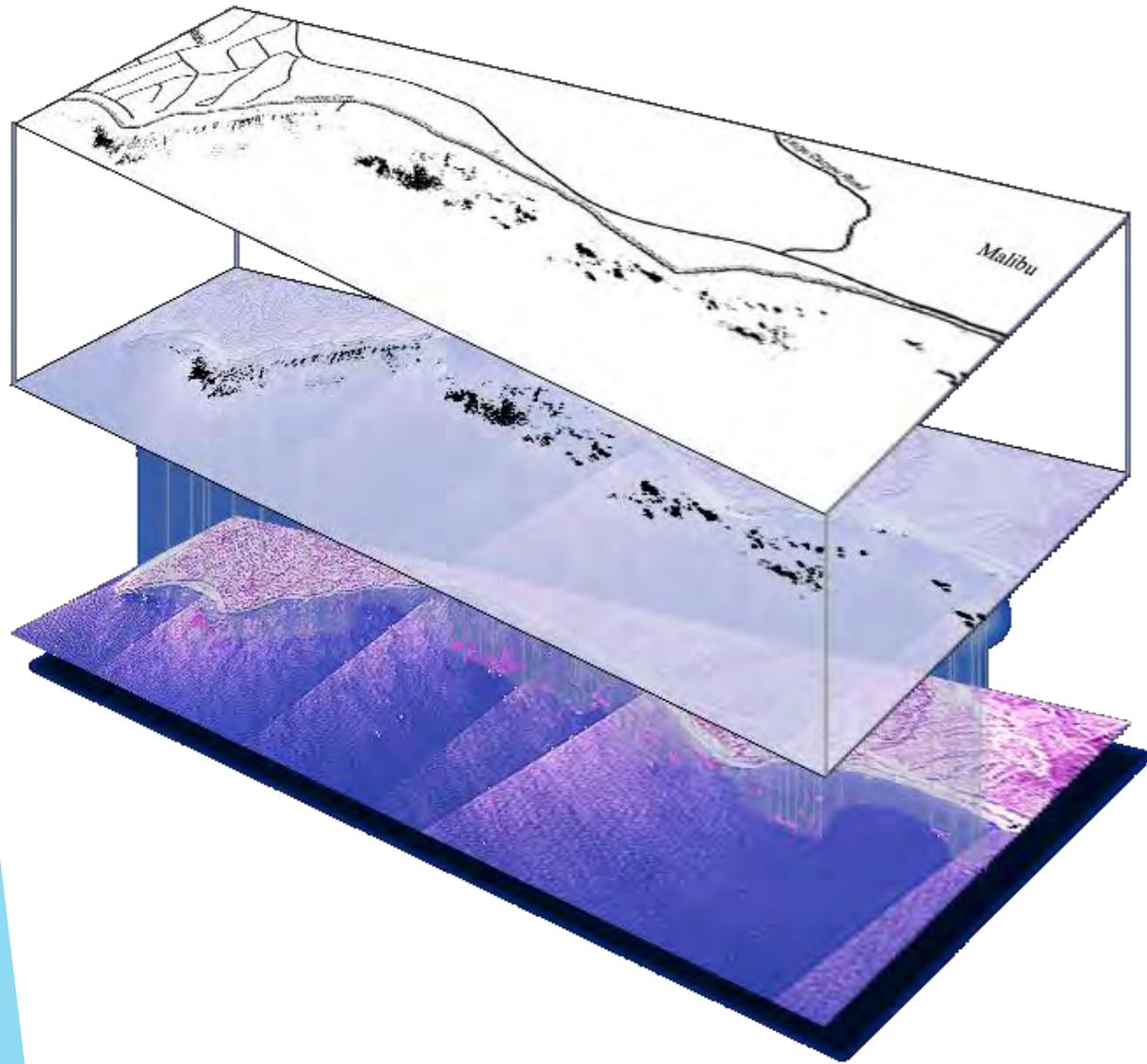
Survey	Date
1 st Quarter	March 31, 2019
2 nd Quarter	July 19, 2019
3 rd Quarter	September 19, 2019
4 th Quarter	December 19, 2019

Maximum Canopy Area San Clemente



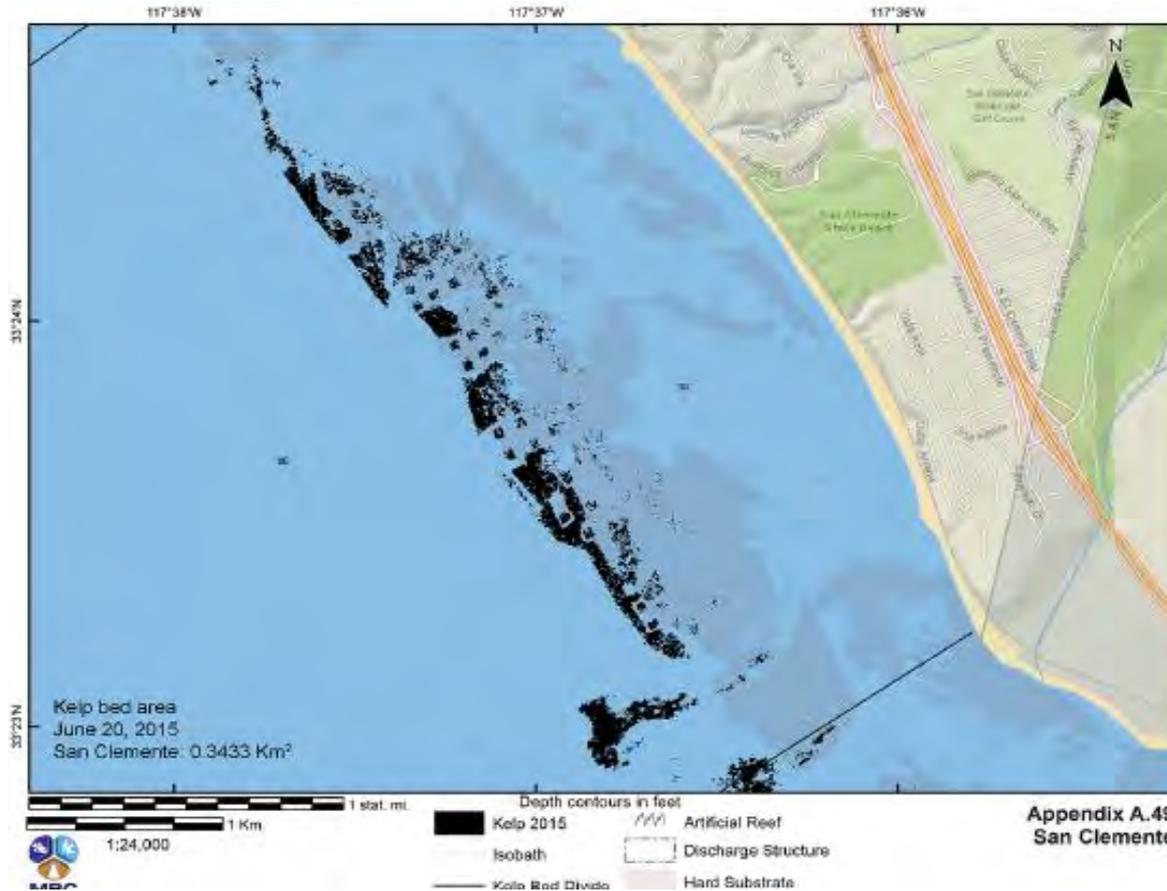
*Images imported to ArcGIS
Several images for some beds
Georeferenced to 3 map features
Surface canopy calculated*





*Photomosaics
Kelp extracted
Layered to
basemaps*

Standardized Basemaps of the Coastline



Region 9 Vessel Survey

- January 4, 15 & 30, 2020
- Surface observations
 - Approximate extent of surface canopy
 - Tissue color, age of fronds, encrustations
 - Subsurface kelp
- In-water diver surveys
- Dana Point/Salt Creek, Encina Power Plant, Leucadia north kelp beds
 - Marine life (e.g., urchins, fish)
 - Age and color of fronds
 - Presence and extent of subsurface giant kelp and other algae



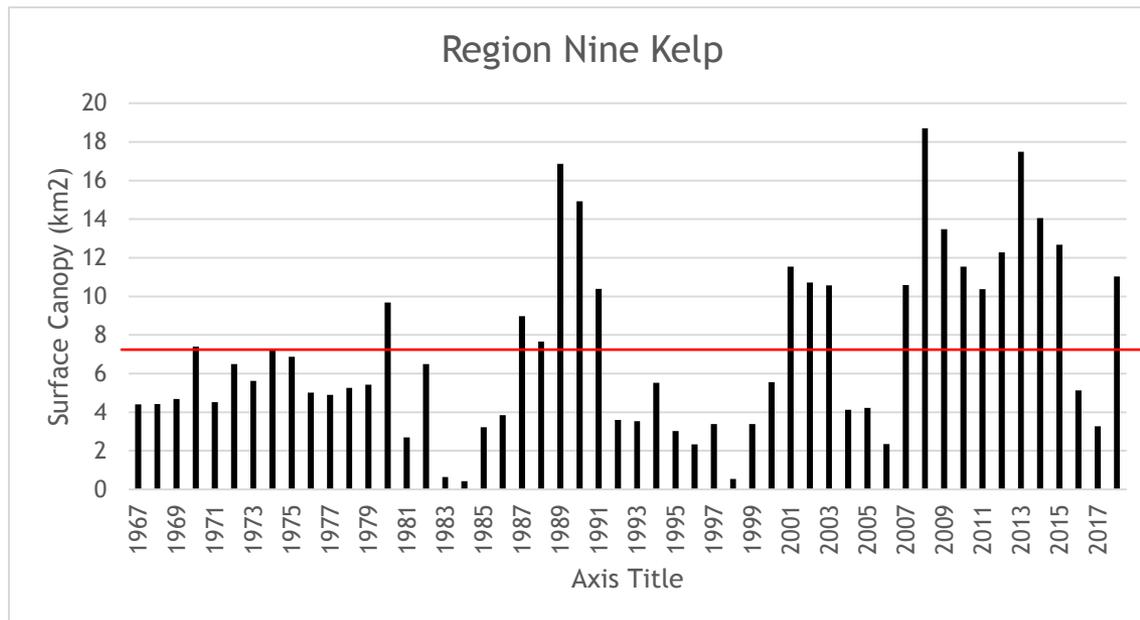
REGION NINE SURVEY RESULTS



Region Nine 2018 Overview

- Total canopy coverage increased substantially
 - 15 beds increased in size
 - 4 beds decreased in size
 - 1 bed disappeared (Carlsbad State Beach)

11.0 km²
3.3 km²
+237%



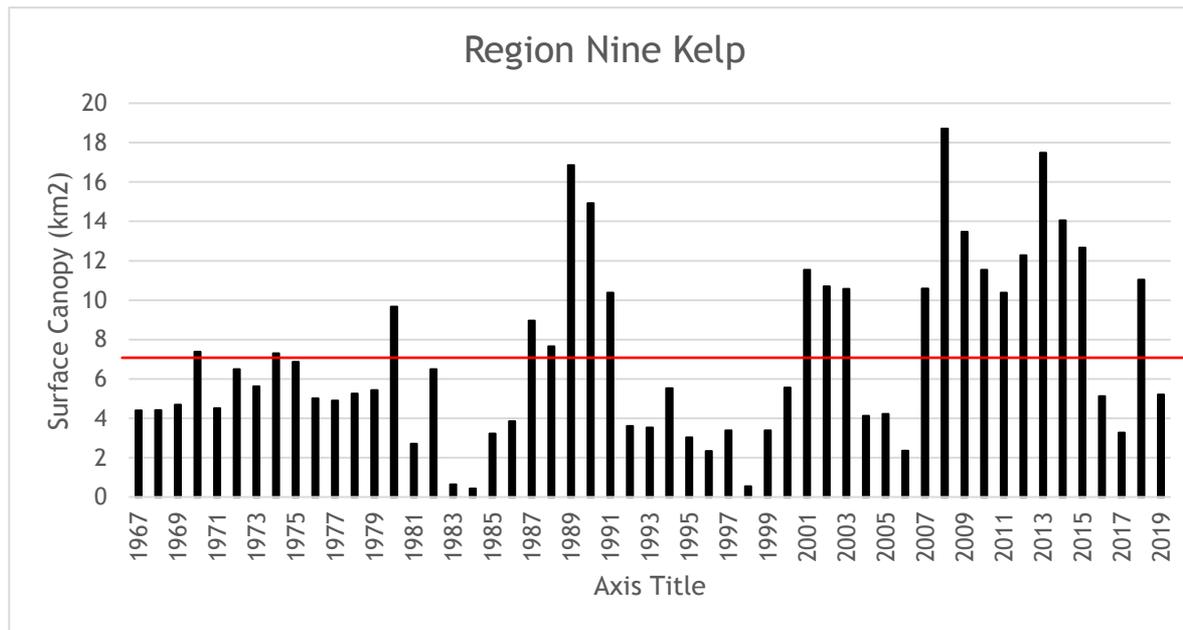
Region Nine 2019 Overview

- Total canopy coverage decreased substantially
 - 18 beds decreased in size
 - 10 beds disappeared

11.0 km²

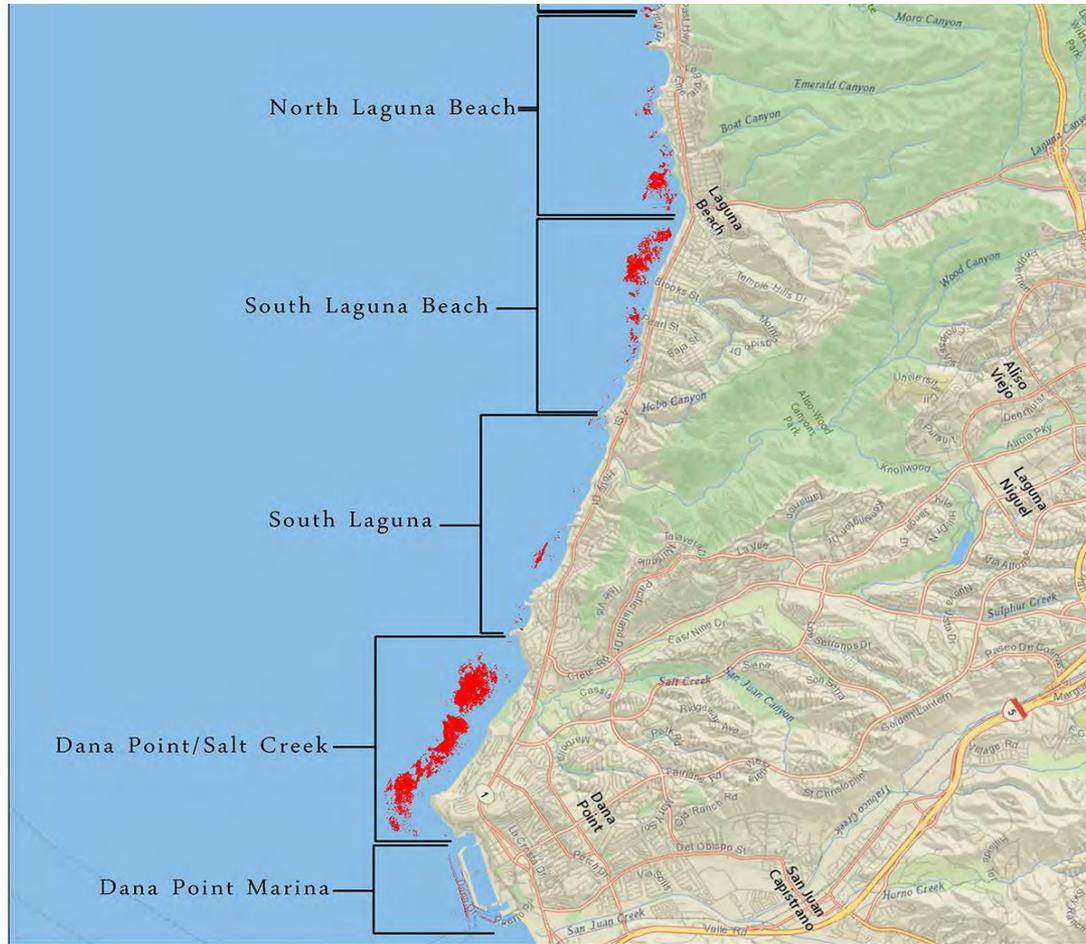
-53%

5.2 km²



	2019 Quarterly Overflights			
	March 31	July 19	September 19	December 19
Kelp Beds				
North Laguna Beach	0.5	0.5	—	0.5
South Laguna Beach	0.5	0.5	—	0.5
South Laguna	—	—	—	—
Salt Creek-Dana Point	—	—	—	—
Dana Marina *	—	—	—	—
Capistrano Beach	—	—	—	—
San Clemente	1.5	1.0	—	—
San Mateo Point	0.5	—	—	—
San Onofre	0.5	0.5	—	—
Pendleton Reefs *	—	—	—	—
Horno Canyon	—	—	—	—
Barn Kelp	—	—	—	—
Santa Margarita	—	—	—	—
Oceanside Harbor *	—	—	—	—
North Carlsbad	—	—	—	—
Agua Hedionda	—	—	—	—
Encina Power Plant	—	—	—	—
Carlsbad State Beach	—	—	—	—
North Leucadia	—	0.5	—	—
Central Leucadia	—	—	—	—
South Leucadia	—	—	—	—
Encinitas	—	—	—	—
Cardiff	—	—	—	—
Solana Beach	—	—	—	—
Del Mar	—	—	—	—
Torrey Pines Park	—	—	—	—
La Jolla Upper	0.5	1.5	1.0	1.0
La Jolla Lower	2.5	3.0	1.0	2.5
Point Loma Upper	3.0	4.0	1.5	3.5
Point Loma Lower	3.0	4.0	1.5	2.5
Imperial Beach	—	—	—	—

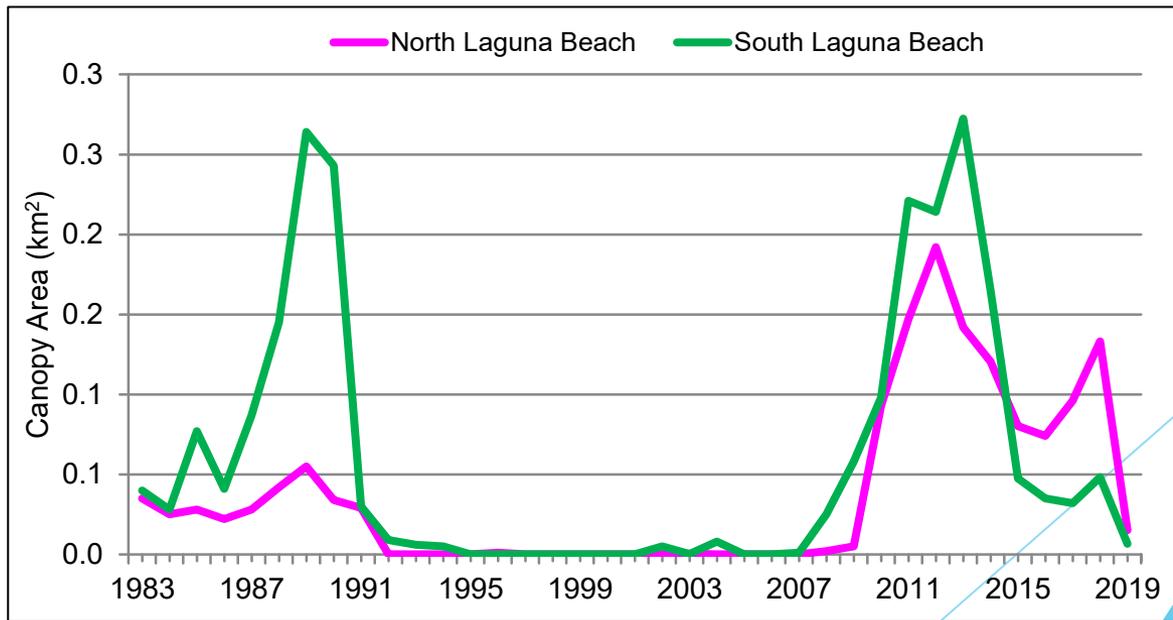
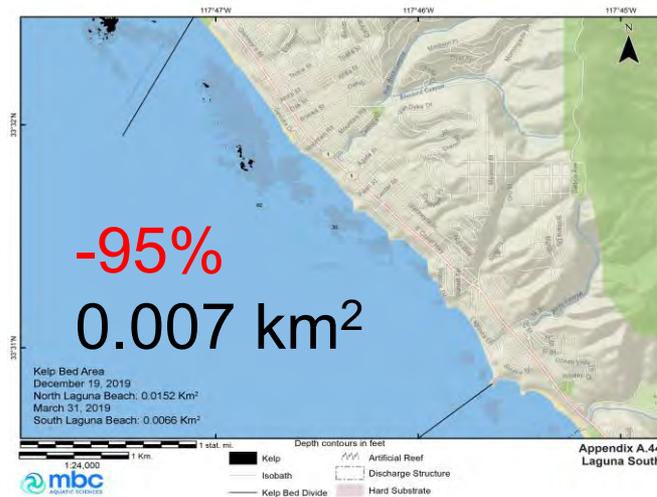
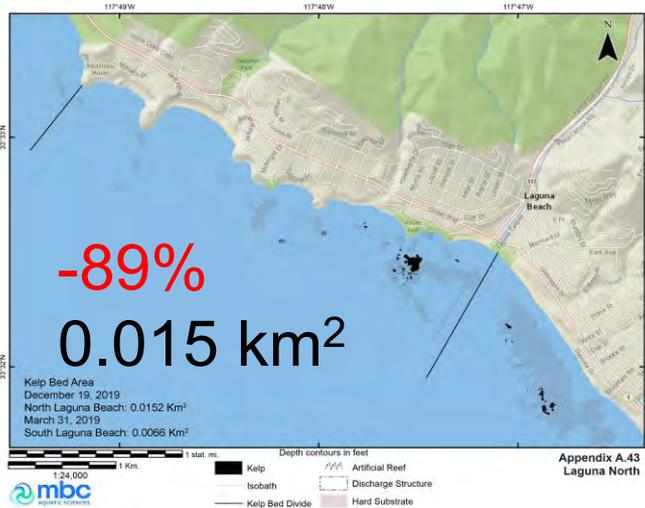
Orange County



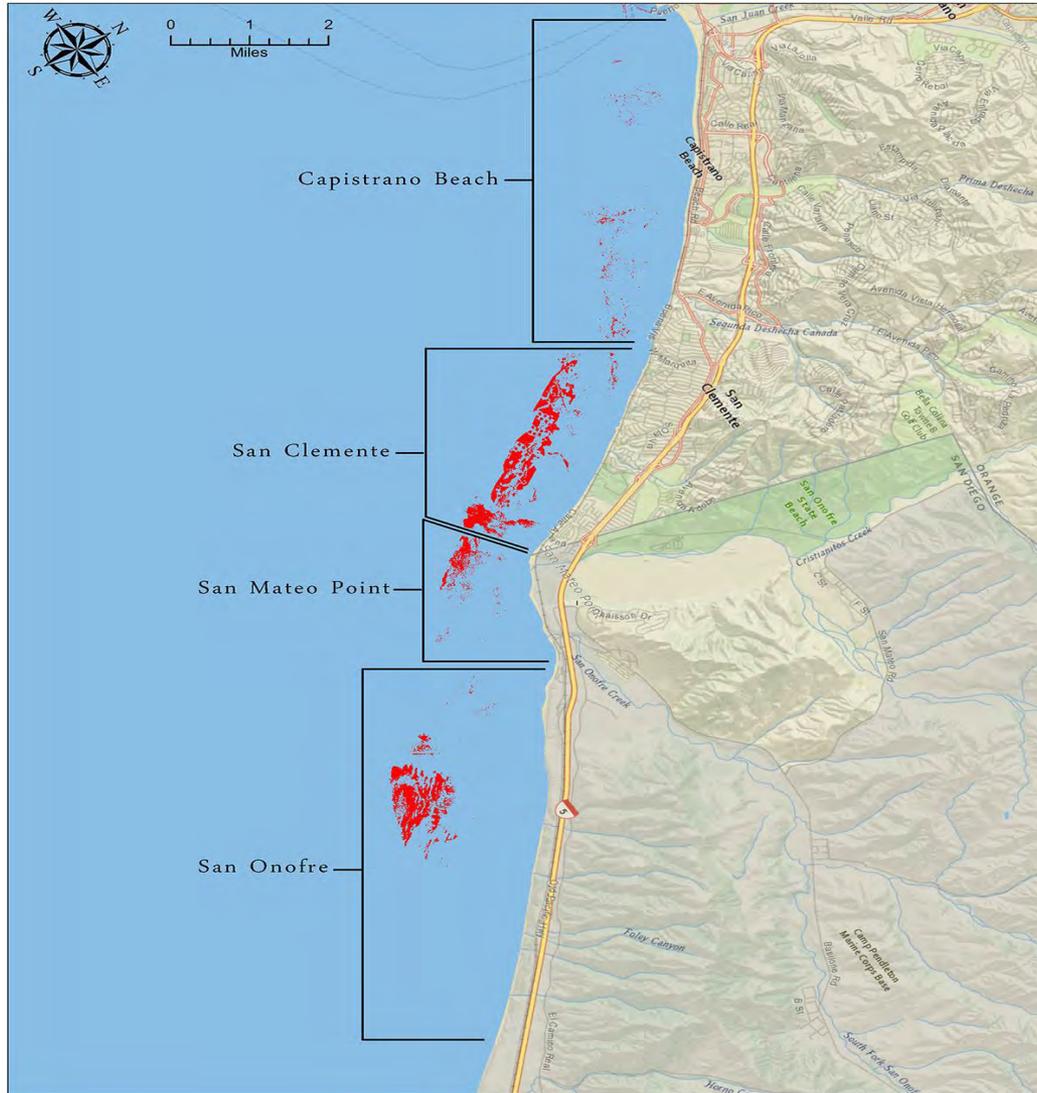
Appendix D.9

North Laguna Beach

South Laguna Beach

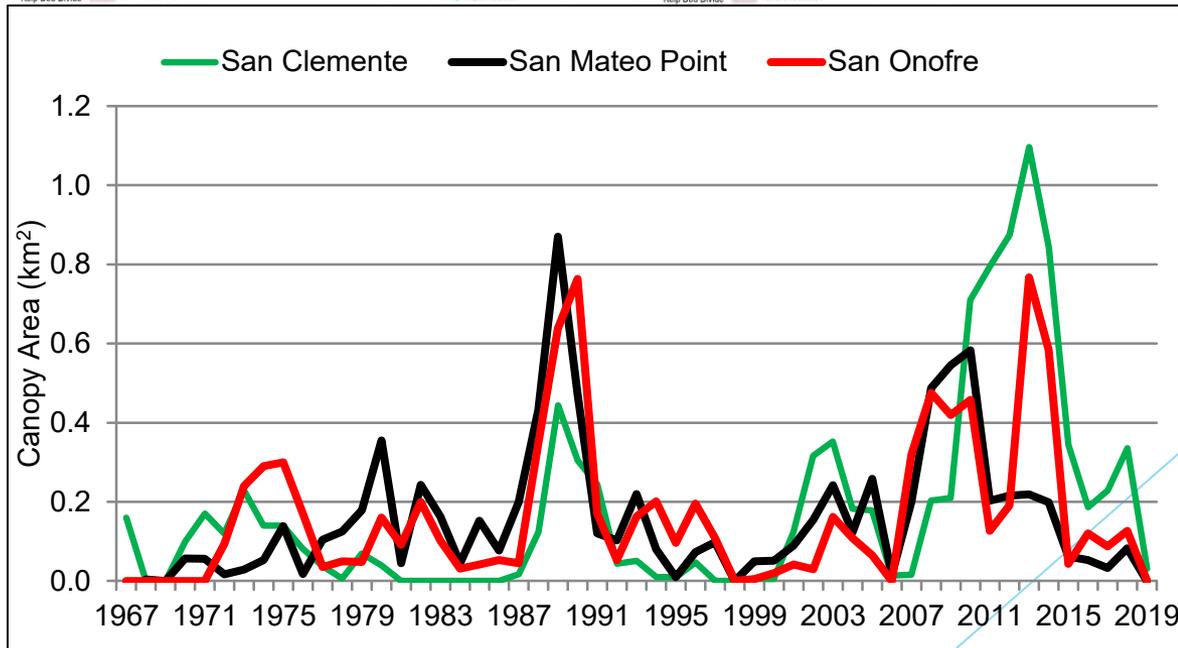
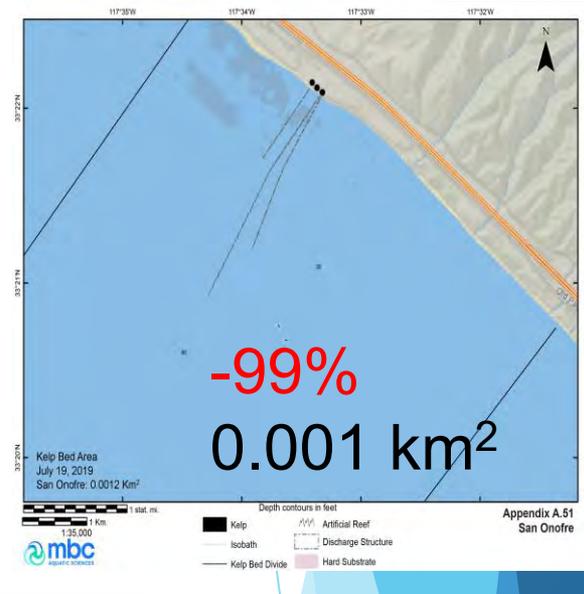
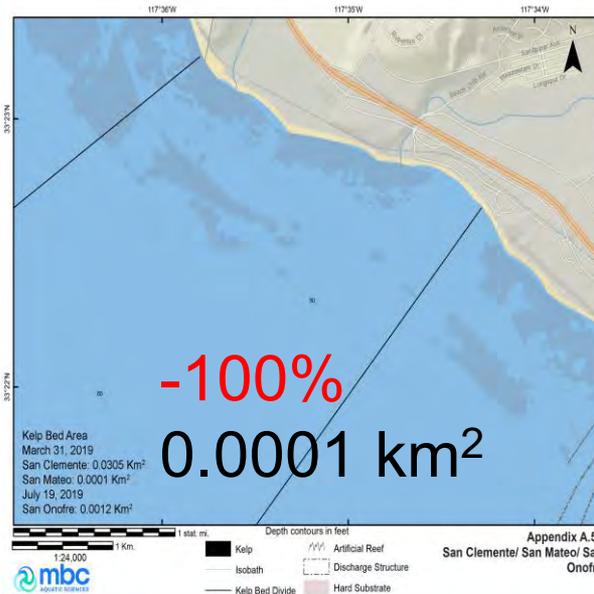
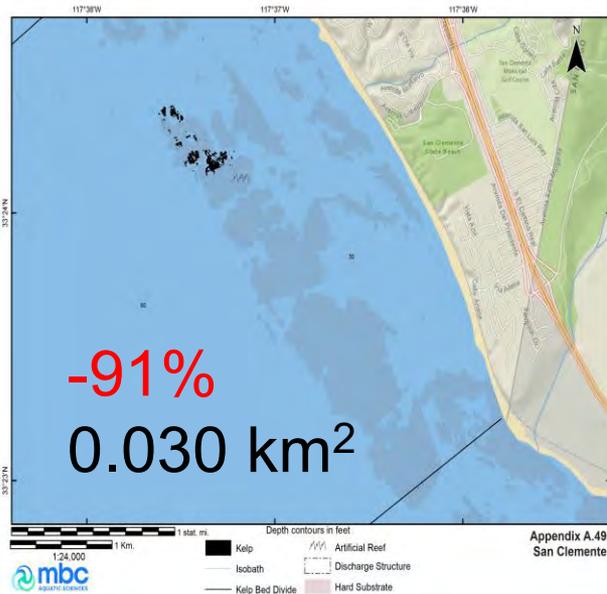


Capistrano Beach to San Onofre

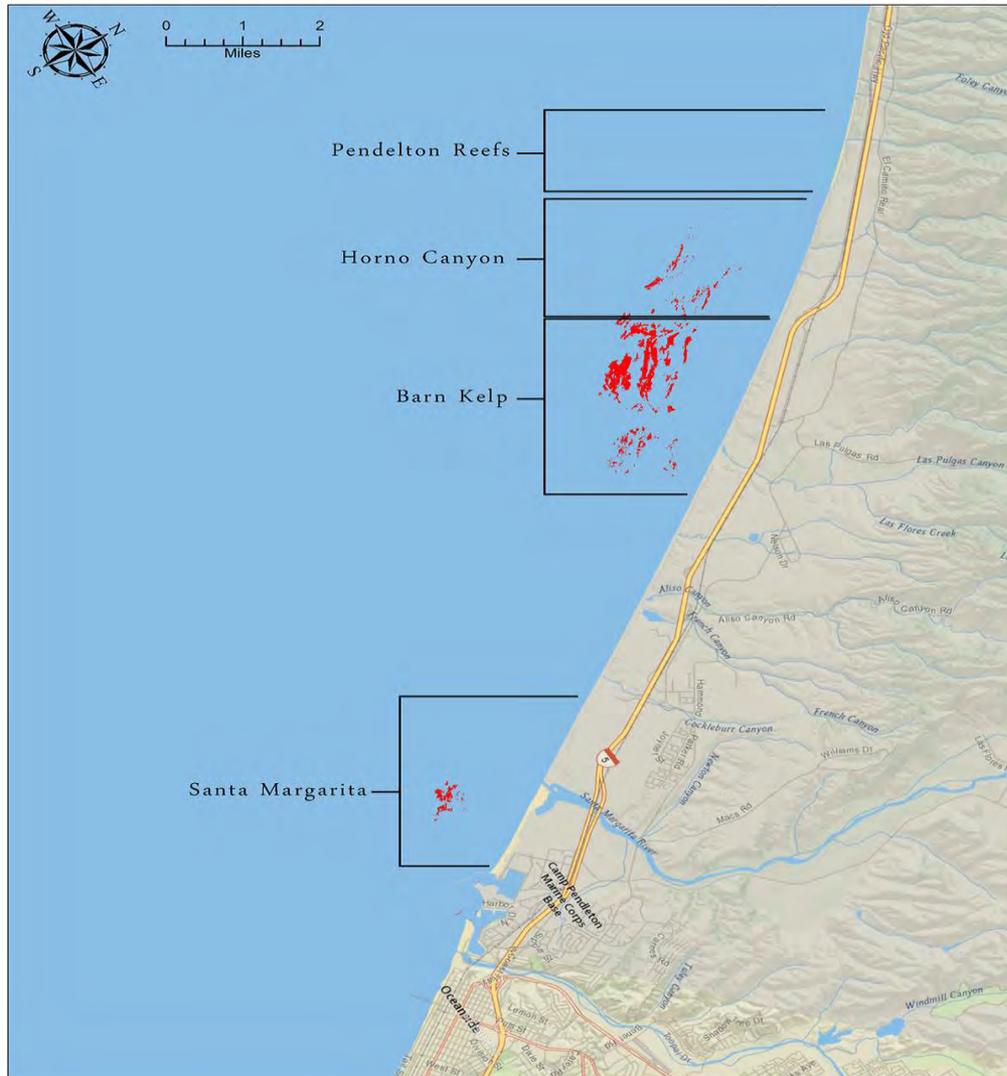


Appendix D.10

San Clemente San Mateo Pt San Onofre

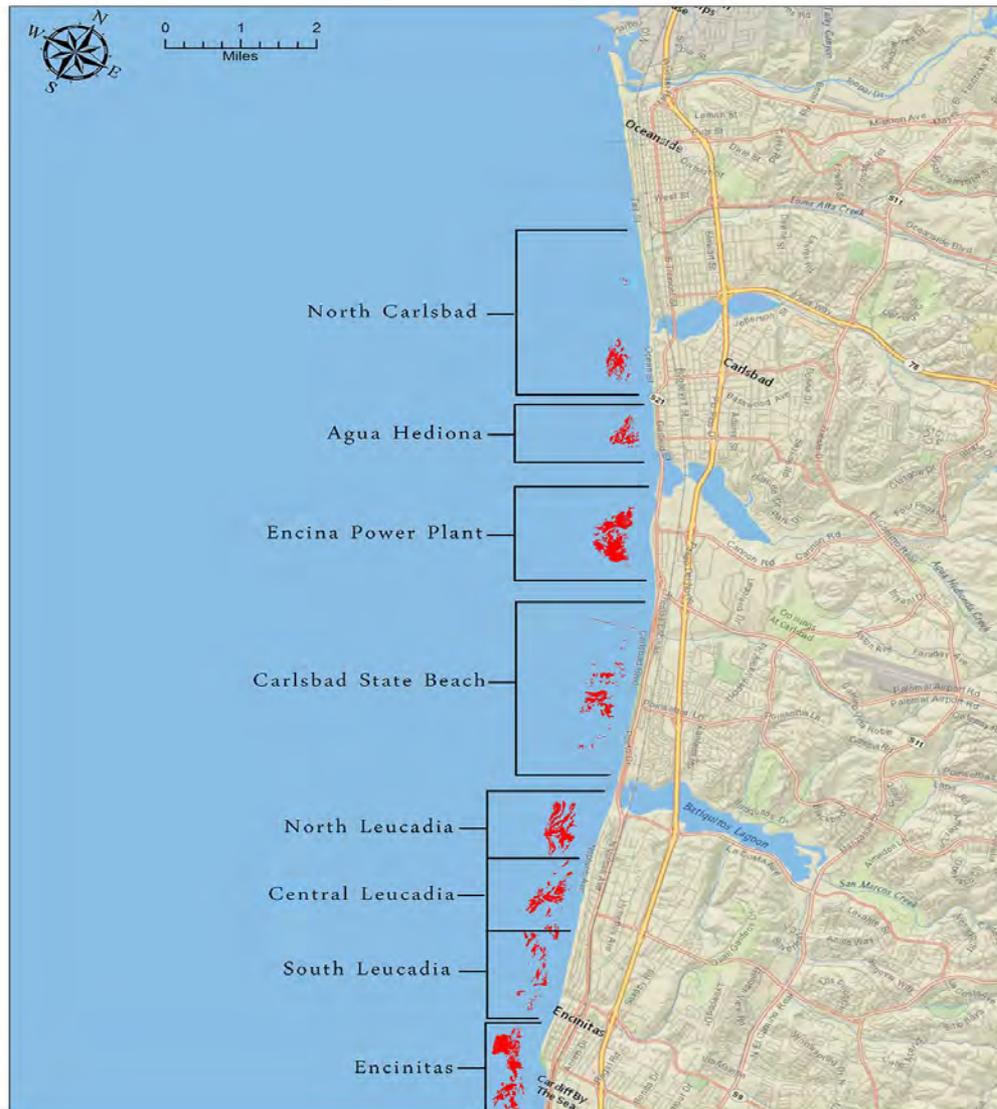


Horno Canyon to Santa Margarita



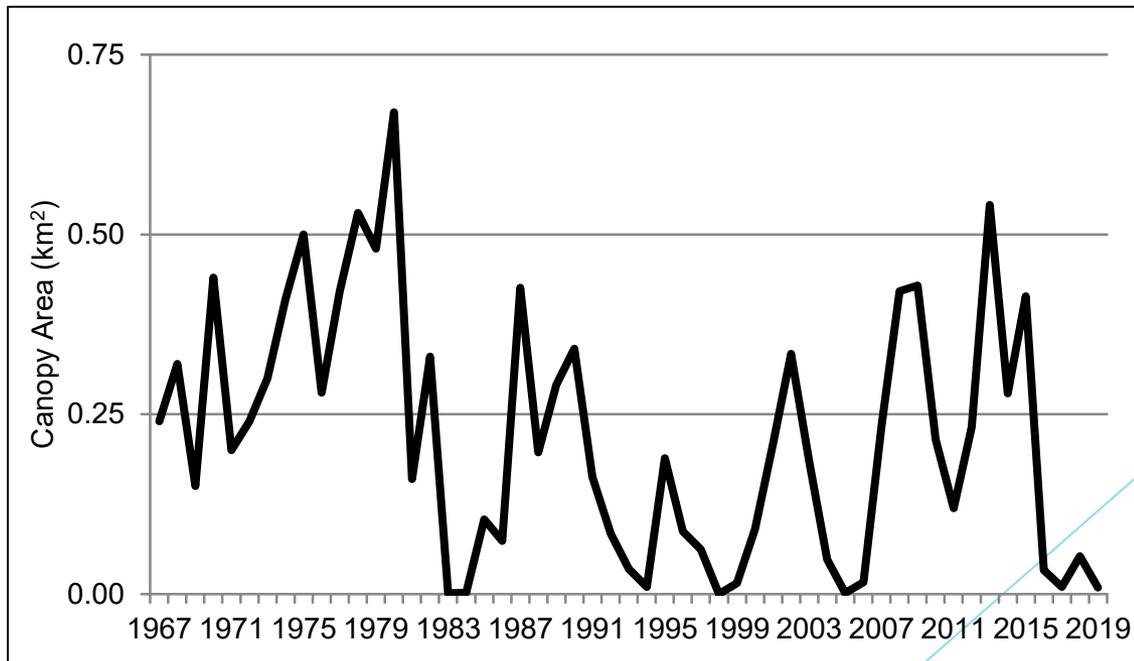
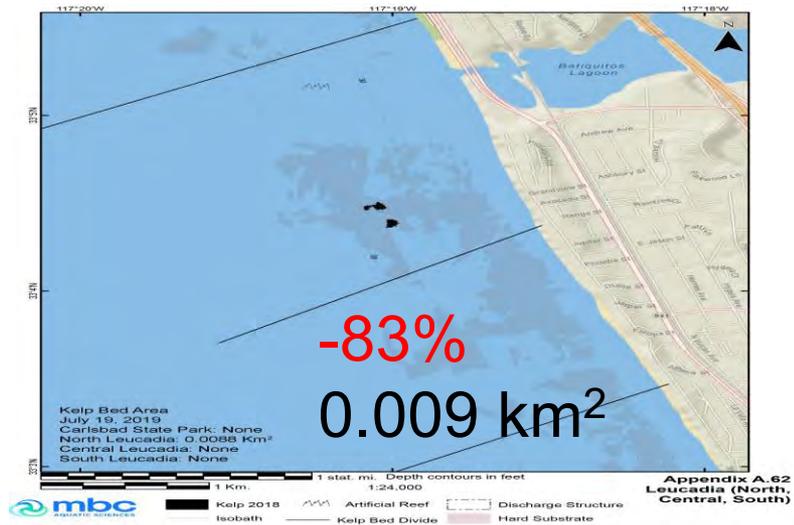
Appendix D.11

North Carlsbad to Encinitas

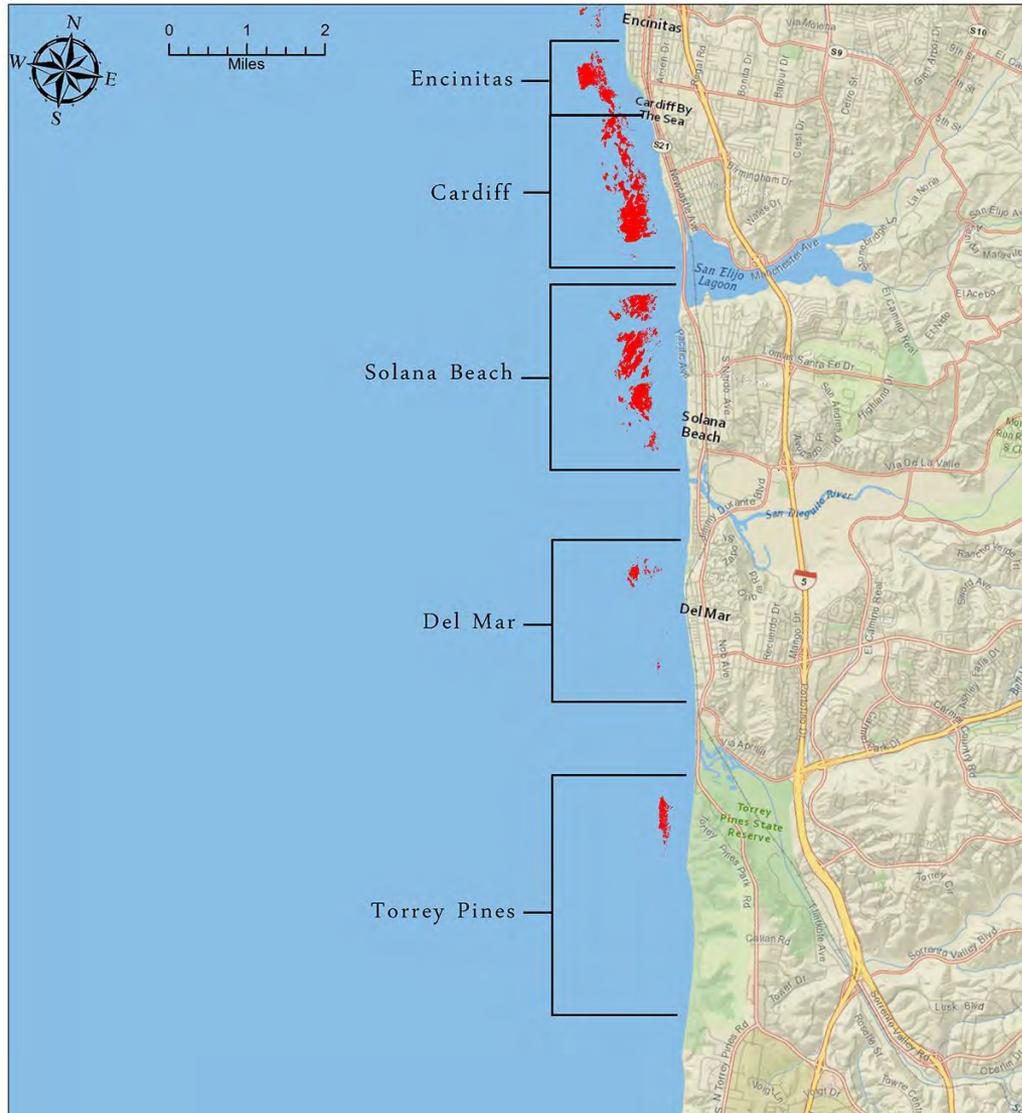


Appendix D.12

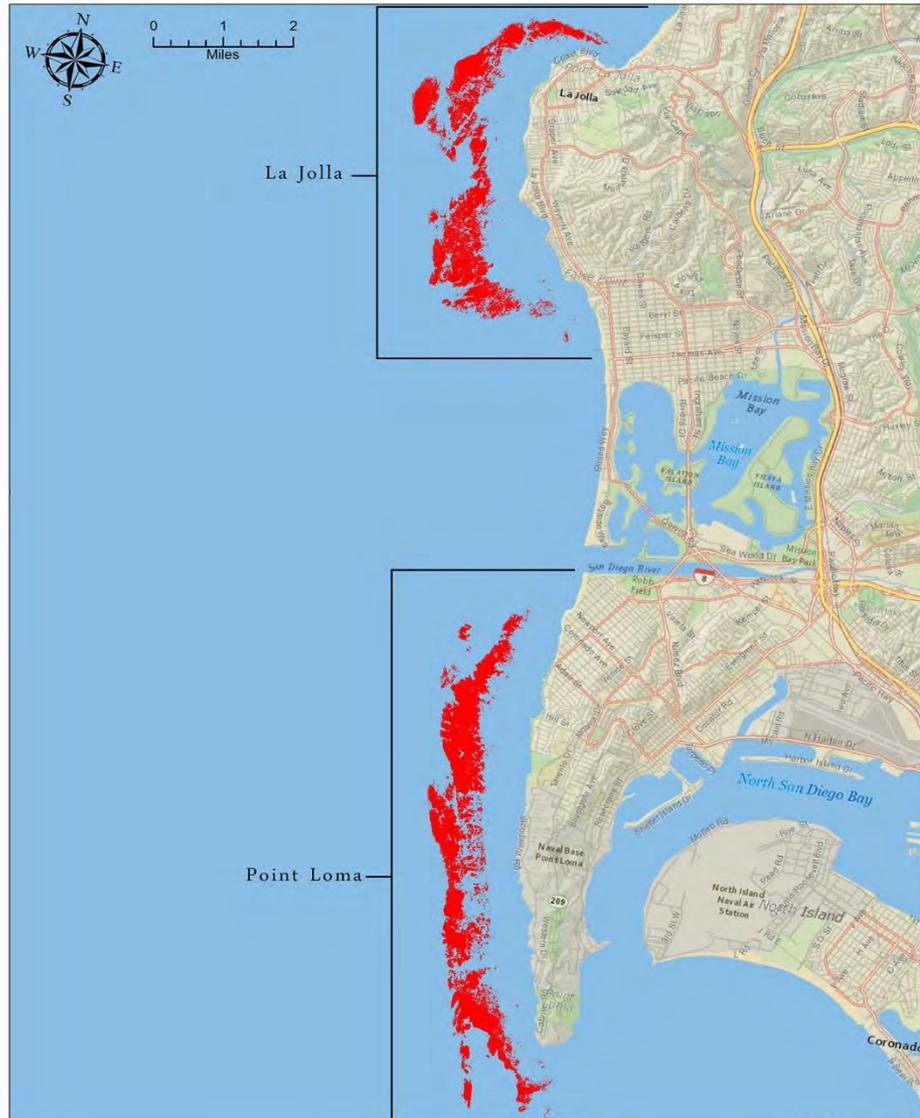
Leucadia - North



Cardiff to Torrey Pines

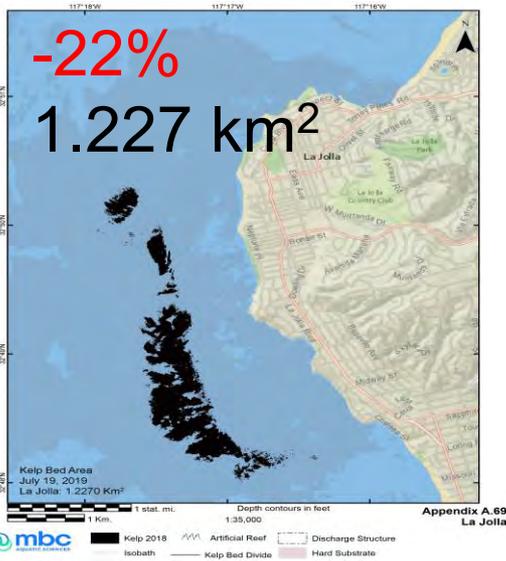


La Jolla and Point Loma

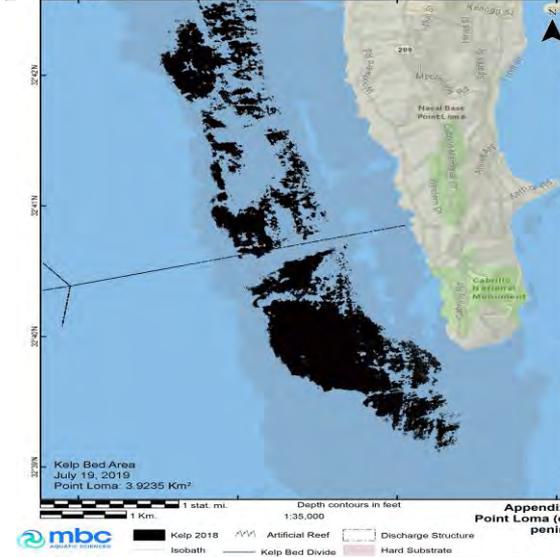
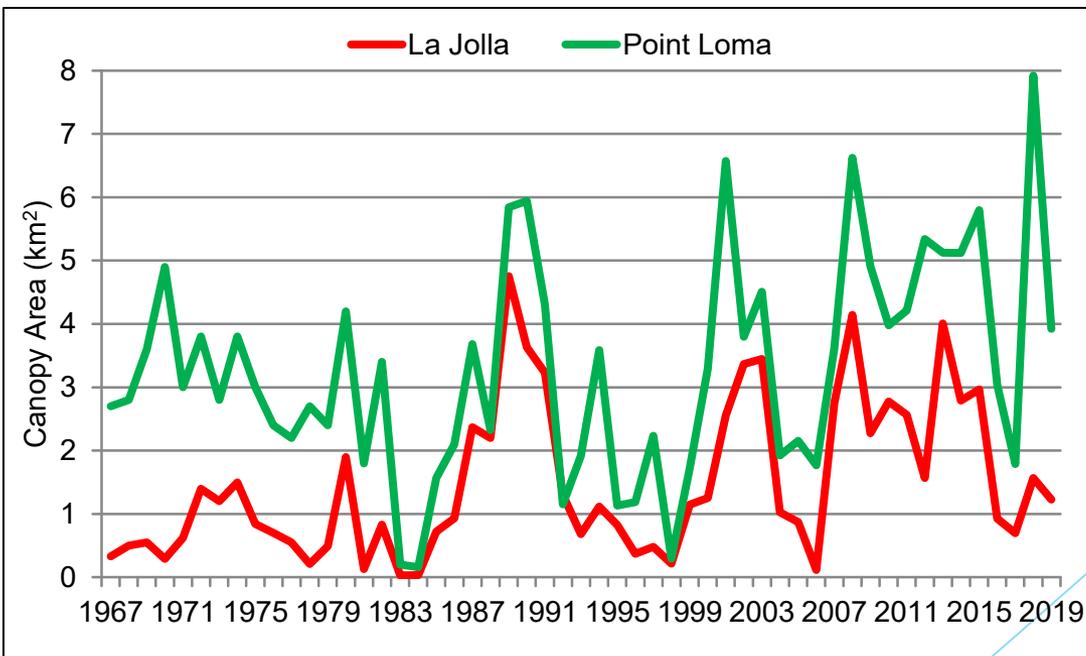
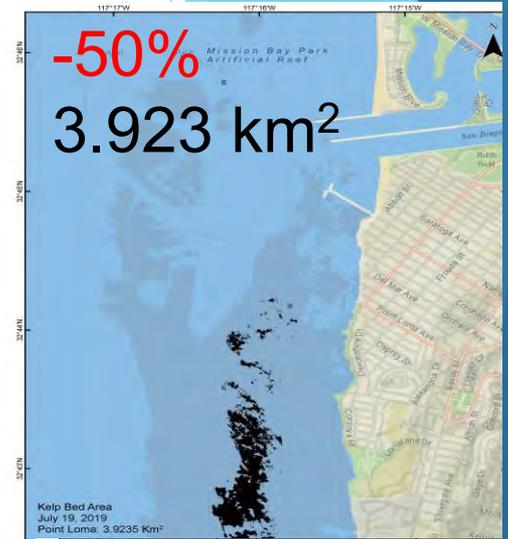


Appendix D.14

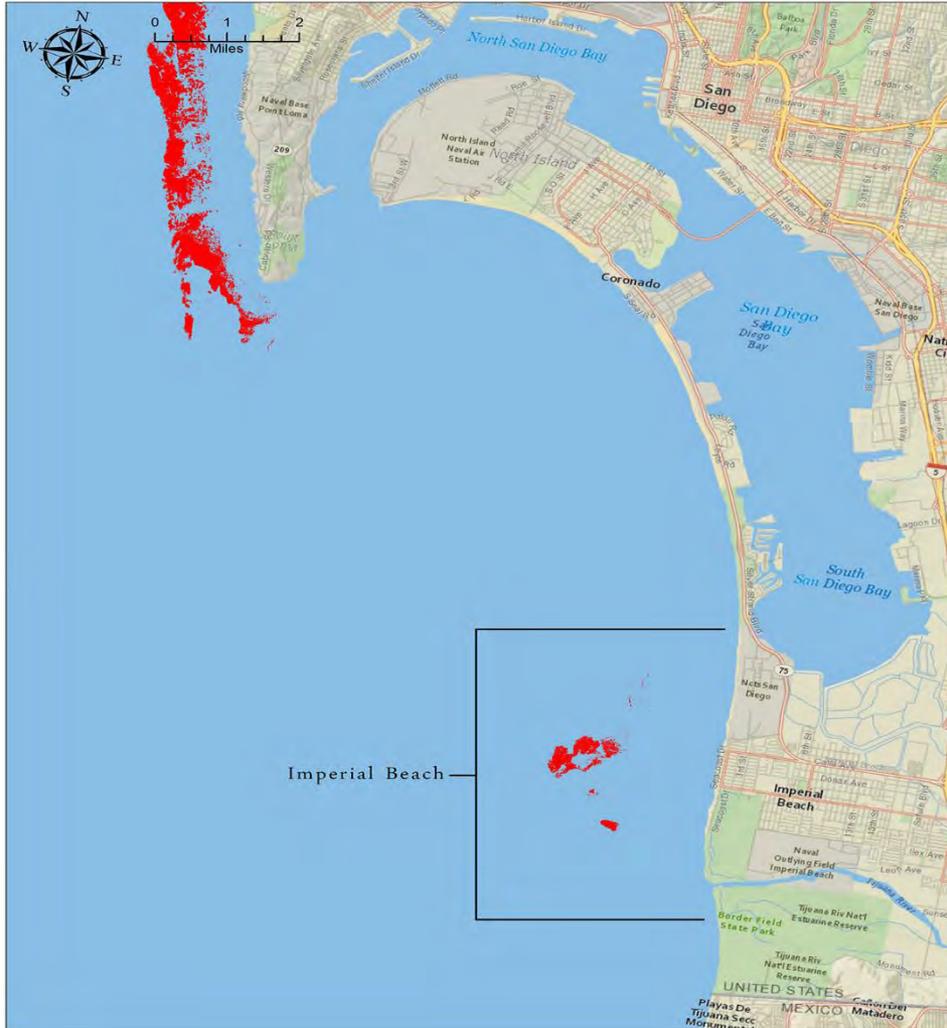
La Jolla



Point Loma



Imperial Beach



Appendix D.15

KELP BED		SURFACE CANOPY AREA IN 2019
North Laguna Beach	Smallest since:	2009
South Laguna Beach		2007
South Laguna		2006
Dana Point/Salt Creek		2006
Capistrano Beach		2005
San Clemente		2007
San Mateo Point		1998
San Onofre		2006
Horno Canyon		2011
Barn Kelp		2006
Encina Power Plant		2006
Encinitas		2005
Cardiff		2005
Solana Beach		1983

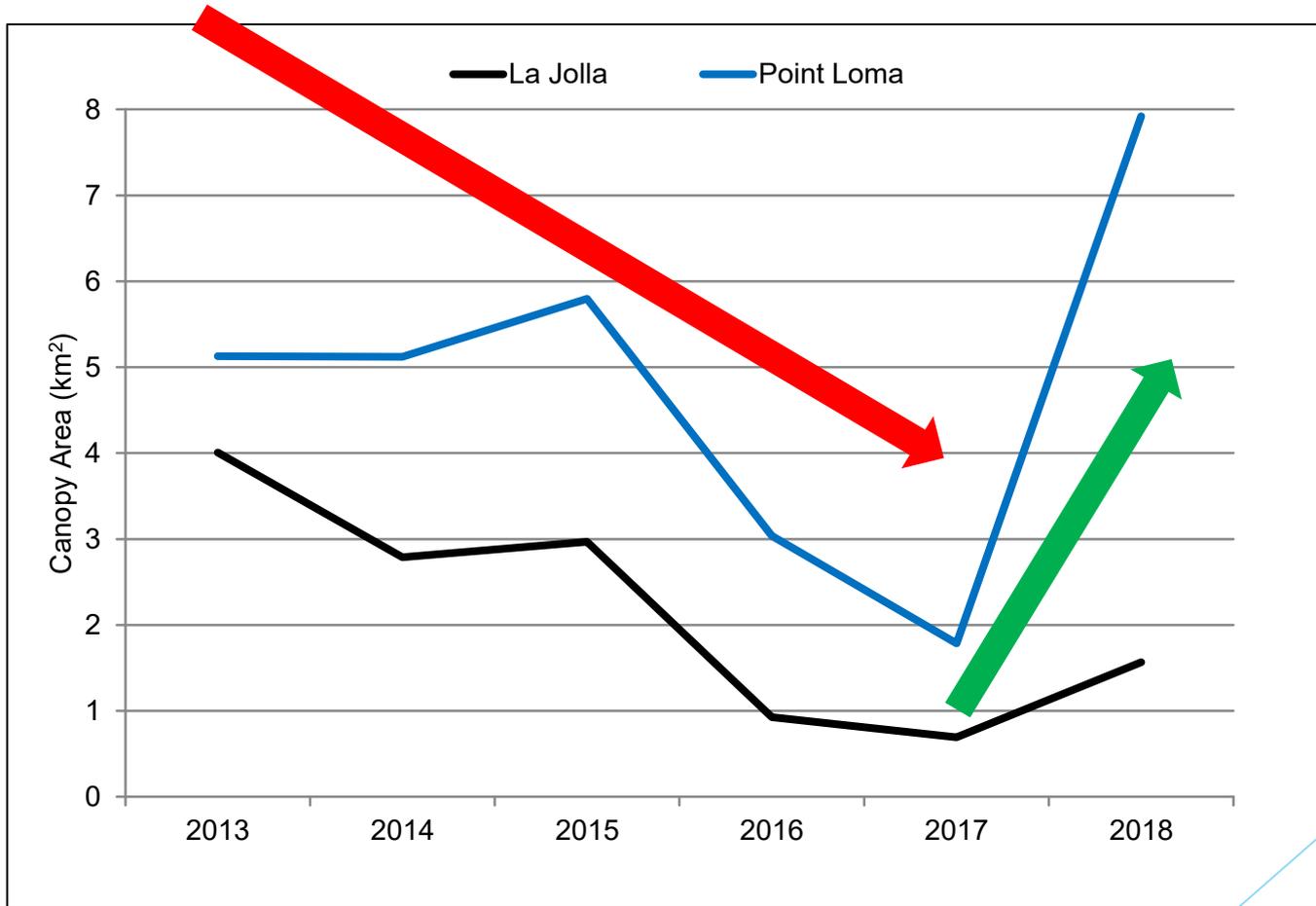
FACTORS AFFECTING KELP BEDS



Parnell, Dayton, Riser & Bulach. 2019. Evaluation of anthropogenic impacts on the San Diego coastal kelp forest ecosystem (2014 to 2019): final report.

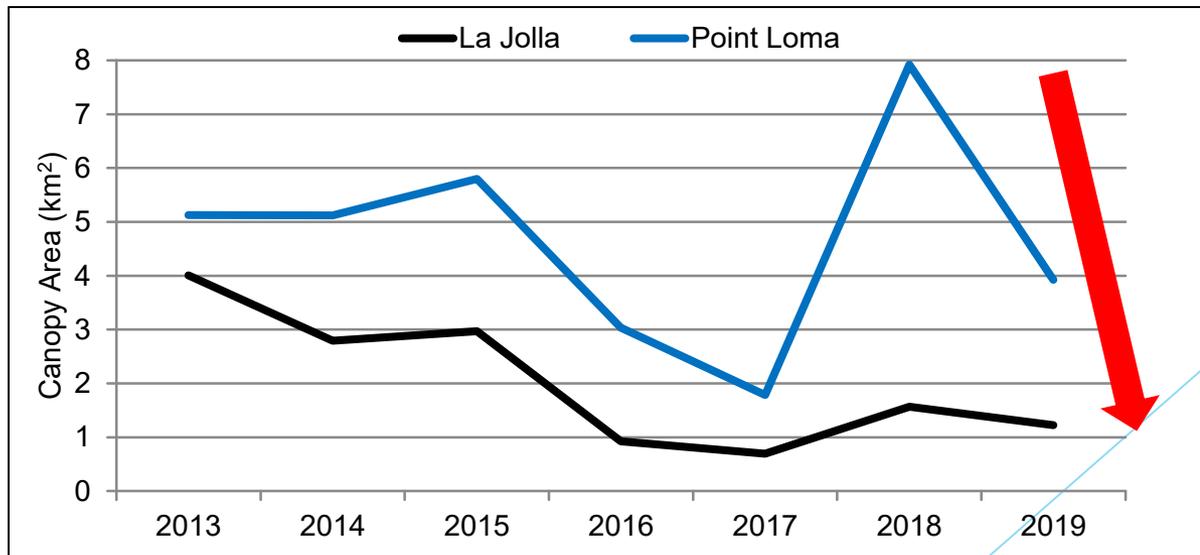
- ▶ SoCal kelp forests subjected to severe temperature and nutrient stress from late 2013 through spring 2017
- ▶ BLOB present during 2014-2015 = anomalously warm surface waters across much of Northeast Pacific Ocean
- ▶ Strong El Niño occurred during fall 2015 and winter of 2016 - just as the BLOB dissipated
- ▶ El Niño/BLOB combo caused longest and warmest period ever observed in 103-year SST time series at Scripps Pier
- ▶ Spring upwelling in 2017 and 2018 brought cool, nutrient-laden waters onto SoCal inner continental shelf creating favorable conditions for giant kelp regrowth

LA JOLLA & POINT LOMA KELP BEDS 2013-2018

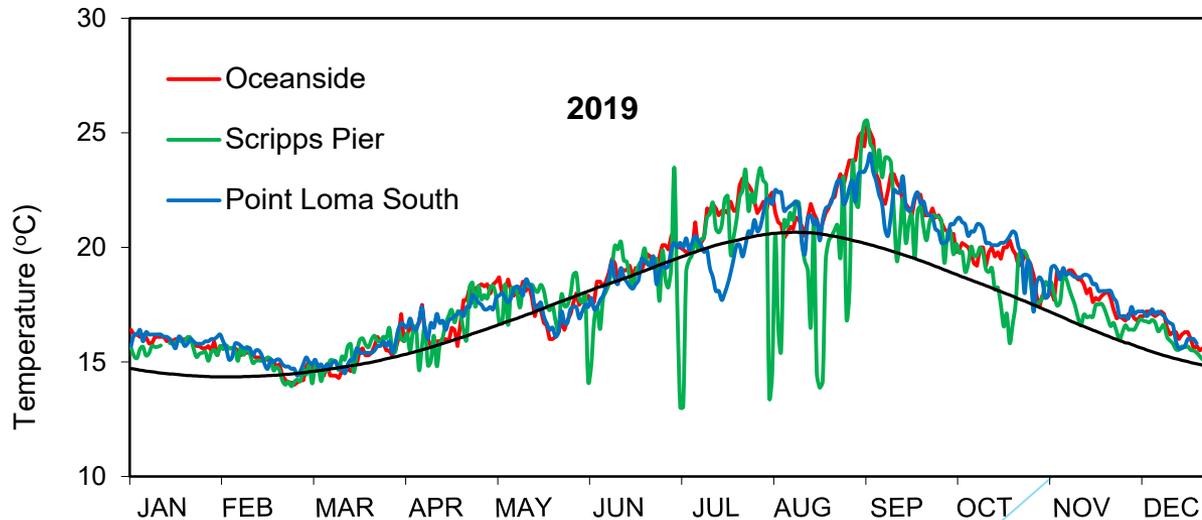
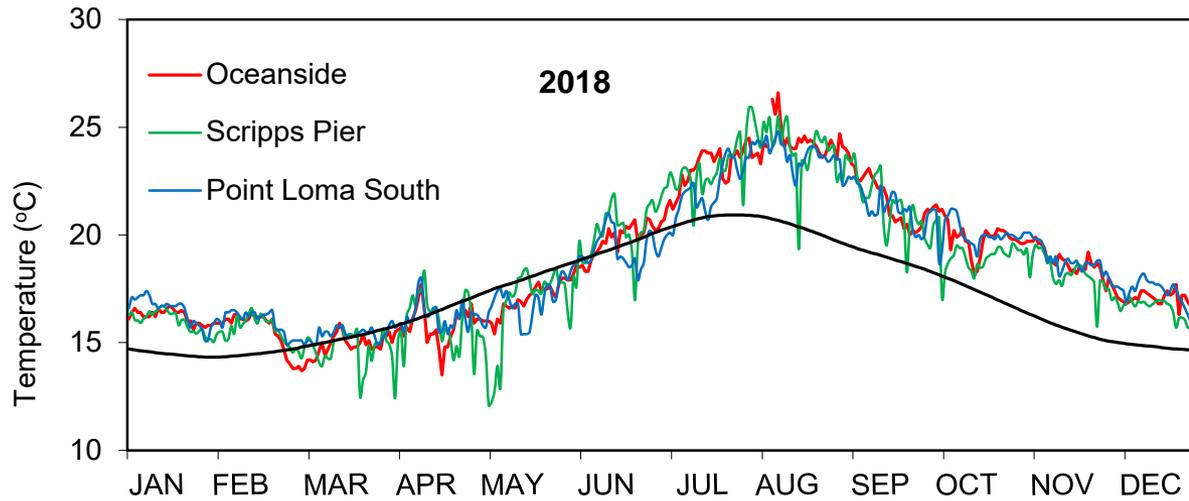


What Happened In 2019?

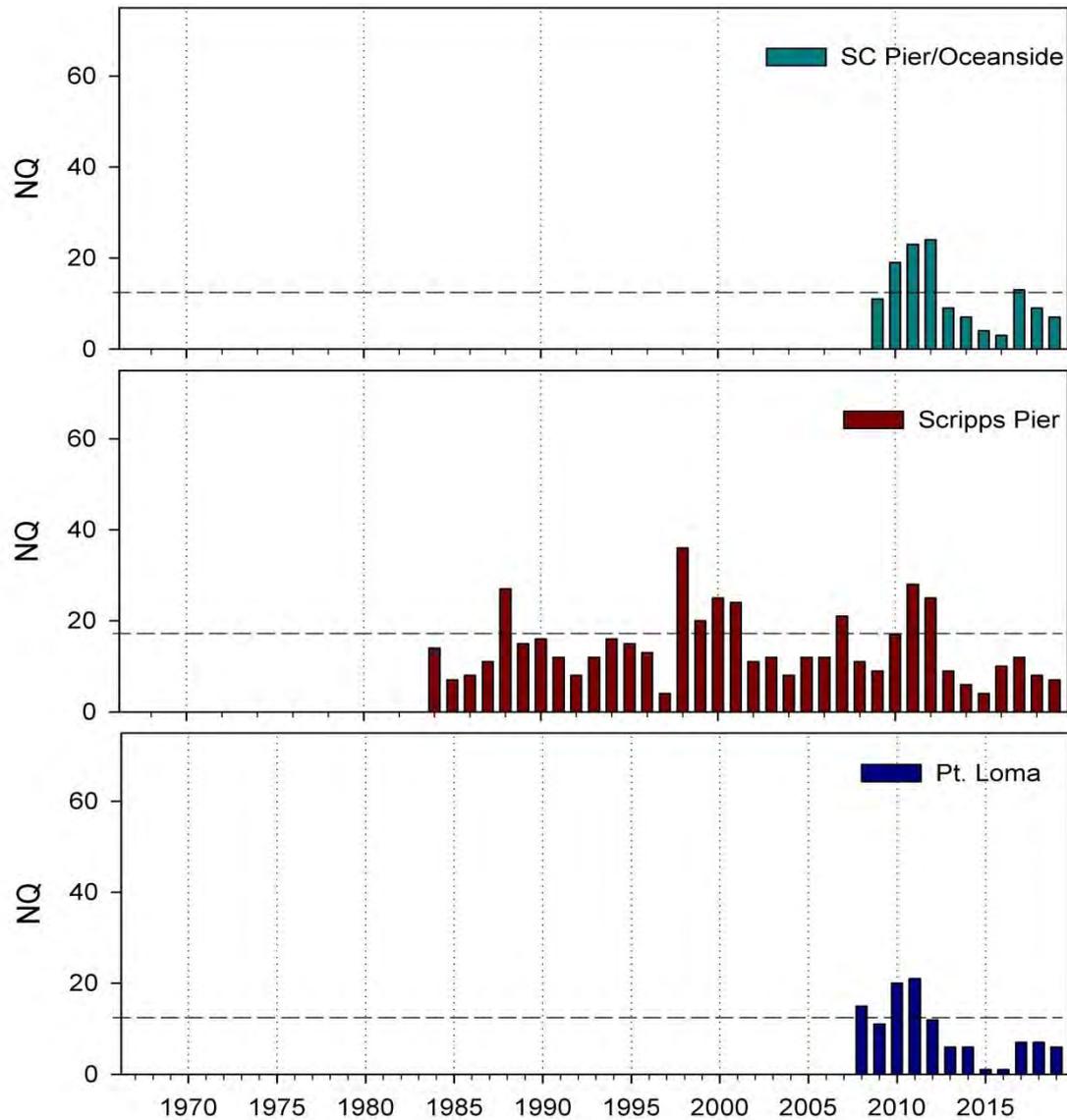
- ▶ In 2018, Region Nine Kelp beds had reached their maximum size in March or June (overflight data)
- ▶ Anomalously warm surface layer present during summer 2018 (Parnell et al, 2019)
- ▶ No surface canopy present throughout most of Region Nine in September or December 2018 (overflight data)
- ▶ Very little surface canopy present throughout most of Region Nine in March or June 2019, and almost none in September or December (except La Jolla and Point Loma)



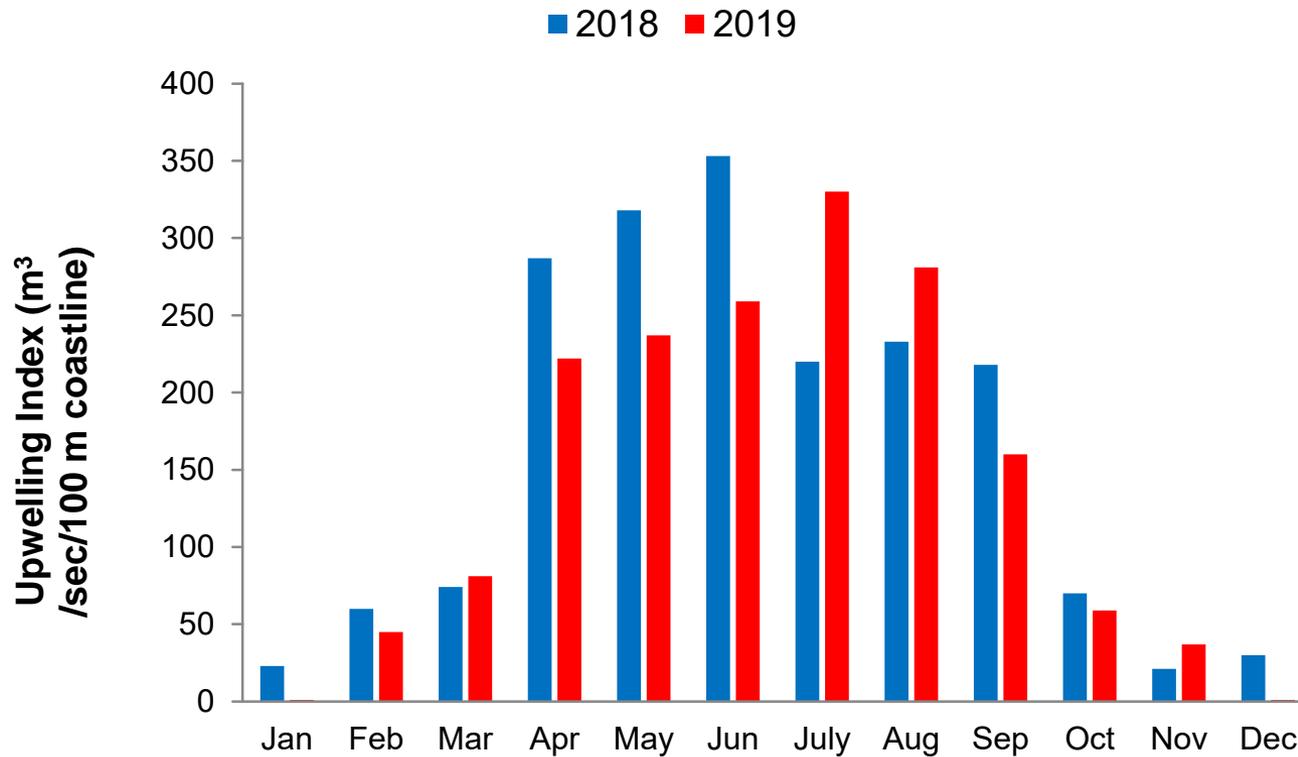
SST Values 2018 Versus 2019



Nutrient Quotient Index Values



Monthly PFEL Upwelling Index



Environmental Indices

- ▶ ENSO – continued in warm phase in 2019
- ▶ PDO – neutral in 2018, but warm regime in 2019
- ▶ NPGO – strongly negative from 2017 through 2019 indicating lower productivity

CONCLUSIONS



Conclusions

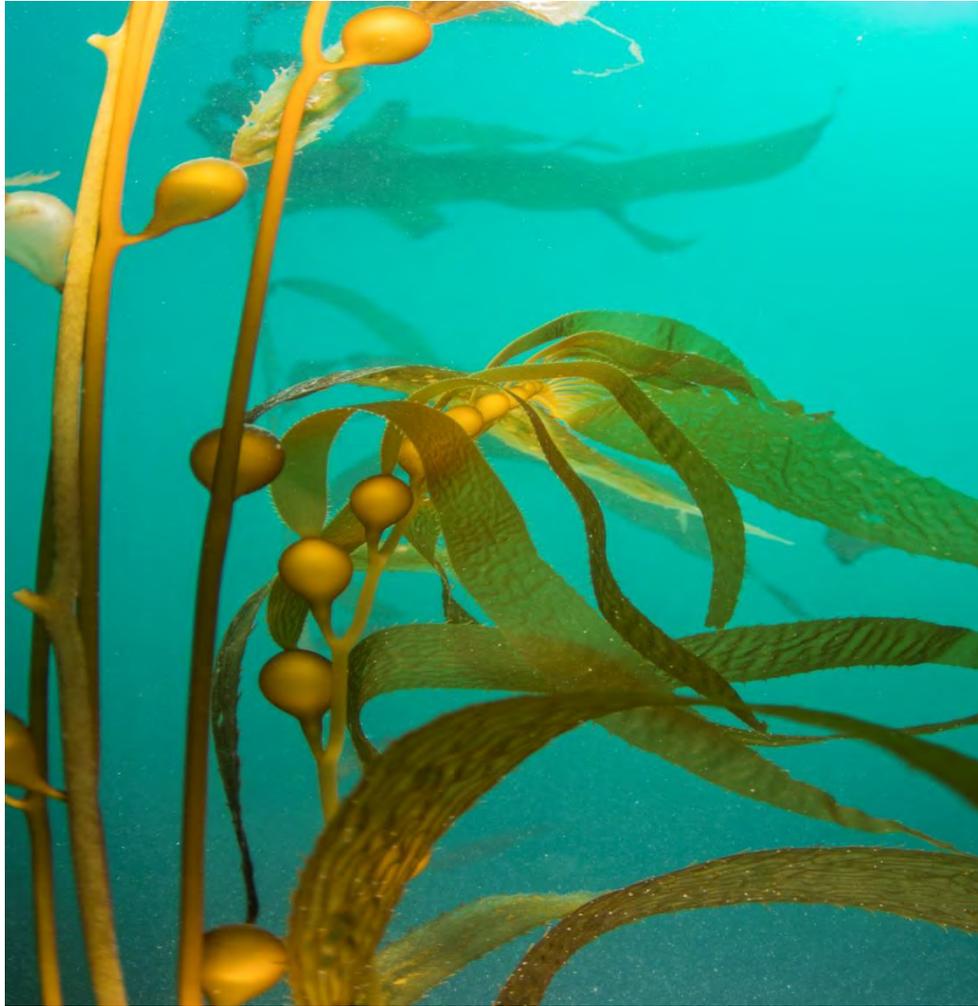
- ▶ 2019 was a bad year for kelp
 - ▶ 53 % decrease in total surface canopy for Region Nine
 - ▶ All kelp beds with visible surface canopy in 2018 decreased in size in 2019
- ▶ 10 kelp beds disappeared
- ▶ High SST values continued in 2019 (higher than normal during March, April, and May)
- ▶ Nutrient quotient values lower in 2019
- ▶ Monthly upwelling index values lower in 2019 during April, May, and June

PREVIEW OF 2020



Kelp Bed	April 15 estimated canopy	July 5 estimated canopy
North Laguna Beach	1.0	2.5
South Laguna Beach	-	0.5
South Laguna	-	-
Salt Creek-Dana Point	0.5	-
Capistrano Beach	0.5	-
San Clemente	-	-
San Mateo Point	-	-
San Onofre	-	-
Horno Canyon	-	-
Barn Kelp	-	-
Santa Margarita	-	-
North Carlsbad	NI	-
Agua Hedionda	-	-
Encina Power Plant	NI	-
Carlsbad State Beach	NI	-
Leucadia	NI	-
Encinitas	-	-
Cardiff	-	-
Solana Beach	NI	-
Del Mar	-	-
Torrey Pines Park	-	-
La Jolla Upper	1.0	1.0
La Jolla Lower	1.0	1.0
Point Loma Upper	2.5	3.0
Point Loma Lower	3.0	3.0
Imperial Beach	-	-

QUESTIONS?



Proposal to study timing on *Sargassum horneri* removal as a technique for eradication

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Background: A report, published in the journal Nature in April 2020, identifies nine key components that are essential to rebuilding the oceans: salt marshes, mangroves, seagrasses, coral reefs, kelp, oyster reefs, fisheries, megafauna and the deep ocean. The authors recommend a range of actions including protecting species, harvesting wisely and restoring habitats (Duarte et al. 2020).

I have been doing ocean restoration work in Orange County since 2002 under the umbrella of several different organizations. In those 18 years, I have trained over 350 volunteer divers to help me with the tasks of restoring giant kelp (2002-2012), restocking and monitoring green abalone (2013-2015) for a study that was published in CDFG journal in 2017, and surveying green abalone intertidally and subtidally (2017-present) to map and calculate density of abalone in Orange county. We are currently raising green abalone for restocking in 2021-2025. In 2009, I started Get Inspired, a nonprofit 501c3 charity organization to continue this work and have partnered with CDFW on several projects. I have taught more than 12,000 students to grow abalone, white seabass, and kelp in custom classroom nurseries for outplanting to restore species along the coast of Orange County. Additionally, through a partnership with Hubbs SeaWorld Research Institute, I started the “seabass in the classroom” program (also in partnership with CDFW). All of these programs have integrated classroom lessons which accompany the culture systems. I go into the schools to teach topics in aquaculture, kelp forest ecology, and ocean chemistry.

Saragssum horneri an invasive species native to Japan and Korea, is now prevalent along the mainland of Western North America from Baja California to Santa Barbara, and at three of the five Channel Islands (Anacapa, Santa Cruz and Santa Barbara) (Marks et al. 2015). Its continued expansion in the eastern Pacific may pose a major threat to the sustainability of native marine ecosystems. (Marks et al. 2015). Marks et al. 2017 recommends that *Sargassum horneri* be tracked, monitored, and studied so that impacts to resources can be assessed and potential management actions, such as eradication, can be evaluated. Kaplanis et al. 2016 mentioned that the rapid and uncontrolled spread of Sargassum has serious implications for their expansion

along the west coast of North America and that the ecological and economic consequences of these invasions require further research.

Forests of *Macrocystis pyrifera* naturally wax and wane throughout the natural cycles experienced along our coast. However, now with the invasion of Sargassum, there is competition for space for kelp to make its comeback resulting in a lack of food for grazers such as abalone which are already experiencing challenges in their recovery. Most herbivores do not prefer Sargassum as a food choice and this perhaps has led to its success (Marks et. al 2020). Through personal observations, Sargassum seems to whether the warm water events and large swells produced from the recent El nino event much better than the giant kelp. It forms such dense forests that fish cannot even swim through it, also limits light penetration to the reef further inhibiting competitors. In some cases, there has been a shift, since our kelp restoration activities, from a *Macrocystis* forest with healthy understory of other alga and encrusting organisms to a desolate Sargassum covered reef. I have an emotional and personal interest in the kelp forests of Orange County, having spent 12 years of my life restoring them. Wheeler North once told me that, “You don’t just go in and restore the kelp and then walk away, it’s going to need to be managed over time” I believe that figuring out the best strategies for managing *Sargassum horneri*, especially, after the devastation of a warm water event, would be useful for the State of California. Up to this point, projects to add to this knowledge based have not met with success due to timing (Marks et al. 2017). So, to help in that effort, I am requesting permission to conduct a pilot project to study the timing of removal of *Sargassum horneri* as a method for controlling it. Just as we manage our terrestrial forests, we may need to start managing our kelp forests. This may especially be helpful to manage the return of *Macrocystis* after a warm water event or significant disruption to the ecosystem. My theory is that if it is cut and/or scrubbed off the reef at its base **just before or when** giant kelp is recruiting, the kelp will be able to regain its “real estate” on the reef and the Sargassum will be outcompeted. To ensure there is “room” on the reef for the kelp to recruit and because we know that Sargassum can recruit throughout the fall and early winter, we will test the timing of eradication to determine the best time for removal for reestablishment of the giant kelp. These tactics may then be employed in the future after a devastating event such as an El Nino, to bring the ecosystem back into balance faster.

Proposed project: Get Inspired team requests permission to conduct this experiment in Crystal Cove SMCA. We have seen a regime shift on this reef. It was once a lush garden of native alga and has recently become a Sargassum pasture with an articulated coralline understory. The premise of this project is that *Macrocystis* has lost its “real estate” or it’s position in this reef community. By timing the removal of *Sargassum* with giant kelp recruitment, we may see the regime shift back to a kelp dominated forest. This SMCA has the least amount of protection, allowing for the take of finfish, lobster and sea urchin. Over the last 5 years, we have observed a loss of diversity of algal species. The recent *2019 Status of the Kelp Beds* report from MBC Aquatic Sciences showed 98% kelp loss in this SMCA. Throughout the project period, we will monitor ocean conditions such as: sea surface temperature, kelp sporophyll release periods and kelp recruitment events whilst conducting targeted Sargassum removal to determine the best time to remove Sargassum to allow for kelp recruitment back on the reef and if it has an impact on

algal composition on the reef. We will notify CDFW the coordinates of sites before we begin as initial surveys will be required. Although “spreading” Sargassum is really no longer a threat in Southern California as it is ubiquitous, care will be taken during the reproductive season to remove the reproductive season to remove the whole plant. During the non-reproductive season, April-October, we will just pull the invasive algae.

Method:

Two study areas will be chosen where we can set up 4 treatment sites in each one. All the treatment sites will be 10m x10m in size. The study area will have *Macrocystis pyrifera* and *Sargassum horneri* present. A HOBO temperature logger will be installed in the study area. This study will be over the expanse of the SMCA in areas where kelp once grew and has disappeared.

All sites will have an initial survey of Sargassum and other native algae with band transects or quadrats (depending on density).

Each dive will have a scoring system for the sporophyll release, kelp recruitment, and Sargassum development stage. We will survey the study area during each dive to determine the spore release quotient on the giant kelp. We will observe each plant give it a score and tally up the scores at the end of the dive. Development stage of Sargassum will also be noted on each dive.

Spore release

- 1- Indicates sporophylls are golden brown and the same color as the kelp blades
- 2- Indicates sporophylls are smooth or darker in color
- 3- Indicates sporophylls are smooth texture, darker in color, and have necrotic ends indicating max spore release is taking place.

Kelp recruitment

- 0- Indicates no sign of kelp recruitment
- 1- Indicates spade shaped brown kelp recruits on the reef (species unknown)
- 2- Indicates giant kelp recruits confirmed on reef site

Once a #2 score is confirmed band transects or quadrats will be conducted to count recruits and determine density. Same treatment will be done on the control site.

Sargassum Development stage (as described by Miller and Engle 2009)

- 1- Fern-like stage <5cm
- 2- Immature, no receptacles
- 3- Fertile, actively reproducing
- 4- Senescent, after reproduction

We will set up 2 control sites in Laguna Beach SMR where no Sargassum removal will occur

Treatment site 1 will have continuous removal of sargassum through the study period

Treatment site 2 we will remove Sargassum starting when there is a dip in sea surface temperatures below 15 C ~December 2020

Treatment site 3 we will remove Sargassum starting in January 2021

Treatment site 4 we will remove Sargassum starting in February 2021

Importance and Benefit: Kelps are a vital California resource and an essential component to our Eastern Pacific ecosystem as seen in the recent Northern California kelp ecosystem collapse. We rely on them to sustain us we use them for fishing, diving, and we have thought enough to protect them with no take zones to allow the ecosystems to flourish. They also are an important habitat and food source for reef species. These critical habitats are facing more and more threats. From warm water events to urchin invasions and invasive species competition, we need to know how to effectively and quickly reestablish these habitats to sustain them for as long as we can. Just like we manage our forests on land, we should be managing the health of these important California ecosystems. We believe that this work is essential to the recovery of our kelp forests and for the management of our kelp ecosystem that will benefit everyone and we respectfully request your permission to investigate these methods.

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Marks LM, Reed DC, Obaza AK. Assessment of control methods for the invasive seaweed *Sargassum horneri* in California, USA. *Manag Biol Invasion*. 2017;8(2): 205–213. 10.3391/mbi.2017.8.2.08

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MBC Aquatic Sciences 2020. *2019 Status of the Kelp Beds Orange and San Diego Counties Prepared for the Region Nine Kelp Survey Consortium*. https://1drv.ms/u/s!AkLZpj2SiR6xpG_MWqg-Hs8Rqsuo?e=ouAaVG

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Presence of *Sargassum horneri* at Todos Santos Bay, Baja California, Mexico: Its Effects on the Local Macroalgae Community

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Abstract

To describe the annual cycle of *Sargassum horneri* in Mexican waters, we selected two sites differing in their degree of wave exposure and sediment type: Rincón de Ballenas (RB), and Rancho Packard (RP). From June 2009 to April 2010 we followed the seasonal changes in *S. horneri* density and biomass along two intertidal transects per site. The effects of this non-indigenous species on the local macroalgae community were assessed by comparing their species composition, density, biomass, species richness, and diversity index in quadrats with and without *S. horneri*. There were significant differences in *S. horneri* density and biomass between sites ($P < 0.001$). At RB the invasive alga density average was 2 ± 0.94 individual m^{-2} , with a mean biomass of 4 ± 0.95 g DW m^{-2} . At RP, *S. horneri* density average was 10 ± 0.96 individual m^{-2} , and mean biomass of 102 ± 0.97 g DW m^{-2} . At RB, the invasive alga promoted a significant reduction in the four selected structural variables, and the corticated macrophytes and the foliose functional forms were severely reduced. At RP, there were only marginally significant effects ($P = 0.06$) of *S. horneri* presence on the local macroalgae community, and higher density, biomass, and diversity values were found when *S. horneri* was present. Most of the functional forms were found, even if the invasive alga was present. At both locations, the highest biomass corresponded to the articulated calcareous functional form. These contrasting results could be due to the fact that the native macroalgae community has already been altered by the early invasion of *S. muticum*, with the most resilient species and functional forms remaining in place. One of the most important changes we noticed is the severe reduction of the canopy forming species at both sites.

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Keywords

Annual Cycle, Community Structure, Diversity Index, Functional Forms, Invasive Alga, Species Richness

1. Introduction

Non-indigenous species (NIS) represent a major concern to marine scientists as the ecosystem in which they arrive in is modified adversely. This change takes place through the ecological interactions they establish with the native species and through direct or indirect physical or chemical changes in the habitat itself. The speed of habitat change is also coupled to the stability or resilience of the ecosystem, so the impact can have different scales in space and time [1].

While experimental work supports the idea that diverse communities show greater resistance to invasion, it is not clear if this results from resource use complementarity, or from an increasing occurrence of suppressive species in more diverse communities [2]. To understand the mechanisms driving this response, interest has shifted from species richness to the functional roles that species or groups of species play. Functional groups are defined as non-phylogenetic grouping of species that perform similarly in an ecosystem based on a set of common biological attributes. Functional groups can be defined in relation to either the contribution of species to ecosystem processes, such as carbon or water cycling, or the response of species to changes in environmental variables, such as climatic variables or disturbance [3]. The number and identity of functional groups within a community may dictate the level of invasibility, implying that the invasion of a coastal habitat will only be promoted through loss of a whole functional group rather than the loss of one or a few members of that group [2].

As marine ecosystems are relatively open, with fewer limits than terrestrial systems to organism dispersal and energy flow, the irreversible impacts of exotic species have profound consequences on ecological systems [2]. Macroalgae are considered to be especially worrying NIMS (non-indigenous marine species) as they may alter ecosystem structure and function by monopolizing space, developing into ecosystem engineers, changing food webs, and spreading beyond their initial point of introduction through efficient dispersal capacities [4] [5]. The success of a non-indigenous species depends on its mode of reproduction, growth rate and dispersive potential [6] [7].

The fucoid genus *Sargassum* is monoecious, highly fecund, and possesses vesicles that allow the reproductive fronds produced annually to drift with currents and inoculate new locations [7]-[9]. Due to its ability to colonize hard and soft substrata, the total area of marine sediments open to occupation by members of the genus *Sargassum* is vast, and cumulative habitat modification could be very significant [10]. The main barrier to colonization of the rock is the presence of algal cover [11].

Once established, these species can accumulate high biomass and thus become a strong competitor for space and light [12]. *Sargassum* invasions have significantly impacted the structure of indigenous algal communities in North America and Europe, through competitive displacement and/or exclusion [7] [13]. Several studies have reported the reduction of functional groups, like the thick leathery and coarsely branched algae and native understory algae through strong competitive interactions with adult individuals of *S. muticum* [7] [14]-[17].

Sargassum horneri is native from Asia, and distributed in Japan, Korea, Hong Kong (China), Chinese Taipei and China Mainland [18]. It was observed in Catalina Island, California, in 2003 [19]. In Baja California, well-established populations of this species were observed in Todos Santos Bay in 2007 [20], from where it had extended along the temperate waters of the Baja California Peninsula [21]. However, no description exists of the population structure of *S. horneri* in Mexican waters. For this reason, we decided to study the annual growth cycle inside the Todos Santos bay. We were also interested in assessing the ecological impact of this non-indigenous algal species on the structure of the local community of macroalgae. For this purpose, we selected two locations that differed in substrate type and wave exposure degree, and measured the seasonal influence of Sea Surface Temperature (SST), Photosynthetically Active Radiation (PAR), and air-exposure hours, on *S. horneri* density and biomass. Simultaneously, we determined the changes in the species composition, density, and biomass of the local macroalgae community. Algal species were classified into functional groups to identify if their number and types differed as a function of the presence or absence of *S. horneri*. We expected *S. horneri* to be better represented in the most exposed site, where its high density and biomass would result in a significant re-

duction of macroalgae, density, and biomass, and a change in species composition. We anticipated the loss or reduction of the canopy forming species, representing the more morphologically complex functional forms.

2. Materials and Methods

2.1. Study Site

The Todos Santos bay is located about 130 km south of the USA-Mexico border, on the northwest coast of the Baja California peninsula, at $\sim 31^{\circ}47'N$; $116^{\circ}43'W$ (**Figure 1**).

The NW oceanic boundary is defined by the ridge of a broad shoal between the Todos Santos islands and the mainland shoreline. The SW boundary is defined as the shortest distance between a prominent point, known as Punta Banda, and the Todos Santos islands, and is marked by a 6 km wide submarine canyon. The bay has a surface area of $\sim 240 \text{ km}^2$. Maximum depth within the bay is $\sim 100 \text{ m}$, except for the canyon, reaching to 400 m and draining down the continental slope [22].

Winds dominate the coastal circulation. Prevailing northwesterly winds, during spring and summer, drive water into the bay from the NW. Only during some winter storms and offshore Santa Ana conditions, water enters from the southwestern [23]. There is an apparent convergence zone within the bay, near the mouth of the Punta Banda estuary, along the eastern shore [24]. Sediment transport into and within the bay follows the same circulation pattern [25]. The bay is under the upwelling influence during periods of NW winds, a prominent feature of much of the Pacific coast of the USA and northern Baja California [26], and some authors have documented the influence of the local upwelling on water properties near the mouth during the springtime upwelling period [22] [27].

The two selected study sites, Rincón de Ballenas (RB), and Rancho Packard (RP), are located in the protected side of the Punta Banda peninsula, which is made up of shale and sandstone, forming high, almost vertical cliffs, which are interrupted locally by small pocket beaches made out of boulders [28]. Wave turbulence and littoral currents separate the material supplied by cliff erosion, allowing only grain sizes greater than 3.5ϕ (coarse fraction) to be deposited on the beach, while smaller sizes (fine fraction) are suspended and transported offshore [28]. Loose gravel predominates at Rincón de Ballenas and hard rock at Rancho Packard (**Figure 1**) [28].

The west coast of Baja California is characterized for having a mixed semidiurnal tidal cycle, with astronomical tides of higher amplitude during winter, season in which the strong storms originate bigger waves. The sum

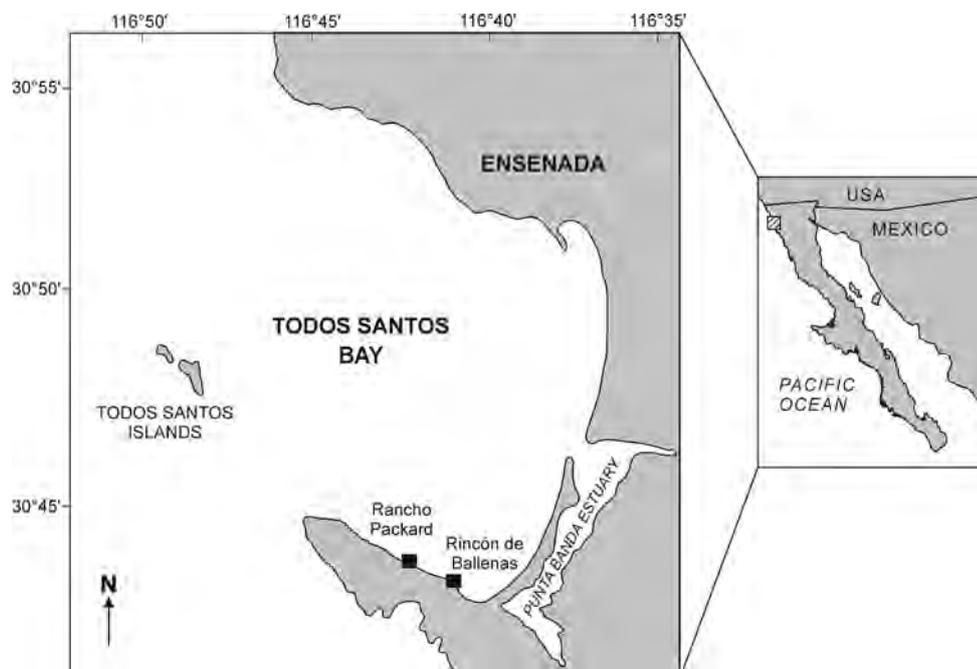


Figure 1. Map of the Baja California peninsula, indicating where the city of Ensenada is located. The inset shows the Todos Santos Bay and the two sampling stations in the protected side of the Punta Banda point.

of these two components results in a larger total wave amplitude. As a consequence, exposure hours are greater during winter. Of our two study sites, RP is more exposed to waves, than RB [29].

2.2. Sampling Design

Sampling took place from June 2009 to April 2010 during the Mean Lower Low Water tidal level (MLLW). Each season was represented by two months: June and July 2009, represented summer; October and November, autumn; December 2009 and January 2010, for winter, and March and April 2010, represented spring.

At each site we installed two transects perpendicular to the shore, separated by about 100 m. Their length and depth varied as a function of the topography. At RB, *S. horneri* was distributed between -0.2 and -0.8 m MLLW, corresponding to the low intertidal level; at RP, its distribution was between $+0.5$ and -0.2 m MLLW, in the middle and high intertidal levels (Figure 2).

2.3. Field Work

In order to cover the whole transect, samples were collected in the following manner; in every visit to the field we placed a 10 m rope along each transect, with marks every 0.5 m. At the beginning of every season, ten 0.25 m² quadrats were collected every meter starting at the 0 distance, and in the second seasonal visit, ten samples were also collected every meter, but starting at the 0.5 m mark.

Sampling was destructive, following the methodology described by [29]. Macroalgae were detached from the

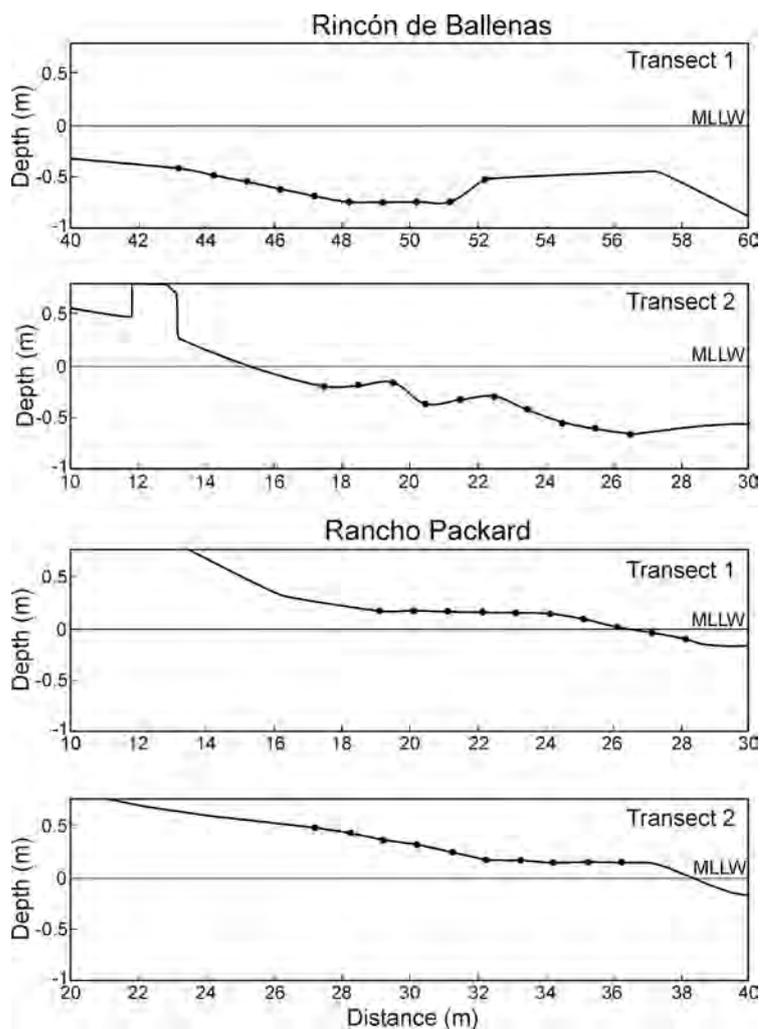


Figure 2. Vertical distribution of *S. horneri* at each of the established transects.

substrate by hand. All collected material was placed in labeled plastic bags, kept cool until arrival to the laboratory. Once in the lab, macroalgae were frozen until processing.

Sea Surface Temperature (SST) and Photosynthetically Active Radiation (PAR) data from May 2009 to May 2010 were downloaded from NASA's Ocean Color Satellite (<http://oceancolor.gsfc.nasa.gov/seadas/>). The use of the MODIS sensor provides a 4 km resolution, so the same data set was used for both sampling sites. For every site and sampling depth, we assessed the seasonal variations in tidal exposure, adding the number of hours that the sea level was lower than the selected reference level (Sea Level Laboratory, CICESE).

2.4. Laboratory Work

Macroalgae were defrosted and rinsed with fresh water to remove salts and sediment. Later, placed in plastic trays, and with tweezers, all epiphytic material, whether vegetal or animal, was removed. Algae were first separated into groups: Chlorophyta, Phaeophyta, and Rhodophyta, and then, all members of each group were identified at the species level. For this, histological cuts were performed, and tissue characteristics were analyzed under microscope. We used the taxonomic keys and classification system of [30]. Density was expressed as No. individuals of each species m^{-2} . Each species was oven dried at $60^{\circ}C$ for 24 hours, and weighed (± 0.1 g) to determine its biomass, expressed as g DW m^{-2} . Average density and biomass values were determined per site, depth, and month. Species were classified infunctional groups following [31], as: filamentous algae, foliose algae, corticated foliose algae, corticated macrophytes, leathery macrophytes, articulated calcareous algae, and crustose algae.

2.5. Data Analyses

S. horneri density and biomass data were analyzed using non-parametric statistics, since data did not followed a normal distribution. Significant differences between sites were explored with the U Mann-Whitney test. Differences among depths and months were analyzed with a one-way non-parametric ANOVA, Kruskal-Wallis. When non-significant differences between depth levels were found, such levels were pooled together to increase the power of the statistical tests [32].

The tendency between density and biomass with sea surface temperature, irradiance, and air exposure hours, was analyzed with the Spearman rank correlation test [33]. For all statistical analyses alpha was set at 0.05, and tests were run using the program STATISTICA 7 for Windows (2002).

2.6. Community Analyses

To determine community diversity, we used two attributes of community structure: species richness (S), and the Shannon-Wiener diversity index (H') [34]:

$$H' = -\sum_i \rho_i (\log p_i)$$

where ρ_i is the proportion of the total count arising from the i th species.

Both attributes were assessed when *S. hornerii* was present, and absent. Differences in H' under both conditions were tested with Hutchinson test [33].

We analyzed the spatial distribution of the functional groups when *S. horneri* was present, or absent, using their biomass values. For this, a non-metric MDS using the package "Vegan" for R platform was used [35].

3. Results

Following the annual cycle of maximum values during summer, and minimum values during winter, surface water temperature varied between $15.5^{\circ}C$ and $21.1^{\circ}C$, and irradiance between 22.8 and $55.2 \text{ mol}\cdot\text{m}^{-2}\cdot\text{d}^{-1}$. Air exposure hours varied by one order of magnitude between sites, with a total of 77 ± 0.5 hours at RB, and 685 ± 53 hours at RP. Winter was the season with more exposure hours at RB, and spring had the highest number of exposure hours at RP.

We found highly significant differences in annual mean *S. horneri* density and biomass between sites ($P < 0.001$), but not between tidal depths at any site ($P > 0.05$), and only at RP there were significant differences between sampled seasons ($P < 0.05$). At RB the invasive alga was present during October, November, and March, with an irregular distribution along the sampled depths. Its average density was 2 ± 0.94 individual m^{-2} , and av-

erage biomass, 4 ± 0.95 g DW m⁻². At RP, *S. horneri* was present during all sampled months, except June, but because of bad weather we could not collect samples in April. Also at this site, *S. horneri* vertical distribution was irregular. Average density was 10 ± 0.96 individual m⁻², with the lowest value during summer, 4 ± 0.9 individual m⁻², and the highest during autumn, 17 ± 0.98 individual m⁻². Average biomass for all the study period at RP was 102 ± 0.97 g DW m⁻², with the lowest values in summer, 12 ± 0.96 g DW m⁻², and the highest in spring, 292 ± 0.98 g DW m⁻².

At RB there were no significant correlations between *S. horneri* density and biomass with the environmental variables, but at RP, *S. horneri* density and biomass were negatively correlated with SST: ($r = -0.34$, $P < 0.001$), and ($r = -0.53$, $P < 0.001$) respectively; biomass was also negatively correlated with PAR ($r = -0.25$, $P < 0.05$). Air exposure hours yield no significant correlations with the biological data set either at RB or at RP.

3.1. Community Structure

A total of 39 macroalgal species was recorded during this study, of which 23 species were Rhodophyta, 11 Phaeophyta, and 5 Chlorophyta. The highest species richness corresponded to RB, with 29 species, while 25 species characterized RP. Highly significant differences between sites were found for macroalgae density ($P < 0.001$), and biomass ($P < 0.001$): 10 ± 0.9 individuals m⁻², and 66 ± 0.98 g DW m⁻², at RB, versus 14 ± 0.97 individuals m⁻², and 120 ± 0.96 g DW m⁻² at RP.

Of the 29 macroalgae species recorded at RB, the Rhodophyta were the most diverse, with 16 species, followed by the Phaeophyta with 9, and the Chlorophyta with 4 species (Table 1).

Table 1. List of macroalgae species found at Rincón de Ballenas, between June 2009 and April 2010, when *S. horneri* was present (+), or absent (-). Their functional form was determined according to [31].

Division	Species	Functional form	Condition
Chlorophyta	<i>Codium fragile</i>	Corticated macrophyte	(-)
	<i>Ulva californica</i>	Foliose	(-)
	<i>Ulva fasciata</i>	Foliose	(-)
	<i>Ulva nematoidea</i>	Foliose	(-)
Phaeophyta	<i>Colpomenia sinuosa</i>	Corticated macrophyte	(-)
	<i>Colpomenia tuberculata</i>	Corticated macrophyte	(-)
	<i>Dictyota flabellata</i>	Corticated foliose	(+)(-)
	<i>Dictyopteris undulata</i>	Corticated foliose	(+)(-)
Rhodophyta	<i>Petrospongium rugosum</i>	Crustose	(-)
	<i>Sargassum muticum</i>	Leathery macrophyte	(+)(-)
	<i>Silvetia compressa</i>	Leathery macrophyte	(+)
	<i>Sphacelaria californica</i>	Filamentous	(-)
	<i>Zonaria farlowii</i>	Corticated foliose	(+)(-)
	<i>Centroceras clavulatum</i>	Corticated macrophyte	(-)
	<i>Chondria californica</i>	Corticated macrophyte	(-)
	<i>Chondria decipiens</i>	Corticated macrophyte	(-)
	<i>Chondrocanthus canaliculatus</i>	Corticated macrophyte	(-)
	<i>Corallina officinalis</i>	Articulated calcareous	(+)
	<i>Corallina polysticha</i>	Articulated calcareous	(-)
	<i>Corallina vancouverensis</i>	Articulated calcareous	(+)(-)
	<i>Cryptopleura ramosa</i>	Foliose	(-)
	<i>Endarachne binghamiae</i>	Corticated macrophyte	(-)
	<i>Hypnea valentiae</i>	Corticated macrophyte	(-)
	<i>Jania crassa</i>	Articulated calcareous	(+)(-)
<i>Jania rosea</i>	Articulated calcareous	(+)(-)	
<i>Laurencia pacifica</i>	Corticated macrophyte	(+)(-)	
<i>Lithotrix aspergillum</i>	Articulated calcareous	(+)(-)	
<i>Pterocladia capillacea</i>	Corticated macrophyte	(+)(-)	
<i>Smithora naiadum</i>	Foliose	(-)	

Densities were higher for *Corallina officinalis*, and *Sargassum muticum*. The highest biomass values corresponded to the red alga *Corallina officinalis*, and to the green alga *Ulva fasciata*. The analysis per group shows that density was slightly higher for the red algae, 8.23 ± 0.73 ; followed by the brown, 7.55 ± 1.32 , and lower for the green algae, with 7 ± 0.86 individuals m^{-2} . With respect to biomass, the green algae showed the highest values with 163 ± 72.8 g DW m^{-2} , followed by the red, 75.76 ± 25.5 , and the brown algae, 41.3 ± 16.09 g DW m^{-2} . At RP, there were 14 species of Rhodophyta, 7 species of Phaeophyta, and 4 Chlorophyta (Table 2).

The red algae with highest density were *Corallina frondescens*, *Centroceras clavulatum*, and *Lithothrix aspergillum*, and the brown algae *Dictyopteris undulata*, and *Petroglossum rugosum*. The species with highest biomass were the red algae: *Lithothrix aspergillum*, *Corallina frondescens*, *C. pinnatifolia*, and *Centroceras clavulatum*. At the group level, density decreased from the brown, to the red, and the green algae: 12.57 ± 2.34 ; 10.66 ± 1.67 , and 6 ± 1.73 individuals m^{-2} respectively. The red algae had the highest biomass: 133.86 ± 32.9 , followed by the brown, 56 ± 7.1 , and the green, 8 ± 1.2 g DW m^{-2} .

3.1.1. Influence of *S. horneri* at RB

The most frequently present macroalgae had the greatest contribution in determining the community structure: *Dictyota flabellata*, *Dictyopteris undulata*, and *Sargassum muticum*, among the brown algae; *Corallina vancouverensis*, *Hypnea valentiae*, *Jania rosea*, and *Laurencia pacifica*, among the red algae. Peak density values were for *Corallina officinalis* and *Laurencia pacifica*, when *S. horneri* was present; when it was absent, highest density values were for *S. muticum*. When *S. horneri* was present, *Corallina officinalis*, and *Jania rosea* had the greatest biomass; when *S. horneri* was absent, peak biomass values corresponded to *Ulva fasciata*, and *Ulva californica*.

Table 2. List of macroalgae species found at Rancho Packard, between June 2009 and April 2010, when *S. horneri* was present (+), or absent (-). Their functional form was determined according to [31].

Division	Species	Functional form	Condition
Chlorophyta	<i>Codium fragile</i>	Corticated macrophyte	(+)
	<i>Codium hubbsi</i>	Corticated macrophyte	(-)
	<i>Ulva californica</i>	Foliose	(-)
	<i>Ulva nematoidea</i>	Foliose	(+)(-)
Phaeophyta	<i>Colpomenia sinuosa</i>	Corticated macrophyte	(+)(-)
	<i>Dictyopteris undulata</i>	Corticated foliose	(+)(-)
	<i>Dictyota flabellata</i>	Corticated foliose	(+)(-)
	<i>Petrospongium rugosum</i>	Crustose	(+)(-)
	<i>Sargassum muticum</i>	Leathery macrophyte	(+)(-)
	<i>Silvetia compressadeliquescens</i>	Leathery macrophyte	(-)
	<i>Zonaria farlowii</i>	Corticated foliose	(+)(-)
Rhodophyta	<i>Amphiroa zonata</i>	Articulated calcareous	(+)(-)
	<i>Centroceras clavulatum</i>	Corticated macrophyte	(+)(-)
	<i>Corallina frondescens</i>	Articulated calcareous	(+)(-)
	<i>Corallina pinnatifolia</i>	Articulated calcareous	(+)
	<i>Corallina vancouverensis</i>	Articulated calcareous	(+)(-)
	<i>Endarachne binghamiae</i>	Corticated macrophyte	(+)
	<i>Hypnea valentiae</i>	Corticated macrophyte	(+)(-)
	<i>Laurencia pacifica</i>	Corticated macrophyte	(+)(-)
	<i>Lithothrix aspergillum</i>	Articulated calcareous	(+)(-)
	<i>Mazzaella affinis</i>	Corticated macrophyte	(-)
	<i>Mazzaella leptorhynchus</i>	Corticated macrophyte	(+)(-)
	<i>Pterocladia caloglossoides</i>	Corticated macrophyte	(+)
	<i>Pterocladia californica</i>	Corticated macrophyte	(+)
<i>Pterocladia capillacea</i>	Corticated macrophyte	(+)(-)	

Macroalgae density showed significant differences between the *S. horneri* presence and absence condition ($P < 0.01$), with a mean of 2.977 ± 4.33 individuals m^{-2} , under presence condition, and 9.647 ± 2.232 individuals m^{-2} when *S. horneri* was absent. The same was true for macroalgae biomass ($P < 0.01$), with a mean of 18.125 ± 28.99 g DW m^{-2} for the invasive alga presence condition, and 76.428 ± 48.75 for the absence condition. Species richness (S) was higher when *S. horneri* was absent, with 28 species, than when the invasive algae was present, 13 species. Also, the diversity index (H'), was higher when *S. horneri* was absent, 0.884, than when it was present, 0.281 ($P < 0.0001$).

There were highly significant differences in macroalgae density through time ($P < 0.01$), with peak values between October and December, with values ranging between 8 and 13 individuals m^{-2} (Figure 3(a)). Macroalgae biomass also showed significant differences through time ($P < 0.01$), with a first peak in November, and a second peak in March, for both presence-absence conditions (Figure 3(b)). Species richness (S) was highest in June, under *S. horneri* absence, and in November, under *S. horneri* presence (Figure 3(c)). The species diversity

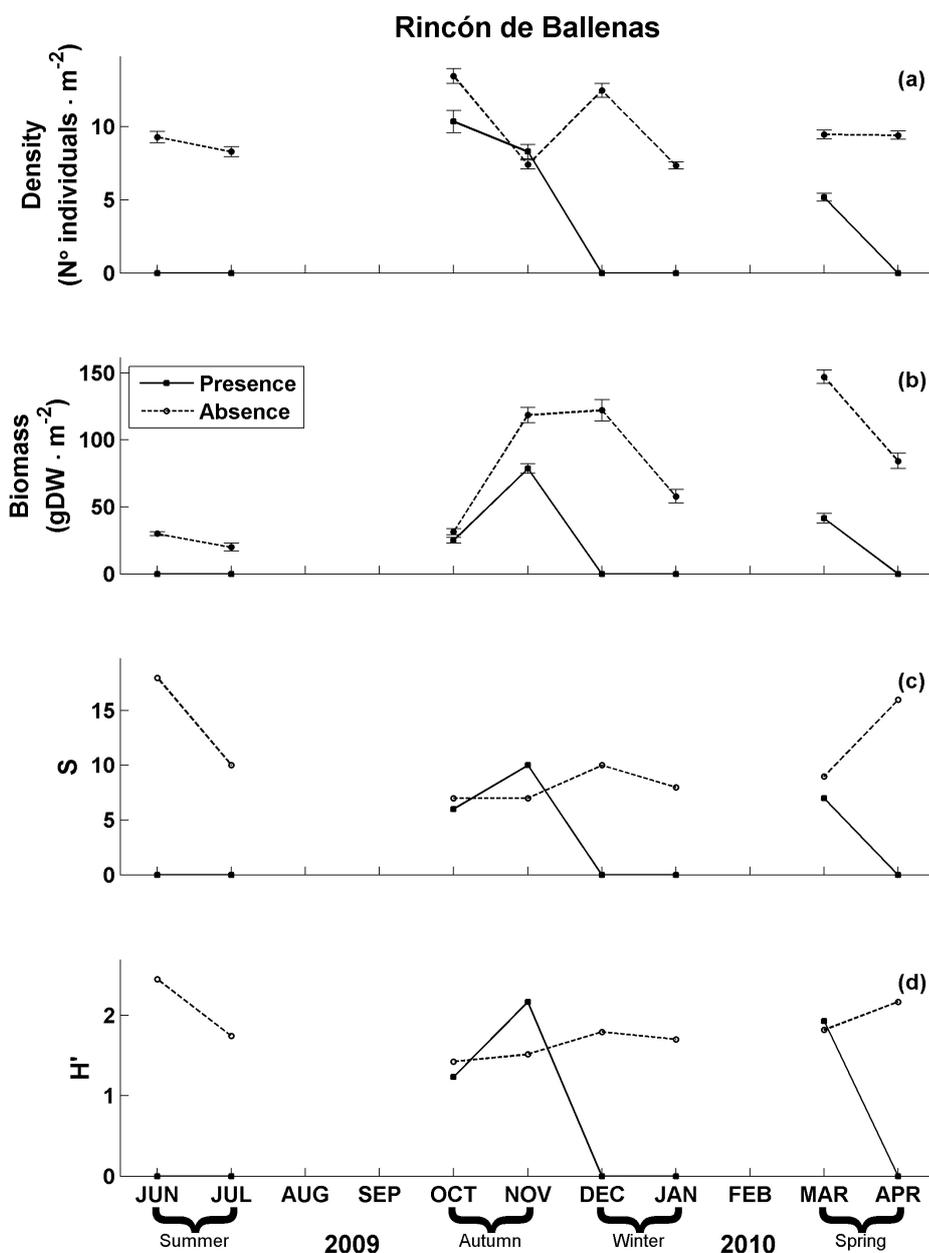


Figure 3. Seasonal variations in the selected structural variables in quadrats with and without *S. horneri* at RB.

index, H' , peaked in November when the invasive alga was present, and in June and April when it was absent, these differences being highly significant ($P < 0.001$) (Figure 3(d)).

The functional form of each species, and whether it was found when *S. horneri* was present (+), absent (-), or under both conditions (+) (-), is indicated in Table 1. It can be noticed that the corticated macrophytes and the foliose functional forms were the most affected, as the 14 species contained in these groups could only be found when the invasive alga was absent.

The MDS shows, to the left, a compact group formed by the low biomass values of all functional groups present; however, to the right, it can be noticed that the articulated calcareous reached the higher biomass values, regardless of whether the invasive algae was present, or absent (Figure 4).

3.1.2. Influence of *S. horneri* at RP

The macroalgae with the highest contribution to the community structure were: *Dyctiopteris undulata*, *Sargassum muticum*, *Zonaria farlowii*, and *Dictyota flabellata*, among the brown algae, and: *Lithothrix aspergillum*, *Centroceras clavullatum*, and *Corallina vancouverensis*, among the red algae.

When *S. horneri* was present, the species with more individuals per m^2 were: *Centroceras clavullatum*, *Corallina frondescens*, *Mazzaella leptorhynchus*, and *Dyctiopteris undulata*. When *S. horneri* was absent, *Lithothrix aspergillum* and *Petrospongium rugosum* were the species with highest densities. When *S. horneri* was present, the algae with the highest biomass values were: *Corallina frondescens*, and *Lithothrix aspergillum*. This last species, also had the highest biomass when *S. horneri* was absent, followed by *Centroceras clavullatum*.

The comparison of macroalgae density between the presence-absence conditions was slightly marginal ($P = 0.06$). Mean values were 9.641 ± 5.52 individual m^{-2} when *S. horneri* was present and 4.880 ± 6.88 individual m^{-2} when it was absent. The same significance level ($P = 0.06$) was found for the biomass comparison, with means of 74.489 ± 60.21 g DW m^{-2} under presence of the invasive alga, and 46.239 ± 82.33 g DW m^{-2} when it was absent. Species richness was similar when *S. horneri* was present, with 22 species, at when it was absent, 21 species. However, there were significant differences in the diversity index, with a higher value when the invasive alga was present, 0.740, than when it was absent 0.676 ($P < 0.005$). We also found that the selected variables showed changes as a function of time, with peak values in October, when the invasive alga was present, and in June when it was absent ($P < 0.001$) (Figure 5).

At this site, most of the species were present independently of the presence of *S. horneri*, with only three species, all with different functional forms, being affected by its presence (Table 2).

The MDS showed that, as in RB, the articulated calcareous group reaches the highest biomass values, followed by the corticated macrophytes (Figure 6).

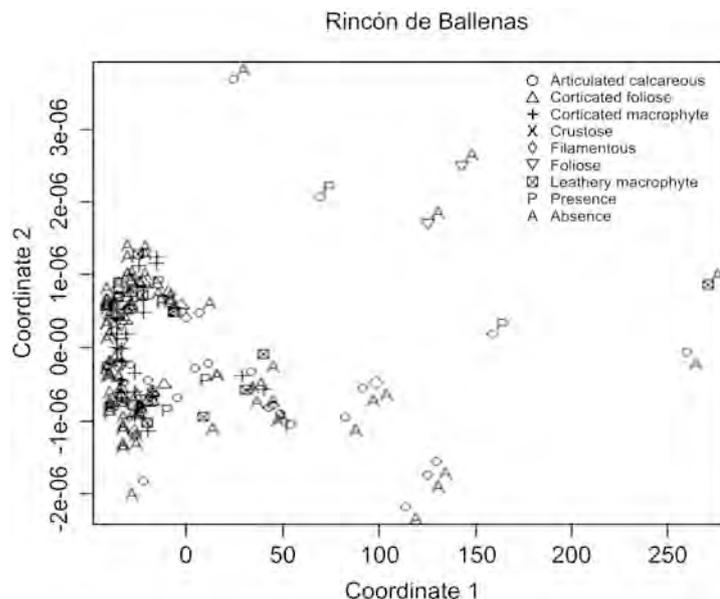


Figure 4. Distribution of the macroalgae functional forms at RB along the two coordinate principal axes.

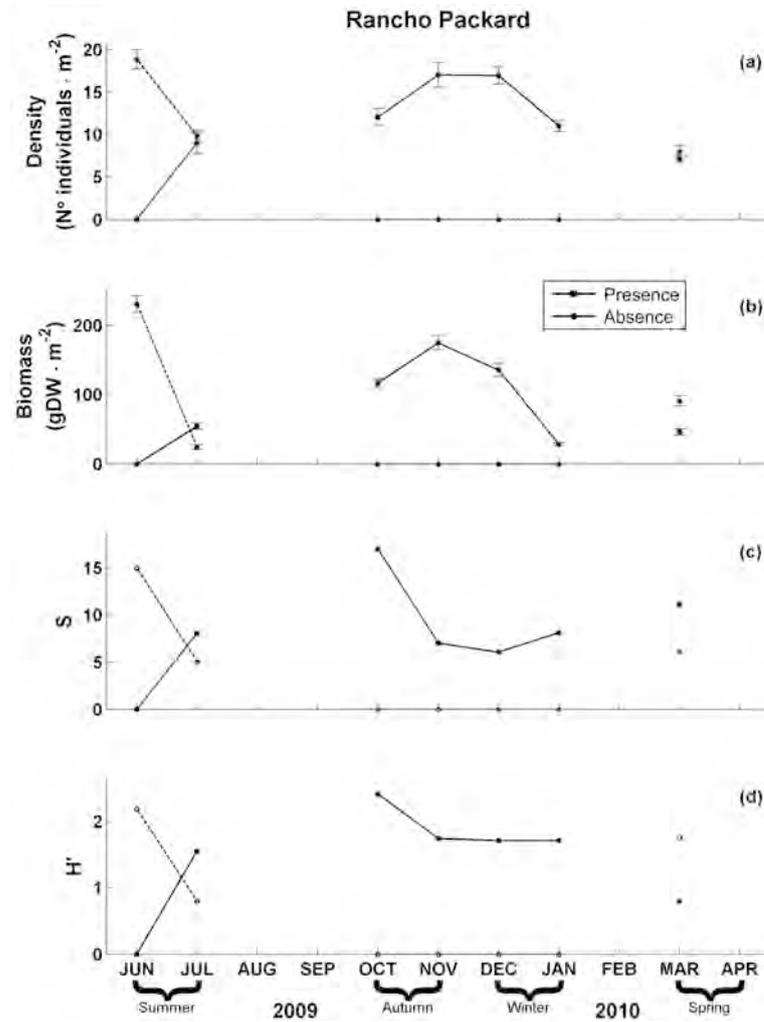


Figure 5. Seasonal variations in the selected structural variables in quadrats with and without *S. horneri* at RP.

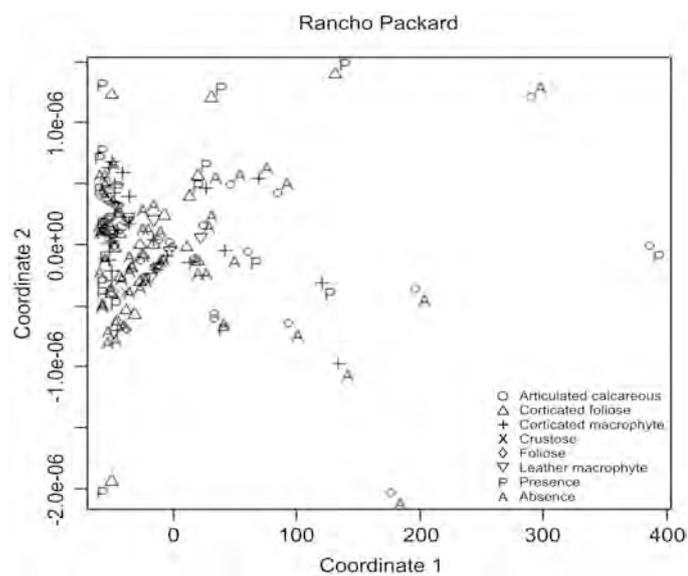


Figure 6. Distribution of the macroalgae functional forms at RP, along the two coordinate principal axes.

4. Discussion

Our results show that there are important site differences in the presence of *S. horneri* inside the Todos Santos bay and its effects on the macroalgae community. At Rincón de Ballenas, the invasive alga was not present all year round, and had low density and biomass values. However, its presence resulted in a significant reduction in the macroalgae density, biomass, species richness and diversity. The corticated macrophytes and the foliose functional forms were severely reduced by the presence of the invasive alga. On the other hand, at Rancho Packard, the presence of *S. horneri* was more continuous throughout the year, and this non-indigenous species reached high density and biomass values. Despite this, there only were marginally significant effects of its presence on the local macroalgae community, and higher density, biomass, and diversity values were found when *S. horneri* was present. Most of the functional forms were represented, even when *S. horneri* was present. Despite these notorious site differences, at both locations, the highest biomass corresponded to the articulated calcareous functional form.

With respect to sites differences, we know that RB is characterized by reduced wave exposure and soft sediment, represented by loose gravel. In contrast, at RP, wave exposure is slightly higher and the substrate is represented by solid rock. Although *S. horneri* has the rare ability to colonize both, hard and soft substrate [10], our data indicate that at the Todos Santos bay, *S. horneri* grows better on hard substrate. This is in agreement with [7], who at Limfjorden, Denmark, found a strong correlation between the cover of *S. muticum* and the presence of hard substrate. Although we did not find significant differences with depth, [7], found that the difference in cover between shallow, 0 - 2 m, and deep, 2 - 6 m, waters, was regulated by the amount of hard substrate.

Water movement has been considered a seasonally important variable which affects standing biomass, thallus size, morphology and, possibly, fertility [36] [37]. Although we did not made direct measurements of water movement, [29] used numerical simulation models to predict ocean surface waves inside the Todos Santos bay, and results of her study show a higher wave energy at RP, where we found the healthier populations of *S. horneri*. However, in Obama bay, Japan, [38] found that the *S. horneri* populations from the sheltered coast had longer primary laterals, and plants had higher weight, than those from the exposed shore. *In situ* measurements of wave exposure are needed at Todos Santos bay, to determine if our two study sites can be considered to be in a protected shore. For the northern coast of Spain [9], found that wave exposure was not significant for *S. muticum* growth and survivorship. In contrast, [39], for the foliose algae of South Wales, and [37], for the populations of *S. polyceratum* in Curaçao, found that foliose algae were more abundant where wave-action was greater and during the cooler months of the year.

Reference [39] also found that algal survival was greater and growth was faster under conditions of increased moisture, decreased emersion, and decreased temperatures and light regimes during low tide. However, in our study we could not find a significant correlation between *S. horneri* density and biomass with tidal exposure, despite the high number of exposure hours at RP. The fact that *S. horneri* grows at the high intertidal at RP (Figure 2), and that at this site exposure hours were greater during spring, could help explain the negative correlation between density and biomass with water temperature and irradiance, in agreement with [39]. The negative correlation between these two biological variables with irradiance could also be due to a high epiphyte load, as found by [40] for three species of *Sargassum* in Hawaii. The negative correlation with water temperature is in agreement with [38] [41] [42], among others, and is characteristic of temperate species.

The ephemeral nature of the individual patches of *S. horneri* at RB, could be understood using the physical and biological arguments that have been presented to explain the colonization and establishment patterns of the genus *Sargassum*: anomalously warm sea water temperatures and their subsequent effects on food web in the region [43]; the disturbance represented by the presence of sand and its negative impact on recruits survival [16]; the unsuitability of smaller stones, gravel and sand, as substrate for grown specimens [7], and highly localized propagule dispersal and settlement [44]. For *Gracillaria verrucosa*, [45] found than an exponential decline in settlement densities and short dispersal distances was partly due to the diffusive environment found in the shallow subtidal.

The strong seasonality that characterizes the genus *Sargassum* has been mainly attributed to sea water temperature, and photoperiod, with regional variations due to latitudinal gradients [43] [46]-[48]. “Autumn-fruiting type” and “spring-fruiting type” populations of *S. horneri* have been described for the Seto Inland Sea, Japan by [47]. For both populations it has been considered that the shortening of day length around the autumn equinox, is the possible cue to start the growth phase, characterized by the rapid increase in thallus length. Simultaneously,

water temperature starts its autumn reduction [47]. However, it is now considered that the difference in seasonality between these two populations does not reflect a phenotypic plasticity, but a genotypic difference [42].

The lifetime of the autumn-fruiting type is considered to have four phases, according to the rates of increase in length and morphogenetic stages: I—formation of early leaves, from December to May; II—differentiation of stems, from May to September; III—rapid elongation of stems and lateral branches, from September to December; IV—senescence phase, after December. In contrast, the spring-fruiting type has two growth phases and a senescence phase: I—from April to September; II—September to March, and III—senescence phase after March [47]. The selected populations of *S. horneri* at Todos Santos bay, corresponds to the spring-fruiting type, like the Japanese populations described by [38], and [47] for the Seto Inland Sea, and the populations of *S. filicinum*, now *S. horneri*, at Long Beach Harbor [19], and the California Channel Islands [49]. The sampled populations in Todos Santos bay show the lowest density and biomass values, when compared to reported values for *S. horneri* (Table 3).

When we analyzed the effects of the presence of the invasive alga on the local macroalgae community, we were surprised by the fact that at RB, where *S. horneri* was only present a few months and, showed low density and biomass values, there were significant differences between the macroalgae community structure when *S. horneri* was present versus when it was absent. The presence of the invasive algae resulted in significant reductions in macroalgae density, biomass, S and H' . On the other hand, at RP, where *S. horneri* had a more continuous presence throughout the year, and reached higher density and biomass values, the comparison between the macroalgae community structure under the presence and absence conditions was only marginally significant. The macroalgae showed higher density, biomass, and H' when *S. horneri* was present.

It seems that the macroalgae community at RB was more susceptible to invasion, than the one at RP. To understand the invasion process, it is necessary to analyze the number and identity of the functional groups present [2]. At RB there was a loss of functional diversity, with most of the species belonging to the foliose and corticated macrophytes functional groups being present only when *S. horneri* was absent (Table 1), while at RP, most of the species, and functional forms, remained when the non-indigenous alga was present (Table 2).

As indicated by [7], during an invasion process, the community structure is affected by the increasing abundance of the invasive alga, and by the changes in the remaining community. After the invasion of *S. muticum* in Limfjorden, Denmark, [7] found that members of the coarsely branched and thick leathery algae tended to decrease consistently over time, as a result of competition. Reference [2] found that canopy species, regardless of their density, suppressed invader biomass, while crustose species promoted invasibility. Turf and subcanopy species effects were similar to those of the canopy species, but less intense [2]. Competitive suppression is mainly due to light competition [17] [51] [52] with space competition becoming important in a later stage [52].

At RB, only two species belonging to the leathery macrophyte functional form (canopy) were present, with most of the corticated macrophytes (subcanopy), and all of the foliose (turf), being gone when *S. horneri* was present. In contrast, at RP, the macroalgae community seems to stand well the presence and abundance of the non-indigenous alga, as most species, and most functional forms remained present, regardless of the presence of *S. horneri*. It is important to note that what we refer to as the local macroalgal community has already being modified, as we found *S. muticum* at both sites. *S. muticum* persist under presence or absence of *S. horneri*, so no competition seems to exist between these two species, but this needs to be assessed in the field.

Table 3. Range of values for abundance, density, and biomass reported for *Sargassum horneri*. Authors are listed chronologically. ND = Not Determined.

Reference	Species	Site	Abundance (No. plants)	Density (No. individuals m ⁻²)	Biomass (g DW m ⁻²)
[38]	<i>S. horneri</i>	Obama Bay, Japan	ND	20	680 (sheltered) 431 (exposed)
[41]	<i>S. horneri</i>	Ohori, Corea	15 (October) - 68 (March)	ND	ND
[19]	<i>S. filicinum</i>	Santa Catalina Island, CA	>30 (April, exposed) 2 - 4 (April, sheltered)	ND	ND
[50]	<i>S. horneri</i>	Gouqui Island, South China Sea	25 (June) - 830 (August)	96 (June) - 3320 (August)	540 (August) 4420 (June)
This study	<i>S. horneri</i>	Todos Santos Bay, Mexico	ND	1 (July) - 10 (March)	3 - 78

The already altered macroalgae communities we found inside the Todos Santos bay, are dominated by the articulated calcareous, functional form with the highest biomass at both sites (**Figure 4** and **Figure 6**). This functional form corresponds to what [2] refer to as turf-forming species, which are recognized for being primary space-holders with limited vertical height (usually ~5 cm length). Algal turf has the ability to monopolize space and persist under a wide range of environmental conditions, and its thickness, rather than its cover, seems to be the most affected by the intensity of disturbance and smothering by sediments [53]. The rapid growth of turf-dominated assemblages provides its capability to compete for space and recover from disturbance [54].

5. Conclusion

Our results do not fully support our hypothesis. As in RP, where the highest density and biomass values of *S. horneri* were found, there was not the significant reduction in macroalgae density, biomass, S and H' we expected; on the contrary, density, biomass, and H' showed higher values when the non-indigenous alga was present (**Figure 5**). This unexpected result could be due to the fact that the native community had already been altered by the early invasion of *S. muticum*, with the most resilient species and functional forms remaining in place. One of the most important changes we noticed is the severe reduction of the canopy forming species at both sites, confirming the fact that the local macroalgae community has already been modified, in agreement with [7]. A long-term monitoring, with more study sites, is needed to fully comprehend the changes that the local macroalgae communities are experiencing along the Baja California peninsula.

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From: Jeff Maassen [REDACTED]
Sent: Thursday, October 1, 2020 4:29 PM
To: FGC <FGC@fgc.ca.gov>
Cc: Ashcraft, Susan@FGC <Susan.Ashcraft@fgc.ca.gov>
Subject: Sargassum Horneri--Request for Commercial Kelp harvest permit

Dear California Department of Fish and Game Commission,

Please see attached request packet for a commercial permit to harvest Sargassum Horneri.

Respectfully,

Lance Maassen

[REDACTED]
[REDACTED]

SARGASSUM HORNERI

Request for Commercial Kelp Harvest permit



Lance Maassen • October 2020

Dear California department of Fish and Game commissioners,

I am a Santa Barbara based Commercial Sea Urchin Fisherman and boat owner Over the last 35 years I have dive harvested within California's Subtidal waters for Sea Urchins from San Clemente Island to Fort Bragg. During this tenure I have collaborated and willingly shared information with CDF&W, UCSB, NOAA, Scripps, SDSU, USC, OPC and others to inform management and research and to hopefully ensure sustainable outcomes for Californias commercial fisheries.

I would like to request the issuance of a permit to Dive- harvest for the Invasive species "Sargassum Horneri". Over the past several years this species abundance appears to be spreading Northward has been observed to be over taking and choking out other indigenous species in the Southern California Bio region.

We are currently in discussions with San Luis Obispo based Kelp harvesting company "Kelpfulca" to collaborate in processing and distribution to explore opportunities utilizing Sargassum including utilizing in food as Seasonings, "Akamoku"(Soup), Beer, Soap and possibly a specialty fertilizer.

Pursuant to Title 14 regulatory compliance I would request some latitude and close collaboration with staff in tailoring some of the regulations and permit fees specifically towards the Hand harvesting of an invasive species which would be necessary to proceed. This would facilitate efficient scaling and enable measured ecological outcomes.

Thank You very much for your consideration,

Lance Maassen



f) All Other Species of Kelp.

(1) Applicant shall apply to the commission, outlining the species to be harvested, amount and location. The commission may set conditions and amount of royalty after review of the application.



Reference Videos:

Youtube video of Sargassum Horneri at Anacapa Island:

<https://www.youtube.com/watch?v=iqo9ASD5GAk>

<https://agris.fao.org/agris-search/search.do?recordID=JP2009005623>

1. § 165. Harvesting of Kelp and Other Aquatic Plants.

2.14 CA ADC § 165 BARCLAYS OFFICIAL CALIFORNIA CODE OF REGULATIONS

Term

Barclays Official California Code of Regulations [Currentness](#)

Title 14. Natural Resources

Division 1. Fish and Game Commission-Department of Fish and Game

Subdivision 1. Fish, Amphibians and Reptiles

Chapter 6. Fish, Commercial (Refs & Annos)

14 CCR § 165

§ 165. Harvesting of Kelp and Other Aquatic Plants.

(a) General License Provisions. Pursuant to the provisions of Section 6651 of the Fish and Game Code, no kelp or other aquatic plants may be harvested for commercial purposes except under a revocable license issued by the department.

(1) Who Shall be Licensed. Each person harvesting kelp and other aquatic plants for profit shall apply each year for a license on 2015 Kelp Harvesting License Application (DFW 658 Rev. 08/14) which is incorporated by reference herein. License applications and a list of laws and regulations governing the harvest of kelp and other aquatic plants (including maps depicting administrative kelp beds) are available on request from the department's Los Alamitos office at 4665 Lampson Avenue, Suite C, Los Alamitos, CA 90720.

(2) Cost of License. See section 6651 of the Fish and Game Code.

(3) Where to Submit Applications. Application forms, together with the fee authorized by Section 6651 of the Fish and Game Code, shall be submitted to the department's Los Alamitos office, 4665 Lampson Avenue, Suite C, Los Alamitos, CA 90720.

(4) License Limitation. All provisions of sections 6650-6680 of the Fish and Game Code, and sections 165 and 165.5 of the commission regulations shall become a condition of all licenses issued under this section to be fully performed by the holders thereof, their agents, servants, employees or those acting under their direction or control.

(b) General Harvesting Provisions.

(1) Weighing of Kelp. A kelp harvester shall determine the weight of harvested kelp or other aquatic plants upon landing or delivery to the harvester's place of business. The harvester may determine the weight of harvested kelp or other aquatic plants by either direct weighing with a state certified scale or a volume conversion that has been approved by the department. If the weight is determined by a certified or licensed weighmaster, the harvester shall obtain a receipt and maintain the receipt in the landing record required under subsection (b)(3) below.

(2) Harvesting Records.

(A) Every person harvesting kelp and other aquatic plants and licensed pursuant to Section 6650 of the Fish and Game Code shall keep a record of the following:

1. Category of plants harvested as defined in subsections 165(c), (d) and (e).
2. The wet weight of harvested kelp or other aquatic plants recorded in pounds or tons (1 ton = 2000 lb).
3. Name and address of the person or firm to whom the plants are sold, unless utilized by the harvester.

(B) The record shall be open at all times for inspection by the department.

(3) Landing Records. Records of landing shall be prepared by all harvesters licensed pursuant to Section 6650 of the Fish and Game Code. Records of landing shall be made in triplicate using Kelp Harvester's Monthly Report forms FG 113 (Rev. 1/97, see Appendix A) and FG 114 (Rev. 1/07, see Appendix A).

(A) The landing records shall show:

1. The wet weight of all aquatic plants harvested in units as defined in subsection (b)(2)(A)2. above.
2. Name and address of harvester.
3. Department of Fish and Wildlife kelp harvester number.
4. Report period, royalty rate, balance of advance deposit (applicable to leased beds), royalty rate amount due and dates of landing.
5. Administrative kelp bed number and, if applicable, marine protected area where plants were harvested.

(B) A duplicate copy of the landing record shall be retained by a kelp harvester for a period of one year and shall be available for inspection at any time within that period by the department. A kelp harvester who harvests kelp from a marine protected area established under subsection 632(b) shall maintain a copy of the landing record on board the harvest vessel for all harvesting conducted during that harvest control period. The original and one copy of the landing record shall be submitted to the department's Accounting Services Branch at 1416 Ninth Street, Room 1215, Sacramento, CA 95814 (or by postal delivery to P.O. Box 944209, Sacramento, CA 94244-2090) on or before the 10th day of each month following the month to which the landing records pertain with the specified royalty required for all kelp and other aquatic plants harvested. Landing records that are mailed shall be postmarked on or before the 10th day of each month following the month to which the landing records pertain. The landing record shall be submitted whether or not harvest occurred.

(C) Failure to submit the required landing records of harvest activity and royalty fees within the prescribed time limit and/or failure to retain the required landing records for the prescribed time period(s) may result in revocation or suspension (including non-renewal) of the harvester's license for a period not to exceed one year. Any revocation, suspension, or nonrenewal may be appealed to the commission.

(4) No eel grass (*Zostera*) or surf grass (*Phyllospadix*) may be cut or disturbed.

(5) No kelp or other aquatic plant may be harvested in a state marine reserve or state marine park as per subsection 632(a).

Commercial harvest of kelp or other aquatic plants may be limited in state marine conservation areas as per subsection 632(b).

(6) It is unlawful to cause or permit waste of any kelp or other aquatic plants taken in the waters of this state or to take, receive or agree to receive more kelp or other aquatic plants than can be used without waste or spoilage.

(c) Harvesting of *Macrocystis* and *Nereocystis* (giant and bull kelp). In this subsection, kelp means both giant and bull kelp.

(1) A kelp harvester may harvest kelp by cutting and removing portions of attached kelp or by collecting unattached kelp.

(2) A kelp harvester may not cut attached kelp at a depth greater than four feet below the surface of the water at the time of cutting.

(3) No kelp received aboard a harvesting vessel shall be allowed to escape from the vessel or be deposited into the waters of this state.

(4) In beds north of Point Montara, *Nereocystis* (bull kelp) may only be taken by hand harvesting. No mechanical harvesters of any kind shall be allowed.

(5) Between April 1 and July 31, a kelp harvester may not harvest bull kelp from a nonleased kelp bed that lies partially or totally within the boundary of the Monterey Bay National Marine Sanctuary extending from Santa Rosa Creek, San Luis Obispo County, northward to Rocky Point, Marin County. This subsection does not preclude the removal of bull kelp from beaches within the Monterey Bay National Marine Sanctuary during the seasonal closure.

(6) Prior commission approval of a kelp harvest plan is necessary before a kelp harvester may use a mechanical harvester to harvest giant kelp.

(A) A kelp harvest plan must identify how a mechanical harvester will be used while avoiding:

1. repetitive harvest from individual giant kelp plants;
2. harvest of bull kelp from those portions of kelp beds that contain both giant kelp and bull kelp; and
3. harvest of giant kelp near sea otter rafting sites used by female sea otters with dependent pups.

(B) All kelp harvest plans shall also include the following:

1. the number of the designated bed or beds as shown in subsection 165.5(j), a description of the kelp bed or portion of the kelp bed requested and the designated number of square miles in each bed or portion thereof;
2. intended use of kelp;
3. amount of kelp proposed to harvest on a monthly and annual basis during the next five years;
4. estimated frequency of harvesting activities for each kelp bed;
5. number of harvest boats, maximum kelp holding capacity in wet tons for each boat, including the operating vessel gross tonnage and fuel tank capacity;
6. harvesting methodology (harvest operation description);
7. all locations (addresses) where kelp landing and weighing will take place;
8. specific details of wet kelp weighing equipment and methods to be used at the landing sites for accurate reporting; and
9. name, address, phone number, and license number of kelp processor and method of transporting the kelp to the processing location.

(C) Kelp harvest plans must be updated and submitted to the commission for approval every five years.

(7) In addition to the license fee, a kelp harvester shall pay a royalty of \$1.71 for each ton (2,000 lb) of wet kelp harvested from a non-leased bed.

(d) Harvesting of marine plants of the genera Gelidium, Pterocladia, Gracilaria, Iridaea, Gloiopeltis or Gigartina which are classified as agar-bearing plants.

(1) General Provisions.

(A) All agar-bearing plants must be harvested by cutting, except that drift or loose plants may be picked up by the harvester. Agar-bearing plants may be cut no closer than two inches to the holdfast and no holdfast may be removed or disturbed. All agar-bearing plants which are removed from a bed must be taken from the water for weighing and processing.

(B) While harvesting agar-bearing plants, it is unlawful to harvest abalone or to have abalone harvesting equipment in possession.

(C) License numbers of the harvesters will be displayed on both sides of the boat from which they are operating in 10-inch black numbers on a white background.

(D) A harvester may use conventional underwater diving gear or SCUBA when harvesting agar-bearing plants.

(2) Kelp Drying Permits. Pursuant to section 6653.5 of the Fish and Game Code, no company or individuals shall reduce the moisture content or otherwise dry agar-bearing plants harvested from waters of the state except under the authority of a kelp drying permit issued by the department. Drying permits shall be issued under the following conditions:

(A) Where Issued. Requests for kelp drying permits shall be submitted to the Department of Fish and Game at the address listed in section 165(a)(3).

(B) Cost of Permit. See subsection 699(b) of these regulations for the fee for this permit.

(C) Permit Review. The department shall return permit application forms to the applicant within three working days of receipt.

(D) Duration of Permits. Except as otherwise provided, kelp drying permits shall be valid for a term of one year from date of issue.

(E) Weighing of Kelp. All agar-bearing marine plants shall be weighed upon landing pursuant to the provisions of subsection (b)(1) of these regulations.

(F) Plant Delivery. Every person taking delivery of agar-bearing marine plants for drying purposes from persons licensed pursuant to section 6650 of the Fish and Game Code or harvesters drying their own plants shall keep a book or books recording the following:

1. A full and correct record of all agar-bearing plants received from other licensed agar harvesters or taken by permittee.
2. Names of the different species.
3. The number of pounds received.
4. Name, address and kelp harvester number of the person from whom the agar-bearing plants were received. The book(s) shall be open at all times for inspection by the department.

(G) Landing Receipts. Receipts shall be issued by all kelp drying permittees to harvesters licensed pursuant to subsection (b)(3) of these regulations and shall show:

1. Price paid.
2. Department origin block number where the agar-bearing plants were harvested.
3. Such other statistical information the department may require.

(H) The original signed copy of receipt shall be delivered to the agar harvester at the time of purchase or receipt of the agar-bearing plants. The duplicate copy shall be kept by the kelp drying permittee for a period of one year and shall be available for inspection at any time within that period by the department, and the triplicate shall be delivered to the department at the address indicated within 10 days after the close of each month, with a royalty of \$17.00 per wet ton (2,000 lbs.) for all agar-bearing seaweed received. Failure to submit the required landing receipts and royalty fees within the prescribed time limit is grounds for revocation of the permittee's drying permit.

(e) Harvesting of marine plants, including the genera Porphyra, Laminaria, Monostrema, and other aquatic plants utilized fresh or preserved as human food and classified as edible seaweed.

(1) General Provisions.

(A) Edible varieties of marine plants must be harvested by cutting or picking, except that drift or loose plants may be picked up by the harvester. All harvested plants must be processed.

(B) Edible seaweed may be harvested from state waters throughout the year, except as provided under section 164.

(C) While harvesting edible seaweed, it is unlawful to harvest abalone or to have abalone harvesting equipment in possession.

(D) A harvester may use conventional underwater diving gear or SCUBA while harvesting edible seaweed.

(2) Harvest of Bull Kelp for Human Consumption. Notwithstanding subsection 165(c) (5)(A), persons operating under the authority of an edible seaweed harvesters license may take, not to exceed, 2 tons (4,000 lbs) of bull kelp per year. The entire plant may be harvested.

(3) Weighing of Edible Marine Plants. All edible marine plants shall be weighed pursuant to the provisions of subsection (b)(1) of these regulations and landing receipts in duplicate issued as per subsection (b)(3).

(4) The original copy of the receipt shall be delivered to the department at the address indicated within 10 days after the close of each month with a royalty of \$24 per wet ton (2,000 lbs.) of edible marine plants harvested from state waters other than San Francisco Bay and Tomales Bay.

F) All Other Species of Kelp.

(1) Applicant shall apply to the commission, outlining the species to be harvested, amount and location. The commission may set conditions and amount of royalty after review of the application.

Note: Authority cited: Sections 6653 and 6653.5, Fish and Game Code. Reference: Sections 6650, 6651, 6652, 6653, 6653.5, 6654, 6656 and 6680, Fish and Game Code.

1. Amendment of subsection (a)(3) filed 10-8-69 as an emergency; designated effective 11-10-69 (Register 69, No. 41). For prior history, see Register 69, No. 15.
2. Certificate of Compliance -section 11422.1, Gov. Code, filed 12-17-69 (Register 69, No. 51).
3. Amendment of subsection (a)(1)(E) filed 6-30-77 as an emergency; effective upon filing (Register 77, No. 27).
4. Certificate of Compliance filed 8-24-77 (Register 77, No. 35).
5. Amendment filed 3-9-81; effective thirtieth day thereafter (Register 81, No. 11).
6. Amendment filed 9-6-85; effective thirtieth day thereafter (Register 85, No. 36).
7. Change without regulatory effect of subsection (e)(3) filed 5-5-86; effective thirtieth day thereafter (Register 86, No. 19).
8. Amendment of subsections (a)(2), (a)(3) and (c)(5) filed 1-27-87; effective thirtieth day thereafter (Register 87, No. 5).
9. Amendment of subsection (c) filed 12-3-90; operative 1-2-91 (Register 91, No. 4).
10. Amendment of subsections (a) and (d) filed 4-18-91; operative 5-18-91 (Register 91, No. 21).
11. Editorial correction of printing error in subsection (c)(3) (Register 91, No. 31).
12. Amendment of subsections (a)(3) and (c)(2), new subsections (c)(5)-(c)(5)(B), subsection renumbering and amendment of newly designated subsection (c)(6), and new subsection (e)(3) and subsection renumbering filed 3-26-96; operative 3-26-96 pursuant to Government Code section 11343.4(d) (Register 96, No. 13).
13. Amendment filed 10-25-2001; operative 11-24-2001 (Register 2001, No. 43).
14. Amendment of subsection (b)(5) filed 3-8-2005; operative 4-7-2005 (Register 2005, No. 10).
15. Change without regulatory effect amending subsection (a)(3) filed 5-5-2005 pursuant to section 100, title 1, California Code of Regulations (Register 2005, No. 18).
16. Amendment of subsections (b)(1), (b)(3), (b)(3)(D)-(F) and (c)(4)(D) filed 8-22-2007; operative 9-21-2007 (Register 2007, No. 34).
17. Editorial correction restoring inadvertently omitted subsection (c)(5) (Register 2011, No. 5).
18. Amendment of section and Note filed 1-14-2014; operative 4-1-2014 (Register 2014, No. 3).
19. Change without regulatory effect amending subsection (a)(1) filed 5-22-2014 pursuant to section 100, title 1, California Code of Regulations (Register 2014, No. 21).
20. Change without regulatory effect amending subsection (a)(1) filed 6-24-2015 pursuant to section 100, title 1, California Code of Regulations (Register 2015, No. 26).
21. Change without regulatory effect amending subsection (b)(3)(B) filed 8-8-2019 pursuant to section 100, title 1, California Code of Regulations (Register 2019, No. 32).

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Research Article

Assessment of control methods for the invasive seaweed *Sargassum horneri* in California, USA

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Editor's note:

This study was first presented at the 9th International Conference on Marine Bioinvasions held in Sydney, Australia, January 19–21, 2016 (<http://www.marinebioinvasions.info/previous-conferences>). Since their inception in 1999, ICMB series have provided a venue for the exchange of information on various aspects of biological invasions in marine ecosystems, including ecological research, education, management and policies tackling marine bioinvasions.

Abstract

Determining the feasibility of controlling marine invasive algae through removal is critical to developing a strategy to manage their spread and impact. To inform control strategies, we investigated the efficacy and efficiency of removing an invasive seaweed, *Sargassum horneri*, from rocky reefs in southern California, USA. We tested the efficacy of removal as a means of reducing colonization and survivorship by clearing *S. horneri* from 60 m² circular plots. We also examined whether *S. horneri* is able to regenerate from remnant holdfasts with severed stipes to determine whether efforts to control *S. horneri* require the complete removal of entire individuals. The experimental removal of *S. horneri* in early winter, just prior to the onset of reproduction, reduced recruitment in the next generation by an average of 54% and reduced survivorship to adulthood by an average of 25%. However, adult densities one year after clearing averaged 83% higher in removal plots and 115% higher in control plots. We attribute these higher densities to anomalously warm water associated with the 2015–16 El Niño that reduced native canopy-forming algae and enhanced the recruitment and survival of *S. horneri*. We did not find any evidence to suggest that *S. horneri* has the capacity to regenerate, indicating that its control via removal does not require the tedious task of ensuring the removal of all living tissue. We developed efficiency metrics for manual removal with and without the aid of an underwater suction device and found the method with maximum efficiency (biomass removed worker⁻¹ hr⁻¹) varied based on the number of divers and surface support workers. Our findings suggest that controlling *S. horneri* via removal will be most effective if done over areas much larger than 60 m² and during cool-water years that favor native algae. Such efforts should be targeted in places such as novel introduction sites or recently invaded areas of special biological or cultural significance where *S. horneri* has not yet become widely established.

Key words: introduced species, management, marine, macroalgae, rocky reef, *Sargassum filicinum***Introduction**

Invasive species are one of the greatest agents of human-induced change to ecosystems worldwide (Pejchar and Mooney 2009). Coastal marine systems are especially vulnerable to introductions of nonindigenous species via trans-oceanic shipping, aquaculture

and the aquarium trade, which have greatly extended the distribution of many marine species outside of their native ranges (Carlton 1989). Marine invasions have steadily increased over the past two centuries (Ruiz et al. 2000) and are expected to continue to rise as global trade expands. Costs associated with the impact and management of invasive species are high, totalling over \$1 billion annually in the USA

(Pimentel et al. 2000), while resources available for management are limited. Therefore, agencies tasked with controlling invasions must be efficient in their management strategies. Exploration of techniques aimed at controlling the spread and impact of marine invasive species and identification of species-specific traits that increase the efficacy of control are urgently needed.

A seaweed recently introduced to southern California, USA, presented an opportunity to test the efficacy of removal in controlling invasive algae on rocky reefs. *Sargassum horneri* (Turner) C. Agardh, 1820 (Fucales) is a large, annual brown alga native to shallow reefs of eastern Asia. It was first discovered in the eastern Pacific in Long Beach Harbor in 2003 and identified as *S. filicinum* Harvey, 1860 (Miller et al. 2007), now considered a synonym of *S. horneri* (Uwai et al. 2009). The species has spread aggressively across 700 km from Santa Barbara in southern California to Isla Natividad in Baja California, Mexico (Marks et al. 2015). It occurs primarily at offshore islands though it has also been found along the mainland and in coastal embayments. In southern California we have observed *S. horneri* growing in the intertidal down to 33 m depth, with its highest densities occurring between 5–15 m. In places where *S. horneri* is established, juveniles can attain high cover with upwards of 1,000 individuals m⁻² during the summer and fall, and these grow to form thick canopies in the winter with dense stands of over 100 adults m⁻² (author's unpublished data). While definitive evidence of ecological impacts on rocky reef systems from *S. horneri* invasion is not yet available (but see Cruz-Trejo et al. 2015), the detrimental effects on native assemblages caused by other invasive seaweeds (e.g., de Villèle and Verlaque 1995; Levin et al. 2002; Casas et al. 2004; Britton-Simmons 2004) suggest management of *S. horneri* is worth exploring (Anderson 2007; Schaffelke and Hewitt 2007; Forrest and Hopkins 2013).

Several life history characteristics of *S. horneri* make it potentially suitable for control by removal. First, it is a large and conspicuous alga consisting of a single main axis with multiple lateral branches that reaches up to several meters high (Yoshida 1983). The annual thallus is anchored by a small holdfast that gives rise to a stipe buoyed by many small gas bladders (Marks et al. 2015). The conspicuous adult thalli allow for efficient identification and removal by divers using SCUBA. Second, *S. horneri* propagates via sexual reproduction. Fertilization occurs in winter on the surface of reproductive structures born on the lateral branches of a mature thallus where embryos are developed and shed (author's unpubli-

shed data). Senescence of the thallus ensues after embryos are shed, completing the annual life cycle. *Sargassum* embryos tend to sink quickly (Gaylord et al. 2002) and the vast majority likely settle within a few meters of the parent thallus (Deyscher and Norton 1982; Stiger and Payri 1999; Kendrick and Walker 1995). Clearing thalli in relatively small areas on the order of tens of square meters may therefore reduce colonization resulting from local dispersal. However, because colonization over longer distances is thought to occur via reproductively mature thalli that are dislodged and set adrift (Yatsuya 2008), any thalli removed must not be released. Asexual reproduction in *S. horneri* via fragmentation or regeneration from remnant tissue has not been studied, although it is known to occur in other fucoid species (McCook and Chapman 1992; Fletcher and Fletcher 1975). Information on the capacity of *S. horneri* to propagate asexually is needed to develop an effective management strategy for controlling its spread.

A new tool that has been developed to help control algal invasions is an underwater suction device. This type of device has been used on coral reefs in Oahu, Hawaii, to reduce densities of invasive algae (Conklin and Smith 2005), and a similar device was recently developed to aid in controlling seaweed invasions on rocky reefs in California. The device has been used to transport *S. horneri* removed from the ocean floor by divers to a platform at the sea surface, where the material can be collected for disposal on land (Meux 2013). However, the effectiveness of this approach in controlling *S. horneri* on temperate rocky reefs and how the efficiency of this method compares to non-mechanical techniques require further investigation.

To inform efforts to manage the spread and impact of *S. horneri*, we removed it from experimental areas and followed colonization and survivorship for one year to address three questions. First, how effective is local removal in controlling populations of *S. horneri*? Second, what is the capacity of the species to regenerate from remnant holdfasts? Third, how much effort is required to remove established populations with and without the aid of an underwater suction device?

Methods

This study was performed on the leeward side of Santa Catalina Island, California, USA on two nearby reefs (Howland Landing: 33.465°N; 118.522°W and Lion Head: 33.453°N; 118.502°W) at 6–8 m depth (Figure 1). We chose these locations because they are representative of the topography of reefs in the area, and have dense populations of *S. horneri*.

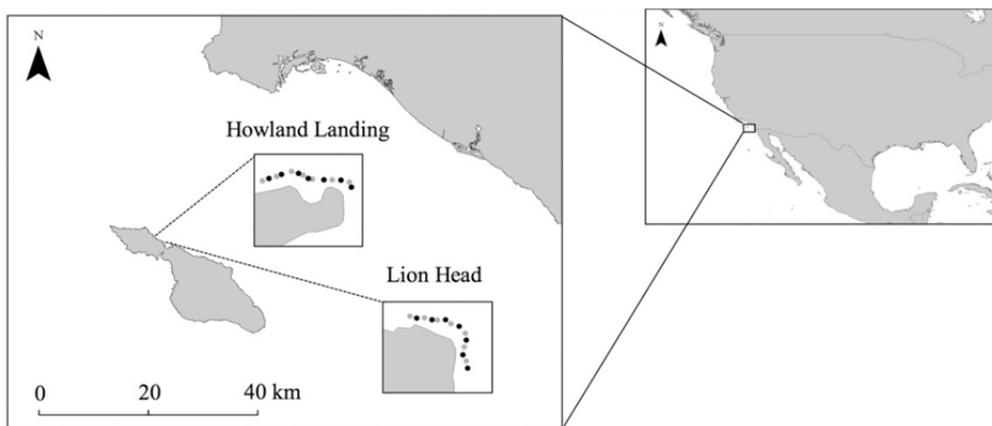


Figure 1. Map of Santa Catalina Island, located 27 km off the coast of southern California, USA. The insets show the distribution of 28 experimental plots spread across two sites: Howland Landing and Lion Head. Dark circles represent removal plots, and grey circles represent control plots.

Removal experiment

To evaluate the effectiveness of *S. horneri* extraction in reducing local populations, we performed a removal experiment and monitored colonization and survivorship of the next generation. We established twenty-eight 60 m² circular plots in areas where *S. horneri* was abundant and assigned plots alternately to either a removal or non-removal (i.e., control) treatment (Figure 1). Fourteen plots were located 15–20 m apart at each of the two study sites.

We extracted *S. horneri* from removal plots in the winter (February 2015) when individuals were at their largest size and lowest densities, but before the vast majority (i.e., 99%) of them became fertile so as to minimize the source of *S. horneri* propagules within the removal plots. Immediately prior to removal we counted the number of *S. horneri* adults (defined as > 5 cm tall) in sixteen 0.25 m² quadrats plot⁻¹ that were placed within each plot at 0, 1, 2 and 3 m from the edge along two perpendicular diameters. To prevent mature thalli from drifting away and starting distant populations, we captured all material removed and transported it to boats anchored at the surface. On deck, workers immediately transferred material into heavy-duty trash bags. We later emptied these bags at an upland location where we left the algae to decompose.

We removed all *S. horneri* from the substrate manually and employed one of two methods to transport it to the surface: mesh bags and lines, or an underwater suction device. The bag and line method involved divers placing *S. horneri* into weighed bags (Figure 2A). Once filled, buoyant bags were released from their weights and attached to lines hanging off

the side of the boat (Figure 2B) and a worker at the surface hauled them onboard. The suction device consisted of a mechanical water pump (Subaru PTX201D Robin Pump) with 7.6 cm-diameter input and output hoses that is operated on the deck of the boat (Figure 2C). Divers fed material into the hose at depth and it was transported to the surface by the movement of a diaphragm (Figure 2D). Regardless of the method used, most individuals were completely removed from the substrate (Figure 2E). However, the holdfasts of some individuals remained after their stipes were severed.

Removal plots were resampled immediately after clearing to confirm all thalli had been removed and to quantify the density of remnant holdfasts. In September 2015, we measured colonization by counting the number of juveniles (defined as < 5 cm tall) in all plots. In February 2016, one year after experimental removals, we counted the number of adults in each plot to assess the effects of removal on population density. Because *S. horneri* grows on rock and the percent cover of rock was consistently high but slightly variable (mean ± SE = 97.9 ± 0.19%) we adjusted estimates of density within each quadrat by the percent cover of rock in that quadrat. Hence *S. horneri* is reported as number m⁻² of rock rather than number m⁻² of sea floor.

We tested the effects of removal on colonization (i.e., juvenile density in September 2015) and population density (i.e., adult density in February 2016) in separate two-way hierarchical ANOVAs with treatment (removal versus control) as a fixed factor and site (Howland Landing versus Lion Head) as a random factor and plots nested within sites. We considered plots independent replicates of treatment effects in cases when the random effect of site was not significant.

Fate of individuals with severed stipes

To determine whether severing a *S. horneri* stipe near its base while leaving the holdfast intact is sufficient to prevent it from regenerating, we followed the fate of individuals after cutting their stipes in March 2015. We attached identifying markers to the reef adjacent to 80 holdfasts and revisited the marked individuals monthly for four months to record whether they remained attached to the substrate and, if so, whether they regenerated new tissue. We also collected observations of the remnant holdfasts in the plots we cleared. Although we were not able to follow these holdfasts individually, we looked for perennating *S. horneri* holdfasts when resampling the plots.

Efficiency of removal

We evaluated the efficiency of removal with and without the aid of the suction device (Figure 2) by quantifying the effort required for each method for a given quantity of *S. horneri* biomass. We did this by recording the removal method being used (i.e., suction device or bags and lines), time spent collecting, number of workers (i.e., scuba divers and surface support person) and amount of biomass removed for each dive. To estimate the biomass removed, we collected the algae into bags as soon as it was brought to the surface and weighed it to the nearest 0.5 kg using a hanging scale. In addition, we measured the rate of transport to the surface using the suction device across a range of stipe lengths to determine if size affected performance. We fed 30 pieces of several stipe lengths that are often naturally observed (30 cm, 60 cm, 100 cm and 150 cm) into the hose and recorded the time it took to bring them up to the surface.

Results

Removal experiment

The density of adult *S. horneri* prior to experimental removal in February 2015 was similar in removal and control plots ($F_{1,1} = 0.98$, $p = 0.504$) averaging 46.4 and 50.4 individuals m^{-2} , respectively (Figure 3A). Adult density differed significantly between the two sites ($F_{1,420} = 26.95$, $p < 0.001$) with density ~55% higher at Howland Landing. Quadrat sampling and visual surveys of entire plots verified that experimental clearing resulted in the removal of virtually all visible thalli in removal plots, but some holdfasts with severed stipes remained. The density of remnant holdfasts immediately after clearing was 46.1% of the initial adult population (mean \pm SE = $22.3 \pm 2.9 m^{-2}$).

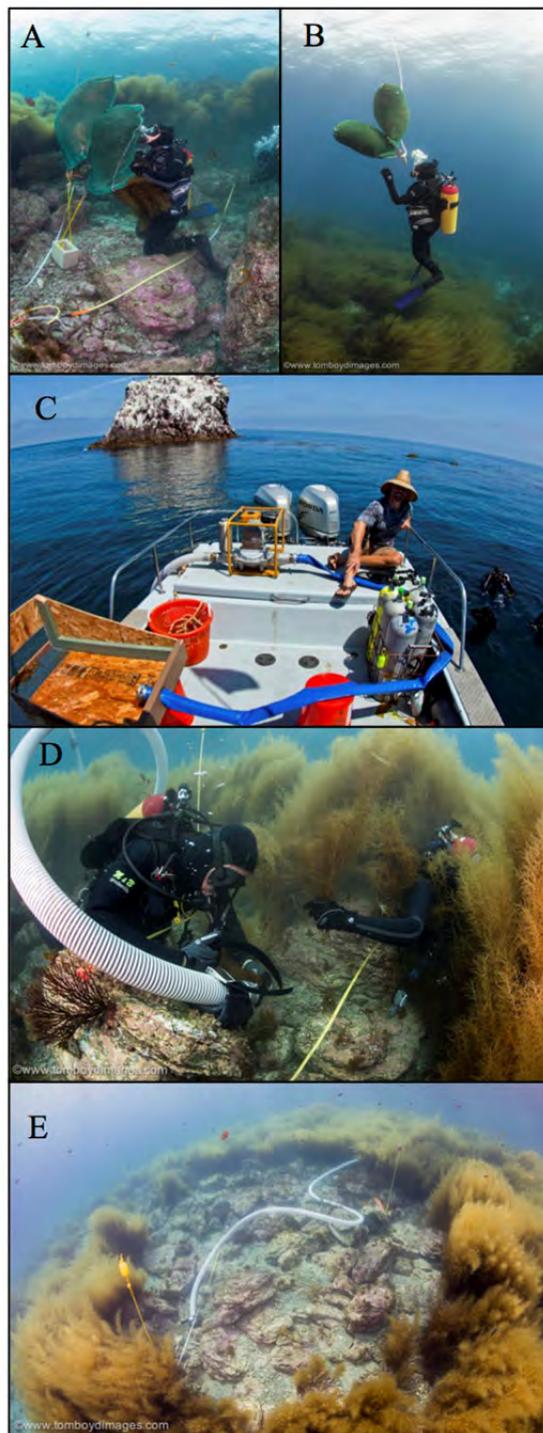


Figure 2. Two methods used to transport *Sargassum horneri* to the surface. Using the bag and line method, a diver fills bags anchored by a cinderblock (A), then clips bags to a line hanging from a boat anchored overhead (B). Using the suction device method, two divers work together to feed *S. horneri* into the hose (C), and a person at the surface collects the material from a sorting table after inspecting it for bycatch (D). After clearing using both methods, plots were left barren of *S. horneri* (E). Photo credits: Tom Boyd (A-B, D-E), Adam Obaza (C).

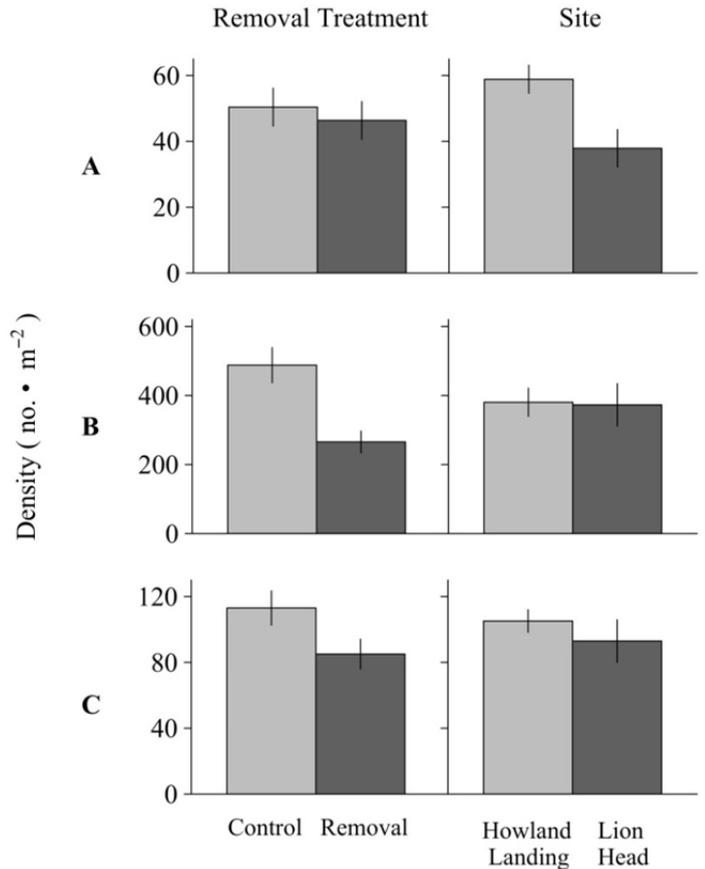


Figure 3. Results of removal experiment showing the average density \pm SE of *Sargassum horneri* (A) adults prior to their removal, (B) juveniles ~220 days after removal, and (C) adults ~366 days after removal (N = 14 plots).

Similarly high densities of recently colonized juveniles were observed in all plots in September 2015, ~7 months after clearing (Figure 3B; $F_{1,420} = 0.08$, $p = 0.775$). Removal had a significant effect on subsequent colonization ($F_{1,26} = 12.95$, $p = 0.001$) as juvenile density was 54% lower in removal plots compared to control plots. The effect of removing *S. horneri* on colonization by juveniles was similar at both sites (treatment \times site: $F_{1,1} = 0.236$, $p = 0.125$).

The reduced densities in removal versus control plots persisted but became less pronounced over time as juveniles grew into adults (Figure 3C). By February 2016, one year after clearing, adult densities averaged 25% lower in removal plots compared to control plots. However, overall adult densities were 83% higher in removal plots and 115% higher in control plots compared to February 2015 prior to removal (Figure 3A versus 3C).

Fate of individuals with severed stipes

Significant tag loss resulted in reduced and unequal sample sizes for estimating survivorship on the different sampling dates, which compromised our ability

to quantitatively evaluate the regenerative capacity of individuals with severed stipes. Nonetheless, the data that we collected indicate that *S. horneri* has little or no capacity for regenerating from remnant holdfasts as none of the individuals with severed stipes that remained tagged generated new tissue. Fifty-six of the 80 tags remained after 31 days and remnants of holdfasts were found for only 20 of these 56 individuals. Remnants of 10 of 14, 4 of 9 and 0 of 8 holdfasts remained after 54, 85 and 113 days, respectively (Figure 4). Furthermore, when we sampled the removal experiment in September 2015, we found no remnant holdfasts, which suggests they had all senesced and disappeared within seven months.

Efficiency of removal

The efficiency of removing *S. horneri* varied by the method used to transport it to the surface and the number of workers. Three workers using the bag and line transport method yielded the slowest average removal rate of 29 kg worker⁻¹ hr⁻¹, while the suction device method with three workers (two divers and one surface support person) yielded an average of 38 kg

worker⁻¹ hr⁻¹ (Figure 5). Limits on the amount of material that can be fed into the hose at any given time resulted in two divers being the optimum number to maximize the transport of algae to the surface. By contrast, the manual transportation method using bags and lines allowed for more divers to work efficiently in the same area. While the overall rate of removal using bags and lines increased with the number of workers, the maximum per capita efficiency was about 45 kg worker⁻¹ hr⁻¹ (Figure 5). The rate of transport using the suction device was highest at intermediate stipe lengths (~60 cm; Figure 6).

Discussion

Our results show that the experimental removal of *S. horneri* reduced the local population in the next generation by ~25% relative to control plots. However, despite this reduction, removing *S. horneri* did not lead to a decline in population density relative to the previous year as adult densities in both the removal and control plots were substantially greater in 2016 than in 2015 prior to removal. These results highlight some of the challenges associated with efforts to reduce established populations of *S. horneri* via removal. Moreover, they suggest that measurable success using removal techniques as a means of controlling *S. horneri* will likely require that removals be done over much larger areas to ensure an adequate reduction in propagule supply, which will be costly. The effect of removing *S. horneri* on its abundance in subsequent generations (as measured by the difference in *S. horneri* density between control and removal plots in the year following removal) was most apparent during the fall when the majority of individuals were juveniles, and became less pronounced in the winter when most were adults. The order of magnitude higher densities that we observed for juveniles compared to adults is consistent with self-thinning induced by intra-specific competition, which is common in large brown algae (Schiel and Choat 1980; Schiel 1985; Dean et al. 1989; Reed 1990). The dampened effect of removal between the juvenile and adult phases suggests removal accelerated the self-thinning process.

The increased density of *S. horneri* that we observed in our removal and control plots may have been due to the unusually warm water resulting from the 2015–16 El Niño. The native canopy-forming kelps *Macrocystis pyrifera* and *Eisenia arboria* commonly found on shallow reefs of Santa Catalina Island thrive in cool, nutrient-rich water. These species largely disappeared from the leeward side of the island during our study while *S. horneri* flourished, as did other species with warm water affinities (e.g.,

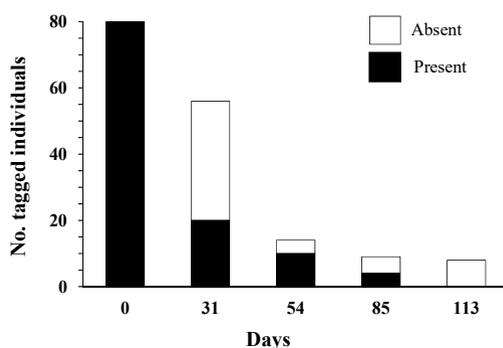


Figure 4. Survivorship of *Sargassum horneri* with severed stipes. Solid bars represent the number of thalli with remnant tissue remaining. Open bars represent the number of tags relocated where holdfasts had senesced. Combined, the bars represent the total number of tags found, and the number of individuals upon which survivorship was based for each sampling period.

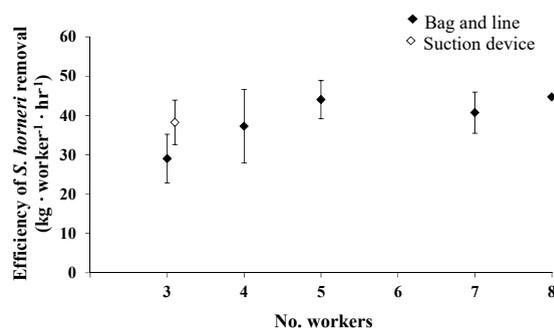


Figure 5. *Sargassum horneri* average removal rate (kg wet biomass worker⁻¹ hr⁻¹) ± SE reported for each removal method. Replication varies by the number of dives with each given number of workers using each method. N = 15 dives with 3 workers using the suction device, and N = 6, 4, 6, 6 and 1 dives with 3, 4, 5, 6, and 8 workers using the bag and line removal method, respectively.

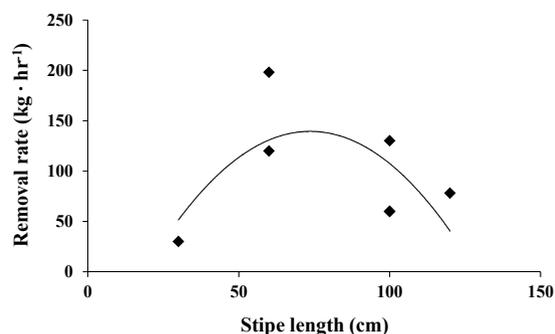


Figure 6. The rate (kg wet biomass hr⁻¹) at which stipes of *Sargassum horneri* were transported by workers using the suction device as a function of stipe length.

Zonaria farlowii, *Dictyota* spp. and *Dictyopteris undulata*). Evidence from the nonindigenous congener *Sargassum muticum*, which became abundant at Santa Catalina Island for several years following the El Niño of 1976 (Coyer 1979), suggests that *Sargassum* spp. with warm-water affinities decline once cooler waters return and large, perennial native kelps become re-established (Ambrose and Nelson 1982) Whether *S. horneri* declines over time remains to be seen, but if the warming observed in 2015–16 is a preview of future conditions, then tropicalization of an algal assemblage that favors *S. horneri* may be the norm.

The efficacy of removing invasive algae could be strengthened by selecting conditions under which native species can exert biotic control on the remaining population, or even by enhancing these controls. Researchers in Hawaii attributed their success in controlling invasive *Eucheuma* spp. and *Kappaphycus* spp. on patch reefs to introducing urchins after performing removals (Conklin and Smith 2005). Once divers reduced the algae below a critical threshold, the herbivores were able to prevent it from growing back. While this is an effective strategy on coral reefs where indiscriminant grazing is acceptable, introducing generalist herbivores is not a viable strategy to control invasive algae on temperate rocky reefs, which are often dominated by a diversity of macroalgae.

An alternative strategy to enhance biological resistance to the regrowth of invasive algae on rocky reefs is to perform removals under conditions favoring the colonization of native species of macroalgae and sessile invertebrates that compete for space and/or light. Resource competition is recognized as an important mechanism structuring communities (MacArthur 1970; Levine and D'Antonio 1999; Tilman 2004), and competition for space and light plays a key role in organizing the benthic community on rocky reefs (Miller and Etter 2008; Arkema et al. 2009). The invasion of a community is thought to be inversely related to species richness due to the enhanced ability of resident species to preempt resources (Elton 1958), and manipulative field experiments have shown that decreasing native diversity increases limited resources and the abundance and survivorship of non-native species in subtidal benthic communities. For example, Stachowicz et al. (2002) found that experimentally increasing sessile invertebrate species richness decreased both the availability of space, the limiting resource in this system, and the abundance of non-indigenous ascidians by buffering against temporal fluctuations in the cover of individual native species. Furthermore, multiple resources might be limiting the success of a non-native species throughout its life

cycle, and higher functional diversity may allow a community to preempt multiple resources more effectively. A native algal community with crustose and turfing algae preempting space and understory and canopy-forming algae preempting light sequentially suppressed the recruitment and survivorship of the nonindigenous seaweed *Sargassum muticum* (Britton-Simmons 2006). The preemption of limited resources by native species of algae and invertebrates in areas where *S. horneri* has been removed could likewise limit *S. horneri*'s ability to re-establish.

Another important factor to consider when controlling invasive algae through removal is the mechanisms by which it recolonizes cleared areas. Many species of invasive algae have the ability to regenerate from miniscule amounts of tissue (e.g., Fletcher and Fletcher 1975; McCook and Chapman 1992) and this characteristic presents a challenge when considering control via removal (Smith 2015). We found no evidence that *S. horneri* has the capacity to regenerate from remnant holdfasts. This suggests that severing stipes, which is far less time consuming than carefully scraping all tissue from the reef, would be an effective and efficient means of reducing *S. horneri* abundance.

Whether an underwater suction device, such as the one tested in this study, would be the preferred method for invasive algae control depends on staff and budget limitations. The bag and lines method is optimal when many workers (i.e., > two divers and one surface support worker) are available. It also requires minimal training and material costs, and so may be preferred with constrictive budgets. A suction device minimizes surface support effort, particularly associated with lifting heavy bags, and offers increased efficiency with a limited number of workers (< 3 divers). Drawbacks of using a suction device include increased start-up costs, logistical challenges associated with equipment transportation and maintenance, and limitations on working depths. In addition, significant time can be spent troubleshooting, such as identifying appropriately sized pieces of algae to reduce the frequency of clogs. However, removal efficiency is likely to improve as operators become more familiar with the device and alter equipment to better suit the target species. Workers in Hawaii designed several models using different kinds of pumps until they identified the optimal configuration for their target species (Conklin personal communication). Therefore, long-term efficiency gains may make a suction device preferable if an extended control effort is expected.

Eradicating problematic species from their novel habitats is most likely to be successful if attempted before they become widely established (Myers et al.

2000; Bax et al. 2003; Hulme 2006). *Caulerpa taxifolia*, a green alga native to the Indo-Pacific region, was introduced in two protected embayments in southern California in 2000 and a rapid response effort successfully eradicated this species (Anderson 2005). The appearance of *S. horneri* off the open coast of North America is the first record of this species outside of its native range in Asia (Marks et al. 2015). While the aggressive spread of *S. horneri* throughout southern California and Baja California, Mexico makes total eradication in this region highly unlikely, *S. horneri* has the potential to spread to other temperate reefs around the globe. Knowledge about the life history and effective methods for controlling *S. horneri* abundance will prepare resource managers in other regions to eradicate new populations before they become widely established. Our study is one of the first on targeted control of an invasive species on the open coast of California. Development of a removal protocol along with awareness generated by this work will better prepare resource managers and the general public for future invasions of *S. horneri* in other regions.

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Supplemental material

The following supplementary material is available for this article:

Table S1. Recruitment and survivorship of *Sargassum horneri* following removal

Table S2. Survivorship of *Sargassum horneri* with severed stipes

Table S3. Per capita removal rate of *Sargassum horneri*

Table S4. Transport rate of *Sargassum horneri* using suction device

This material is available online for download from the Long Term Ecological Research Network Data Portal, <http://dx.doi.org/10.6073/pasta/a812d149f4d6e9cd5662d4c44eaedd2>

Impacts of the non-native alga *Sargassum horneri* on benthic community production in a California kelp forest

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ABSTRACT: The arrival of *Sargassum horneri* throughout the Southern California Bight and the Baja Peninsula has raised concern regarding kelp forest resilience and ecosystem function following the invasion of this non-native species. To understand how *S. horneri* impacts native algal abundance and community production, we removed *S. horneri* from experimental plots over a period of 11 mo. We measured impacts on native algal communities and community productivity using SCUBA surveys and benthic chambers equipped with oxygen, temperature, and light sensors. We observed a nearly 4-fold increase in recruitment of *Macrocystis pyrifera* and a 9-fold increase in adult *M. pyrifera* stipe density in *S. horneri* removal plots, but no discernable changes in net community production among treatments. We found ephemeral increases in gross community production and community respiration in the non-removal plots that coincided with periods of peak *S. horneri* biomass. To understand the temporal dynamics of community production, we deployed benthic chambers across a rocky reef dominated by *S. horneri*. Here, temporal variation in community production was most strongly related to corresponding variation in water temperature and changes in *S. horneri* biomass related to its annual lifecycle. Overall, our study indicates that *S. horneri* presence contributed to ephemeral increases in gross community production and community respiration, but it did not affect net community production. Moreover, *S. horneri* removal can lead to increases in native algal abundances given favorable abiotic conditions. We suggest that *S. horneri* thrives in a disturbed ecosystem rather than being a driver of ecosystem change.

KEY WORDS: Kelp forest · Non-native algae · Community production · Community respiration · *Sargassum horneri*

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1. INTRODUCTION

Kelp forests support diverse and productive communities in temperate marine ecosystems, making them ecosystems of paramount importance (Schiel & Foster 2015). Primary production in these forests ranges between 400 and 1900 g C m⁻² yr⁻¹, making them comparable to terrestrial rainforests (Mann 1973, Melillo et al. 1996, Reed & Brzezinski 2009). Temporal variation in kelp forest primary production is driven by corresponding changes in environmental conditions and primary producer biomass, which

may be strongly dependent on the life-history characteristics of the dominant algae (Miller et al. 2011, Rodgers & Shears 2016, South et al. 2016). Measurements of community and primary production can be a meaningful metric of ecosystem function and instrumental in helping us understand ecosystem dynamics, especially in cases where production patterns are altered by the invasion of non-native autotrophs.

Disturbances that decrease abundance or alter the distribution of native species can aid in the establishment of non-native (i.e. invasive) species (Valentine & Johnson 2003, MacDougall & Turkington 2005, Jauni

et al. 2015, Eviner 2016). Invasive algae can proliferate, especially after native algal assemblages are experimentally reduced (Valentine & Johnson 2003, 2004) or naturally disturbed (Scheibling & Gagnon 2006, Britton Simmons & Abbott 2008, Thomsen et al. 2019), resulting in ecosystem-wide impacts including changes to native community structure (Williams & Smith 2007) and primary production (Tait et al. 2015). Understanding these changes can help elucidate the larger impacts of species invasions and inform managers on the best practices to mitigate them.

Invasive species are often documented as having long-term negative impacts on ecosystems they invade (Simberloff & Von Holle 1999). However, invasive species can also have ephemeral, and sometimes positive, impacts on an ecosystem (South et al. 2016, Flory et al. 2017, Quijón et al. 2017). *Undaria pinnatifida* has spread globally since it was first documented as invasive in Wellington Harbor, New Zealand in 1987 and has since been declared impossible to eradicate (Hay & Luckens 1987, Stuart 2002). *U. pinnatifida* is most likely to colonize in disturbed environments and have negative or neutral impacts on native algal abundance and diversity; however, these impacts are transient through time and space (Valentine & Johnson 2003, 2004, South et al. 2016). Transient impacts are likely a product of phenology. Specifically, peak *U. pinnatifida* biomass does not overlap with that of native macroalgae, decreasing the potential for competition (Hay & Villouta 2009, South et al. 2016). While community production and carbon subsidy supply may be greater during the periods when *U. pinnatifida* is abundant, total annual community production is likely lower than in areas where it has not invaded or where the native algae are able to recover (Tait et al. 2015). Similar positive ephemeral impacts on community production have been observed with other non-native annual species including *Sargassum muticum* (Pedersen et al. 2005), *Gracilaria vermiculophylla* (Thomsen & McGlathery 2007, Nejrup & Pedersen 2010), and *Codium fragile* (Thomsen & McGlathery 2007).

The fucoid alga *S. horneri* (= *S. filicinum*) (Turner) C. Agardh (Fucales, Phaeophyceae) is an invasive brown alga that originates from subtidal (2 to 20 m) rocky reefs of western Japan and South Korea (Umezaki 1984, Miller et al. 2007). *S. horneri* has high fecundity, broad dispersal capabilities, and wide-ranging temperature tolerances, which together increase its chances of establishing and proliferating in novel ecosystems (Umezaki 1984, Miller et al. 2007, Marks et al. 2018). *S. horneri* was first reported in Long Beach Harbor, California, USA, in 2003 (Miller et al. 2007) and has since spread along the California and Baja Cali-

fornia, Mexico, coasts (Riosmena-Rodríguez et al. 2012, Marks et al. 2015). *S. horneri* is now routinely observed from Isla Natividad, Baja California Sur, Mexico, to Santa Barbara, California, USA (a distance of approximately 850 km), where it has established dense populations and often overlaps in niche space with the foundational giant kelp *Macrocystis pyrifera* (Riosmena-Rodríguez et al. 2012, Marks et al. 2015, 2018). Reductions in *M. pyrifera* populations in Southern California, due to a prolonged period of warm water (e.g. Reed et al. 2016) and an El Niño-Southern Oscillation (ENSO) (e.g. Edwards 2004) that occurred during 2014 to 2016 (Edwards 2019), likely facilitated the establishment of *S. horneri* (Marks et al. 2017). In areas where *S. horneri* has replaced *M. pyrifera*, fish recruitment and distribution may be negatively impacted (Ginther & Steele 2018, Srednick & Steele 2019). However, differences in fish distributions are likely mediated by the alga's height rather than its non-native identity (Srednick & Steele 2019). Relatively little is known about the interactions between *S. horneri* and native algae or potential impacts to ecosystem function, as this is the first known spread of *S. horneri* outside of its native range (Marks et al. 2015).

This study examined how *S. horneri* impacts community production on a subtidal rocky reef at Catalina Island, California, USA. We used a combination of SCUBA diver surveys and benthic chambers to quantify temporal variation in algae diversity and community production in areas invaded by *S. horneri* and in areas where *S. horneri* was removed from experimental plots. We asked: (1) How does *S. horneri* influence native algal assemblages and patterns of benthic community production? (2) What drives temporal variation in community production in ecosystems invaded by *S. horneri*? We predicted that native algal abundances and net community production (NCP), gross community production (GCP), and community respiration (CR) will be greater in *S. horneri* removal plots compared to control plots. We predicted that temporal variation in community production over the course of the study will be positively associated with the *S. horneri* annual life cycle, specifically, periods of greater *S. horneri* biomass will have greater rates of NCP, GCP, and CR.

2. MATERIALS AND METHODS

2.1. Study site

We studied how the *Sargassum horneri* invasion affected rocky reef communities on the leeward side of

Catalina Island, just outside of Big Fisherman's Cove and adjacent to the Wrigley Marine Science Center. This 9 to 18 m deep rocky reef lies within the Blue Cavern State Marine Conservation Area (SMCA) and has been historically dominated by the kelps *Macrocystis pyrifera* and *Ecklonia arborea*. Recently (ca. 2006), *S. horneri* arrived on Catalina Island and has since increased in abundance, becoming the dominant alga on this and other rocky reefs across the island (Miller et al. 2007, Marks et al. 2015, G. Sullaway pers. obs.). This expansion was likely facilitated by a period of anomalous warm water and a strong ENSO that occurred between 2014 and 2016, during which time *M. pyrifera* and *E. arborea* densities decreased (Di Lorenzo & Mantua 2016, Marks et al. 2017). The understory algal assemblage on the study reef is composed largely of the brown algae *Zonaria farlowii*, *S. palmeri*, *Dictyota flabellata* and *Dictyoptertis undulata*, the fleshy red algae *Plocamium cartilagineum* and *Rhodymenia californica*, and the geniculate coralline algae *Bossia orbigniana* and *Calliarthron cheilosporioides*. The substrate is composed largely of non-geniculate coralline algae, *Lithothamnion* spp., and rocky cobble.

2.2. Benthic chamber construction

To measure rates of community production on the rocky reefs invaded by *S. horneri*, we deployed collapsible benthic chambers modeled after those described by Haas et al. (2013) and Calhoun et al. (2017) (Fig. 1). These chambers create closed systems over the benthos where oxygen, irradiance, and temperature sensors track photosynthesis and salient ocean conditions within a known volume of water (see also Miller et al. 2009, 2011, Rodgers et al. 2015). While many past experiments examining community production in kelp forest ecosystems have relied on laboratory experiments that do not incorporate natural fluctuations in abiotic conditions, recent studies have identified techniques that measure community production *in situ*, thereby increasing ecological realism (Tait & Schiel 2011, Rodgers et al. 2015, Olivé et al. 2016). For example, *in situ* chamber designs have been developed for estimating community production by individual species (Rodgers et al. 2015, Olivé et al. 2016) and whole benthic communities (e.g. Miller et al. 2009, Haas et al. 2013). In general, estimates of NCP, GCP, and CR for the benthos can be made by measuring changes in dissolved oxygen (DO) within chambers that are placed *in situ* over macroalgae and invertebrate communities. Then, by linking temporal

changes in oxygen concentrations within the chambers to incident temperature and irradiance conditions, we can relate variation in these measures to community structure and abiotic conditions (Miller et al. 2009). Further, because these chambers encompass whole benthic communities, species interactions (e.g. shading) and invertebrate respiration are incorporated into production measurements. These interactions are often not captured in laboratory experiments but are pertinent to understanding NCP (Bracken & Williams 2013).

We constructed benthic chambers using 0.106 cm polycarbonate plastic triangle sheets glued to fiberglass-reinforced vinyl panels using Weld-On® glue. The chambers were 1.2 m tall, and frames were reinforced using aluminum rods and stainless-steel cable, which held the interior volume and benthic area of each benthic chamber constant at 192 l and 0.575 m², respectively. Flexible polycarbonate walls prevented boundary layer formation by transferring wave energy into the interior of each benthic chamber. To ensure chambers were held to the seafloor, we secured anchor chain to vinyl skirts (skirts were



Fig. 1. (a) Benthic chambers deployed with temperature, oxygen, and PAR sensors. (b) Sensor stand with PAR sensor at the top and dissolved oxygen/temperature sensor at the bottom

30.5 cm long) around the chamber perimeters. We initially verified wave energy transfer and the effectiveness of chamber-substrate seals by observing fluorescein dye movement within the chambers and ensuring it did not leak to the surrounding environment (authors' pers. obs.). We used 'clod card' dissolution rates as a proxy for relative rates of water movement inside and outside chambers (Doty 1971, for further details and methods see Appendix). Mean dissolution inside chambers was 6% lower than outside the chambers, but this difference was not significant, indicating that the chamber walls were flexible enough to allow for water movement comparable to the environment outside the chamber (paired Wilcoxon test, $p = 0.055$). We constructed PVC sensor arrays that were equipped with a photosynthetically active radiation (PAR) sensor (Odyssey Dataflow Systems) and a DO (mg l^{-1}) and temperature ($^{\circ}\text{C}$) sensor (MiniDOT logger, PME) (Fig. 1). PAR sensors were placed approximately 0.5 m above the benthos and recorded every 10 s, while oxygen and temperature sensors were placed approximately 0.2 m above the benthos and recorded every minute. Following each chamber deployment, we determined the average hourly oxygen concentration and calculated changes in DO between each consecutive hour, which provided a slope that estimated NCP within each chamber (Miller et al. 2009). We used measurements at night (i.e. in the dark) to estimate respiration by the benthic communities (CR). Finally, we added NCP and CR to estimate GCP (see also Miller et al. 2011 and Olivé et al. 2016).

2.3. Assessing changes in algal assemblages and community production following *Sargassum horneri* removal

To assess how the removal of *S. horneri* affects native algal abundances and patterns of community production, we marked 6 circular plots (each 6 m diameter and 28.3 m^2) near the seawater intake pipes at Two Harbors on Catalina Island in May 2016. These experimental plots were allocated into 3 pairs (hereafter blocks), which were spaced approximately 15 m from each other along a 10 m isobath. One of the 2 plots per block was randomly assigned as a *S. horneri* removal plot, where all *S. horneri* were removed by hand using SCUBA every other month (i.e. in May 2016, August 2016, October 2016, February 2017, and April 2017) until April 2017. During these removals, all *S. horneri* were disposed of on land to ensure we did not further facilitate spread of the

alga. Prior to each *S. horneri* removal, all algae within each plot were quantified along 4 radially directed transects that were allocated in a 'spoke and wheel' design, where each spoke consisted of a 3 m transect originating from the center of the plot. Two 0.5 m^2 quadrats were placed at randomly selected positions along each transect while excluding the center meter of the circle, resulting in a total of 8 quadrats per plot. Within each quadrat, all algae were identified to species and enumerated. If *M. pyrifera* were present, all stipes were counted 1 m above their holdfasts and juveniles were counted by individual holdfasts.

To estimate the impact of *S. horneri* removal on benthic community production, we placed benthic chambers near the center of each experimental plot for 24 h. At the end of 24 h, all algae within the chambers were identified to lowest taxonomic level and enumerated, but otherwise left intact to allow for continued community production measurements within the plot. The abundance of all other species of macroalgae within the chamber footprints was recorded and biomass was estimated based on abundance–biomass relationships as determined at a nearby (<2 km away) 10 to 13 m deep non-SMCA rocky reef. Specifically, individuals of each species (*Z. farlowii* [$n = 35$], *D. undulata* [$n = 35$], *D. flabellata* [$n = 30$], *M. pyrifera* [$n = 15$], *E. arborea* [$n = 7$], and *S. palmeri* [$n = 30$]) were collected and weighed to determine an abundance–biomass relationship that could be used to non-destructively estimate algae weights.

2.4. Characterizing temporal patterns of community production in ecosystems invaded by *Sargassum horneri*

To understand temporal patterns of community production in ecosystems invaded by *S. horneri*, we haphazardly placed 10 to 12 chambers on the benthos along the 9–12 m isobath for 24 h periods in order to encompass full diurnal cycles. If a benthic chamber was randomly placed over an algal assemblage that included *M. pyrifera* (which was infrequent as *M. pyrifera* was initially rare at the study site), the chamber was sealed around the kelp stipes at its apex so that only the bottom ~1 m of thalli was included inside the chamber. These deployments were repeated every other month from March 2016 to April 2017, although logistic constraints and failures of the chamber-benthos seals resulted in lower sample sizes in summer and fall deployments (Table 1). During each deployment, the chambers

Table 1. Sample size for each benthic chamber deployment grouped by season

Season	Associated months	Sample size
Spring	March–May	23
Summer	June–August	12
Fall	September–November	10
Winter	November–March	17

were opened and flushed with fresh seawater every 4 h so that changes in water chemistry within them did not alter photosynthetic rates. After 24 h, all *S. horneri* within the chamber footprints was collected and weighed, and weight of native algae was estimated using previously described abundance biomass relationships due to SMCA permit restrictions.

2.5. Statistical analyses

All univariate statistical analyses were done using R-Studio (R Core Team), and all multivariate analyses were done using PRIMER-E ver. 6.0 (Clarke & Warwick 1994). All univariate data were assessed for normality and equality of variances by graphical examination of residuals and a non-significant result in Levene's tests, respectively. Multivariate data were evaluated for normality using Draftsman plots.

To evaluate the effects of *S. horneri* removal on native algal abundance through time, algal abundance data were first 4th root transformed to downweigh the influence of overly abundant species. Given the large amount of zeros in the data set, it was not possible to calculate Bray-Curtis similarities for all pairs of samples, and thus a Euclidean distance-based matrix was generated to characterize similarities among all sample pairs. Then, a 3-way blocked PERMANOVA with sample month and removal treatment considered as fixed variables and block considered as a random variable with replication was used to evaluate differences in algal abundances within quadrats among factor levels. Tests of significance between removal treatments were based on Monte Carlo tests due to the small number of unique permutations for that factor. Following this, pairwise permutation contrasts were used to evaluate differences between removal treatments during each month separately. These were accompanied by estimates of multivariate dispersions using the multivariate dispersion (MvDisp) procedure in Primer-E to evaluate how within treatment variability (i.e. among sample quadrats) compared between the removal treatments. SIMPER analyses were used

to estimate the relative importance (% contribution) of each algal species to the observed differences between the removal treatments during each sample month. Lastly, non-metric multidimensional scaling (nMDS) plots were generated to examine the relative similarities in algal assemblages between the removal treatments on each sample date. *S. horneri* abundance was removed from this analysis so we could specifically evaluate community-wide differences outside of the manipulated treatment. Separate 2-way Model I blocked ANOVAs were used to evaluate the effects of *S. horneri* removal on NCP, CR, and GCP. For each ANOVA, data were square root transformed to correct problems with homoscedasticity. Time and Treatment (Control and Removal) were considered fixed variables, and Block was considered a random variable.

To understand temporal changes in production metrics (NCP, GCP, and CR) and abiotic conditions on the reef invaded by *S. horneri*, data were log transformed to satisfy assumptions of normality and equality of variances, and separate 1-way ANOVAs were used to determine differences in each production metric, sea water temperature, *S. horneri* biomass, and mean PAR through time. For the analyses, we grouped time into 4 'seasons' based on trends in *S. horneri* biomass and in the hope of making the plots easier to interpret; however, we do not draw conclusions related to season as we were unable to replicate seasons.

3. RESULTS

3.1. Assessing changes in algal assemblages and community production following *Sargassum horneri* removal

The native algal assemblages within the *Sargassum horneri* removal and control plots varied through time (PERMANOVA: Month, pseudo- $F_{3,168} = 7.27$, $p < 0.001$) and between treatments (Treatment, Pseudo- $F_{1,168} = 3.40$, $p = 0.052$), but these factors interacted such that the relative differences between treatments varied among months (Month \times Treatment, pseudo- $F_{3,168} = 3.87$, $p = 0.024$; Table 2). When examined within each month separately, the removal and control plots did not differ during the first 2 time points (August, $t = 1.48$, $p = 0.159$; October, $t = 1.456$, $p = 0.237$) but they did differ during the latter time points (February, $t = 3.568$, $p = 0.004$; April, $t = 2.863$, $p = 0.002$; Fig. 2, Tables 2 & 3). This resulted in the assemblages being markedly different 11 mo after the clearings were established (Fig. 2). Further, the within-plot (i.e.

Table 2. (a) Blocked-PERMANOVA results examining dissimilarities in native algal assemblages between *Sargassum horneri* removal and non-removal control plots and (b) post-hoc comparison of Month \times Treatment interaction

a) PERMANOVA						
Source	df	SS	MS	Pseudo- <i>F</i>	p(perm)	% Variation
Treatment	1	149.13	149.13	3.4	0.05	4.80 %
Block	2	55.31	27.657	3.8	0.002	1.40 %
Month	3	758.77	252.92	7.26	0.001	1.97 %
Treatment \times Block	2	87.71	43.857	6.04	0.001	5.00 %
Treatment \times Month	3	372.68	124.23	3.866	0.024	16.70 %
Block \times Month	6	208.84	34.806	4.797	0.001	7.50 %
Treatment \times Block \times Month	6	192.79	32.13	4.428	0.001	13.50 %
Res (= Quadrat)	168	1219	7.256			

b) Post-hoc comparisons: Month \times Treatment			
	<i>t</i>	p(perm)	
August	1.484	0.159	
October	1.456	0.237	
February	3.568	0.004	
April	2.863	0.002	

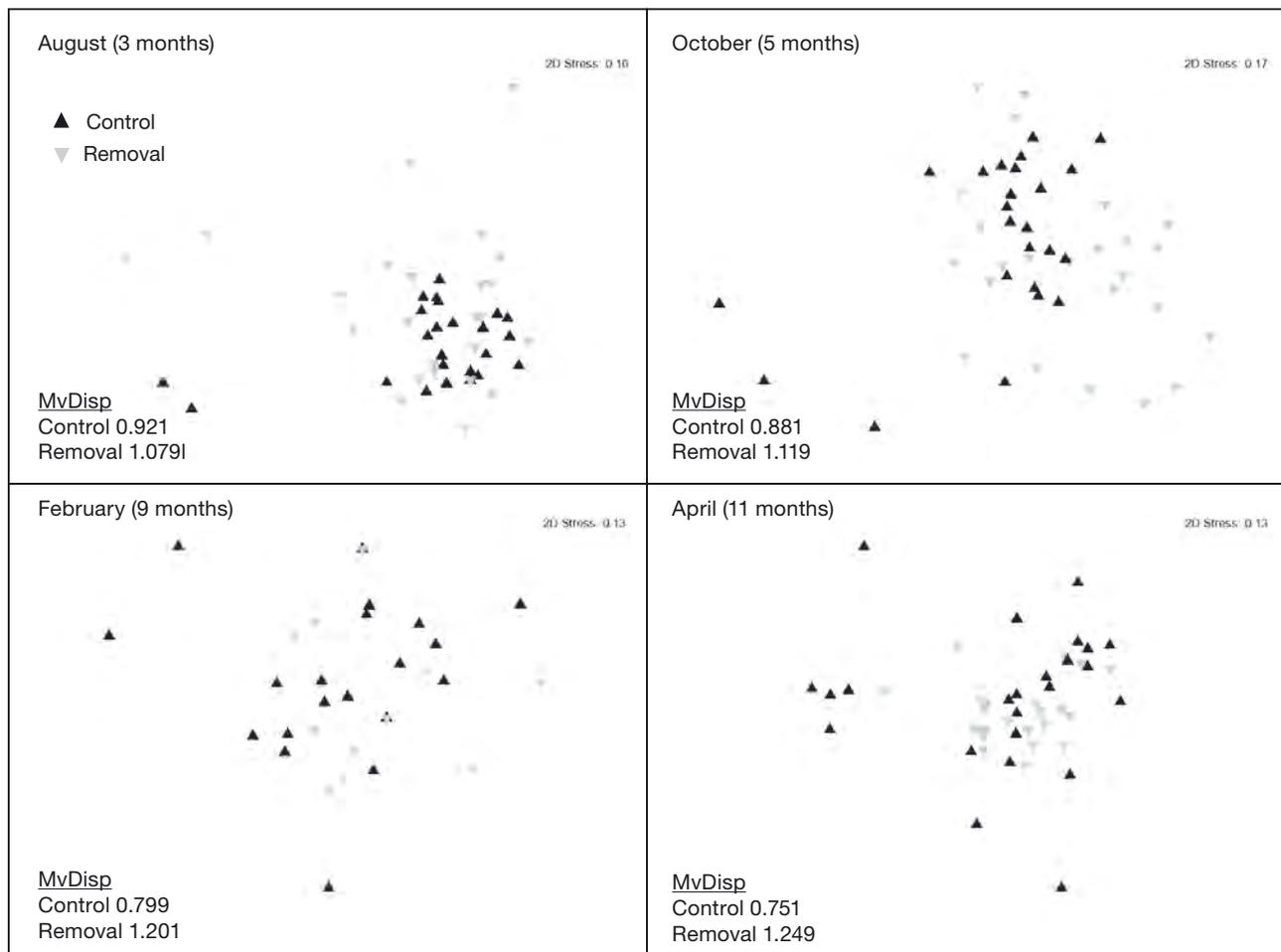


Fig. 2. Non-metric multidimensional scaling plot examining differences among algae communities in *Sargassum horneri* removal and control plots through time. Multivariate dispersion (MvDisp) indicates the variation in assemblages among sample unit, larger numbers indicate greater variability

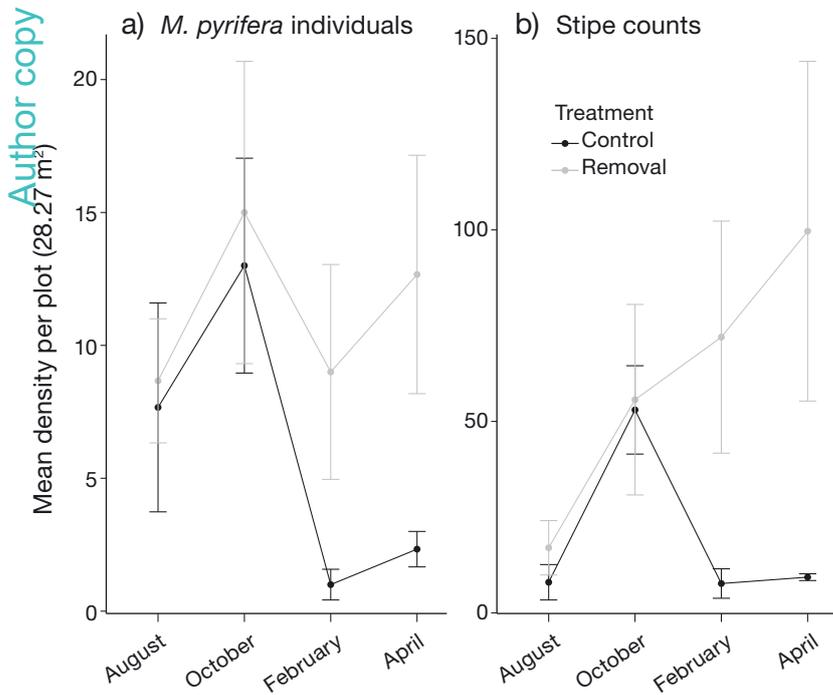


Fig. 3. Number of *Macrocystis pyrifera* individuals and stipes (mean \pm SE) per plot through time

quadrat to quadrat) variation in these assemblages was greater in the removal plots and appeared to increase slightly over the experiment, while at the same time it was lower and appeared to decrease slightly in the control plots (quantified using MvDisp, Fig. 2). Specifically, juvenile kelps were 3.7 times more abundant in the *S. horneri* removal plots than the control plots (2.92 ± 0.47 ind. vs. 0.79 ± 0.29 ind. per 28.3 m², respectively; mean \pm SE) after 11 mo. Additionally, mean densities of *M. pyrifera* individuals and stipes were greater in removal plots than control by the conclusion of the experiment (Fig. 3). Likewise, geniculate coralline algae, understory brown algae such as *Zonaria farlowii* and *Dictyopteris undulata*, and fleshy red algae such as *Plocamium cartilagineum* were more abundant in *S. horneri* removal plots upon conclusion of the experiment (Table 3).

Interestingly, observed differences in algal community composition did not translate to strong differences in community production between *S. horneri* removal and control

Table 3. Average abundance and % contribution from SIMPER analysis of algae between control and removal plots through time. Diss: dissimilarity

Survey month	Species	Average abundance				
		Removal	Control	Diss/SD	% Contribution	% Cumulative
August (3 mo)	<i>Corallina</i> spp.	0.294	0.577	0.45	27.51	27.51
	<i>Dictyopteris undulata</i>	2.22	2.13	0.69	15.38	42.89
	<i>Sargassum palmeri</i>	0.899	1.11	0.91	13.96	56.86
	<i>Zonaria farlowii</i>	0.827	0.929	0.93	12.78	69.64
	<i>Rhodomenia californica</i>	0	0.477	0.39	9.76	79.39
	Juvenile kelp	0.267	0.25	0.74	4.15	83.55
	<i>Eisenia arborea</i>	0.197	0.142	0.5	3.67	87.22
	<i>Cystosiera osmundaceae</i>	0	0.184	0.33	2.97	90.19
October (5 mo)	<i>Zonaria farlowii</i>	1.66	1.18	0.82	19.7	19.7
	<i>Dictyopteris undulata</i>	1.22	1.8	0.91	19.66	39.36
	<i>Sargassum palmeri</i>	0.792	1.13	0.81	19.22	58.57
	<i>Corallina</i> spp.	0.072	0.431	0.39	15.31	73.89
	<i>Macrocystis pyrifera</i>	0.309	0.042	0.45	5.06	78.94
	<i>Rhodomenia californica</i>	0.101	0.185	0.36	4.94	83.88
	<i>Sargassum muticum</i>	0.317	0	0.48	4.9	88.78
	<i>Colpomenia peregrina</i>	0.244	0	0.33	4.08	92.87
February (9 mo)	<i>Zonaria farlowii</i>	0.786	0.285	0.88	31.73	31.73
	<i>Dictyopteris undulata</i>	0.498	0.339	0.62	26.09	57.82
	<i>Sargassum palmeri</i>	0.55	0.451	0.78	22.09	79.91
	<i>Eisenia arborea</i>	0.208	0.083	0.59	6.46	86.37
	<i>Corallina</i> spp.	0	0.142	0.28	6.29	92.66
April (11 mo)	<i>Corallina</i> spp.	0.922	0.215	3.85	28.61	28.61
	Juvenile kelp	1.22	0.392	2.75	20.04	49
	<i>Dictyopteris undulata</i>	1.42	1.11	2.47	18.33	67.34
	<i>Zonaria farlowii</i>	1.08	0.618	1.87	13.88	81.22
	<i>Plocamium cartilagineum</i>	0.203	0	0.667	4.95	86.17
	<i>Sargassum palmeri</i>	0.269	0.368	0.594	4.41	90.58

plots over the course of the experiment (Fig. 4, Table 4). We did not find statistically significant differences among treatments for any production metric (2-way blocked ANOVAs: NCP–Treatment $F_{1,15} = 2.739$ $p = 0.119$; GCP–Treatment $F_{1,15} = 1.30$ $p = 0.272$; CR–Treatment $F_{1,15} = 3.73$ $p = 0.071$). However, patterns emerged indicating that *S. horneri* may have an additive impact on overall metabolism at times when it is rapidly growing and has high biomass. For instance, in fall, GCP and CR were 1.6 and 1.9 times greater, respectively, and *S. horneri* was 7.7 times denser in the control plots than the removal plots (control: 147 ± 30 ind. m^{-2} ; removal: 19 ± 2.8 ind. m^{-2}) (Fig. 4). This increase in metabolism range occurred when *S. horneri* was growing and likely created an additive rather than substantive contribution to GCP and CR. This change in magnitude was not reflected in the overall

NCP (Fig. 4), indicating that all 3 measurements of production are valuable when evaluating ecosystem function through community production.

3.2. Characterizing temporal patterns of community production in ecosystems invaded by *Sargassum horneri*

Overall, community production on a subtidal reef invaded by *S. horneri* followed a temporal pattern

Table 4. Results from a 2-way randomized blocked-ANOVA test for differences in production metrics between removal and control plots. NCP: net community production; GCP: gross community production; CR: community respiration. Statistically significant values ($p < 0.05$) in **bold**

Production metric	p:Season	p:Treatment	p:Replicate	p:Season × Treatment
NCP	0.130	0.119	0.480	0.805
GCP	0.048	0.272	0.500	0.552
CR	0.022	0.071	0.749	0.636

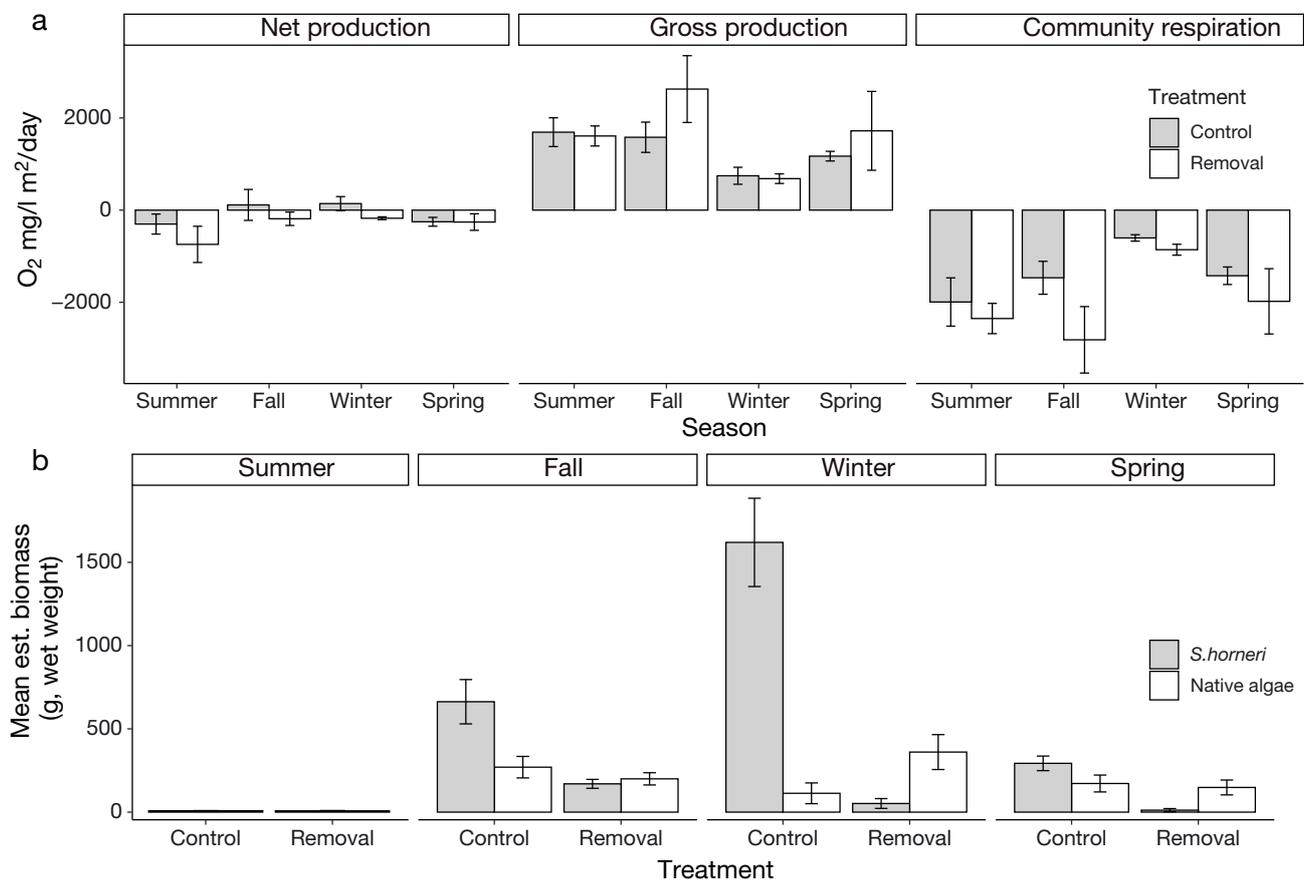


Fig. 4. (a) Seasonal mean net community production, gross community production, and community respiration, and (b) algae biomass between control and *Sargassum horneri* removal plots by season (± 1 SE)

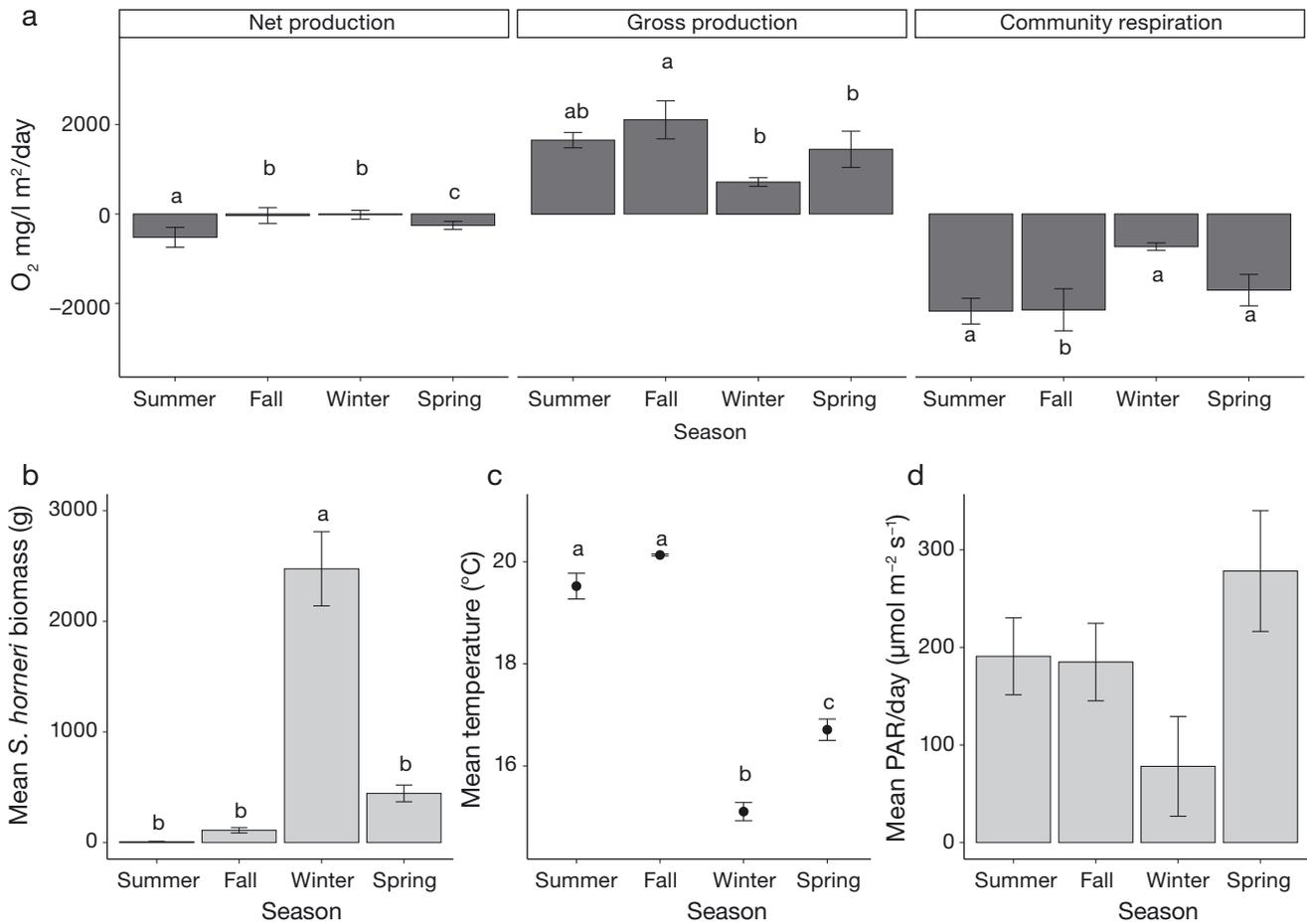


Fig. 5. Means of (a) community production (net and gross) and respiration, (b) *Sargassum horneri* biomass, (c) ocean temperature, and (d) photosynthetically active radiation (PAR) on a reef invaded by *S. horneri* (± 1 SE). Different letters denote significant differences according to Tukey's post-hoc test

Table 5. Results from univariate ANOVA testing for differences in production metrics (NCP, GCP, CR) and environmental variables (biomass, temperature, and PAR) on a reef invaded by *Sargassum horneri*. Statistically significant values ($p < 0.05$) in **bold**. See Table 4 for definitions of production metric abbreviations

Dependent variable	p:season
NCP	<0.001
GCP	<0.001
CR	<0.001
<i>S. horneri</i> biomass (g)	<0.001
Temperature ($^{\circ}\text{C}$)	<0.001
PAR ($\mu\text{mol m}^{-2} \text{s}^{-1}$)	0.107

that was coincident with the annual life cycle of the alga. Specifically, GCP and CR were both greatest during fall (GCP: $2709.5 \pm 320.06 \text{ mg O}_2 \text{ l}^{-1} \text{ m}^{-2} \text{ d}^{-1}$; CR: $-2368.8 \pm 272.2 \text{ mg O}_2 \text{ l}^{-1} \text{ m}^{-2} \text{ d}^{-1}$), and lowest during winter (GCP: $1038.6 \pm 538.8 \text{ mg O}_2 \text{ l}^{-1} \text{ m}^{-2} \text{ d}^{-1}$; CR: $-748.15 \pm 138.1 \text{ mg O}_2 \text{ l}^{-1} \text{ m}^{-2} \text{ d}^{-1}$; mean \pm SE)

(Fig. 5a, Table 5). Additionally, the greatest difference between GCP and CR, which may be the best estimate of the impact of the alga on ecosystem function (M. S. Edwards et al. unpubl.), occurred during the fall when the *S. horneri* population was rapidly growing and the alga presumably had heightened metabolic activity and increasing biomass (Gómez & Wiencke 1996). Mean benthic PAR varied temporally but these differences were not significant through time (Fig. 5d). Temperature was greatest in summer and fall (Fig. 5c, Table 5), coinciding with higher rates of GCP (Fig. 5) on the reef.

4. DISCUSSION

We found that *Sargassum horneri* presence contributed to ephemeral increases in GCP and CR that were not reflected in NCP, indicating that all 3 measurements of production are valuable when evaluat-

ing ecosystem function through community production (Edwards et al. 2020). Additionally, *S. horneri* removal led to increases in native algal abundances during a period of favorable abiotic conditions, while cooler nutrient-rich waters, in concert with *S. horneri* removal, likely facilitated observed increases in native algal abundances in our *S. horneri* removal plots. Compared to our removal plots, control plots exhibited minimal *Macrocystis pyrifera* recruitment or native algal growth, indicating that once *S. horneri* is established, it can impede native algal abundance. While *M. pyrifera* proliferates in water temperatures ranging from 8 to 17°C, (North et al. 1986, Schiel & Foster 2015), *S. horneri* has been documented in temperatures ranging from 18 to 22°C and has established in the Eastern Pacific in temperatures ranging from 14 to 22°C (Chu et al. 1998, Marks et al. 2015). Our study began on the tail end of an anomalously warm water period (Reed et al. 2016), when *M. pyrifera* was nearly absent from the subtidal community at Catalina Island. At this time, mean water temperatures at our study site in summer 2016 were $19.5 \pm 0.25^\circ\text{C}$ (Fig. 5c) and we did not observe *M. pyrifera* in our study site. When water temperatures cooled down to $15.1 \pm 0.17^\circ\text{C}$ in winter 2016, we began to see *M. pyrifera* growth (Fig. 3). During this time, we observed a 9-fold increase in *M. pyrifera* stipe density and a 3-fold increase in juvenile kelp abundance in areas where *S. horneri* had been removed. *M. pyrifera* densities at our study site (Removal: $3.68 \pm 1.64 \text{ ind. m}^{-2}$; Control: $0.34 \pm 0.03 \text{ ind. m}^{-2}$, Fig. 3a) were overall much lower than what McAlary et al. (1998) observed on Catalina Island in the mid-1990s (10 to 12 ind. m^{-2}) or what Edwards (2004) observed for several kelp forests throughout the southern California mainland (i.e. 6 to 12 ind. m^{-2}). Moreover, the dominant understory brown algae observed during this study, *Zonaria farlowii* and *Dictyopteris undulata*, both have an affinity for warm water conditions (Marks et al. 2017). We suggest that suitable abiotic conditions in concert with *S. horneri* removal are related to the increase in native algae and especially *M. pyrifera* abundance at our site. It appears that *S. horneri* may be taking advantage of an environmental shift and acting as a ‘passenger’ to abiotic change rather than driving an ecosystem shift (MacDougall & Turkington 2005).

In this study, we estimated community production, which means we cannot explicitly identify the extent to which production is related to changes in algal abundance associated with *S. horneri* removal versus changes to the heterotrophic communities as they are associated with certain algal species. We hypothesize

that differences in production between removal and control plots are strongly related to changes in algae production rather than changes to the meso-consumer community. *S. horneri* is a structurally complex species that likely supports higher invertebrate abundances than *M. pyrifera* per unit area. Studies on the morphologically similar *S. muticum* suggest that it supports high abundances of invertebrates compared to less structurally complex native species (Gestoso et al. 2010). However, it should be noted that *S. horneri* has a chemical anti-fouling component that discourages invertebrate settlement. This may limit invertebrate abundances on the algae (Cho 2013), which could lead to lower than expected invertebrate abundances. There are likely differences in invertebrate communities between *S. horneri*-dominated assemblages and native assemblages, but this is not expected to significantly shift the results of this study; if anything, it may result in our underestimating rates of production from algae. This is a clear area for future research and would contribute to a greater understanding of *S. horneri* impacts on ecosystem function.

We hypothesize that the lack of differences in NCP between experimental treatments is due, in part, to compensatory production by understory algae. Similarly, Miller et al. (2011) did not observe differences in production between *M. pyrifera* removal plots compared to control plots. They hypothesized that compensatory production by understory algae and phytoplankton occurs with canopy removal and may dampen expected variability in production (Miller et al. 2011). Sub-canopy algae, such as *S. horneri*, do not form an extensive surface canopy comparable to that of *M. pyrifera* and this may limit the ability for compensatory production (Pfister et al. 2019). While an ecosystem’s ability to compensate for these types of shifts requires further research, it indicates an impressive capacity to buffer short-term disturbances (Miller et al. 2011, Lamy et al. 2019). Differences in community production between *S. horneri* removal and control plots were ephemeral. Compared to *S. horneri* removal plots, control plots had greater GCP and CR (Fig. 4) in fall when *S. horneri* had the greatest growth rates and was increasing in biomass (Fig. 4, Marks et al. 2018). While NCP is often used to measure changes in community production and ecosystem function, we found support for the idea that GCP and CR may better reflect changes to ecosystem function (Edwards et al. 2020).

Community production metrics (NCP, GCP, and CR) were temporally variable in an ecosystem fully inundated with *S. horneri*, and this variability is likely related to changes in *S. horneri* biomass and

water temperature. The annual life history of *S. horneri* means that biomass varies significantly over the course of a year. This annual lifecycle is notably different from that of the perennial *M. pyrifera* and other native understory algal species found in southern California subtidal rocky reefs (excluding *Desmarestia ligulata*). Research in other subtidal macroalgae ecosystems have found that the biomass of the dominant species drives relative rates of community production (Miller et al. 2011, Rodgers & Shears 2016). In our study, *S. horneri* was the dominant species for the majority of the year throughout the ecosystem, except when senescent in summer. We found that the highest CR rates occurred during periods with warmer water and greater *S. horneri* growth and biomass (Fig. 5a,c). Similarly, Tait & Schiel (2013) found that natural fluctuations in temperature led to increased community respiration and decreased NCP. In our study, increased respiration was balanced by increased GCP, so we did not see decreased rates of NCP. However, this compensation may not occur over longer time scales and has the potential to alter ecosystem NCP (Tait & Schiel 2013).

When considering management action that focuses on non-native alga removal, it is important to note that our removal plots were relatively small (28.3 m²). Marks et al. (2017) recommends plots larger than 60 m² for effective *S. horneri* management. This requires a large amount of time, effort, and funding. Native algae responses to non-native species removal may have varied results depending on oceanographic conditions. Removal of a non-native species in ocean conditions not amenable to native species growth may result in an ecosystem lacking structure and community production, and in fact may further facilitate non-native species establishment. Instead, we recommend removal occurs in conditions favorable to native algae growth.

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Appendix. Relative water motion measured with clod cards
— methods and results

We constructed clod cards using plaster of Paris (mixed sensu Doty 1971) in tapered cylindrical molds and dried them for 48 h at 25°C. Next, we glued plaster to tiles and took initial weight. We deployed the cards inside and outside chambers for 24 h ($n = 27$) over the course of 3 consecutive days at the onset of this project. After 24 h, we removed the tiles and dried them for 48 h at 25°C.

We subtracted the pre- and post-deployment weights to get the dissolution rate in 24 h and compared this based on card location (inside or outside chamber) using a paired Wilcoxon test. We used a non-parametric test because both a Shapiro-Wilk test and quantile plot indicated that data were slightly heteroskedastic. We found that mean dissolution inside the chambers was 6% lower than outside the chambers, but this difference was not significant, indicating that the chamber walls were flexible enough to allow for substantial water movement (paired Wilcoxon test, $p = 0.055$; Fig. A1).

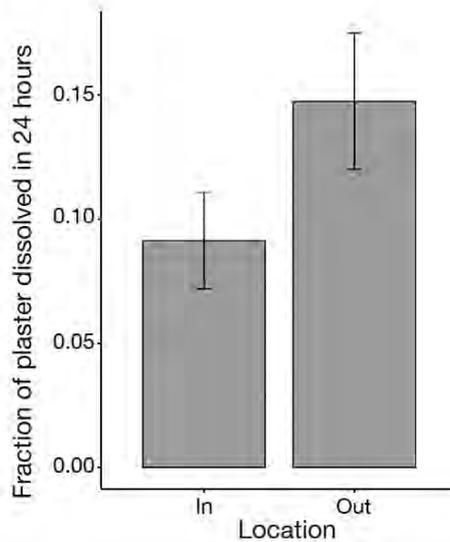


Fig. A1. Clod card dissolution rates inside and outside of chamber (paired Wilcoxon test, $p = 0.055$)



Implementing the MLMA Master Plan

Prioritizing Key California Invertebrate Fisheries



Debbie Aseltine-Neilson and Tom Mason
California Department of Fish and Wildlife

Marine Resource Committee Meeting
November 10, 2020

Fishery Prioritization

Goal: Identify state-managed fisheries with the greatest need for management attention to inform implementation of the MLMA Master Plan scaled management approach

*Focus CDFW's limited resources on greatest management need

Prioritization Method

Prioritization process consists of 2 components:

1. Productivity Susceptibility Analysis (PSA): Assess productivity of a species and its susceptibility to current fishing practices
 2. Ecological Risk Assessment (ERA): Assess impacts of fishery on bycatch and habitat
- Scores from PSA, Bycatch, and Habitat added to get total score/ranking

Productivity Susceptibility Analysis

- Developed through collaboration between CDFW and partners (2015-2016)
- Applied to 45 state-managed fisheries with greatest catch or landings (15 invertebrates)
- Interim priority list in 2018 Master Plan based on PSA results only

Ecological Risk Assessment

- Iterative tool development, involving partners and stakeholders
- Tool further refined by CDFW to be more streamlined and intuitive

Fishery Prioritization Status

- Finfish
 - Prioritization results presented to Commission in Fall 2019
 - High priority fisheries ran through scaled Management Approach and results provided to FGC in Feb 2020
- Today- Prioritization for 15 invertebrate fisheries

Results: Combined PSA and ERA

Species	Sector	Gear	Rank Total	PSA Rank	ERA Rank Bycatch	ERA Rank Habitat
Giant Red Sea Cucumber	Comm	Trawl	5	2	2	1
Ocean Pink Shrimp	Comm	Trawl	7	3	3	1
Ridgeback Prawn	Comm	Trawl	7	4	2	1
Spiny Lobster	Comm	Trap	7	2	3	2
Spot Prawn	Comm	Trap	9	4	3	2
Kellet's Whelk	Comm	Trap	9	4	3	2
Dungeness Crab	Comm	Trap	10	5	2	3
Spiny Lobster	Rec	Hoop net	10	3	4	3
Warty Sea Cucumber	Comm	Dive (Hand)	10	1	4	5
Red Abalone	Rec	Dive (Hand/Iron)	10	1	4	5
Rock Crab	Comm	Trap	10	5	3	2
Dungeness Crab	Rec	Trap	11	5	3	3
Pacific Geoduck Clam	Rec	Shovel	11	3	4	4
Red Sea Urchin	Comm	Dive (Hand/Rake)	13	4	4	5
Pismo Clam	Rec	Clam Fork	14	5	4	5

PSA Results

Species	Sector	Gear	PSA Rank
Warty Sea Cucumber	Comm	Dive (Hand)	1
Red Abalone	Rec	Dive (Hand/Iron)	1
Giant Red Sea Cucumber	Comm	Trawl	2
Spiny Lobster	Comm	Trap	2
Ocean Pink Shrimp	Comm	Trawl	3
Spiny Lobster	Rec	Hoop net	3
Pacific Geoduck Clam	Rec	Shovel	3
Kellet's Whelk	Comm	Trap	4
Spot Prawn	Comm	Trap	4
Red Sea Urchin	Comm	Dive (Hand/Rake)	4
Ridgeback Prawn	Comm	Trawl	4
Rock Crab	Comm	Trap	5
Dungeness Crab	Comm	Trap	5
Pismo Clam	Rec	Clam Fork	5
Dungeness Crab	Rec	Trap	5

ERA Results

Species	Sector	Gear	Ecological Risk	ERA Rank Bycatch	ERA Rank Habitat
Ridgeback Prawn	Comm	Trawl	3	2	1
Giant Red Sea Cucumber	Comm	Trawl	3	2	1
Ocean Pink Shrimp	Comm	Trawl	4	3	1
Dungeness Crab	Comm	Trap	5	2	3
Spot Prawn	Comm	Trap	5	3	2
Rock Crab	Comm	Trap	5	3	2
Spiny Lobster	Comm	Trap	5	3	2
Kellet's Whelk	Comm	Trap	5	3	2
Dungeness Crab	Rec	Trap	6	3	3
Spiny Lobster	Rec	Hoop net	7	4	3
Pacific Geoduck Clam	Rec	Shovel	8	4	4
Warty Sea Cucumber	Comm	Dive (Hand)	9	4	5
Red Sea Urchin	Comm	Dive (Hand/Rake)	9	4	5
Red Abalone	Rec	Dive (Hand/Iron)	9	4	5
Pismo Clam	Rec	Clam Fork	9	4	5

Priority for Scaled Management

Consider for Scaled Management

Ridgeback Prawn

Ocean Pink Shrimp*

Giant Red Sea Cucumber

Spot Prawn

Rock Crab

Warty Sea Cucumber

Red Sea Urchin

In progress/other considerations

Red Abalone* (FMP in progress)

Spiny Lobster* (FMP in place)

Dungeness Crab* (complex management system)

Implementation: Scaled Management

- Scaled management addresses the questions:
 - What happens next for fisheries identified as higher priority?
 - What is the appropriate management action?
- Seeks to match the level of management effort with the management needs and complexity of the fishery
- During process, also will address fisheries or factors not contemplated in the prioritization process (e.g. emerging fishery or resource concerns)

Next Steps

- Determine what scale of management is appropriate
- Update the MLMA Work Plan and provide an update to the Commission





Thank You

Debbie.Aseltine-Neilson@wildlife.ca.gov, Tom.Mason@wildlife.ca.gov



CDFW- D.Stein



California Pink Shrimp Fishery Management Plan

Marine Resources Committee

November 10, 2020

Sonke Mastrup
Environmental Program Manager
Sonke.Mastrup@wildlife.ca.gov

Why Develop a Fishery Management Plan?

In the process of developing underlying elements required for Marine Stewardship Council (MSC) certification, CDFW determined we had the required elements of a FMP



Photo: ODFW



Photo: CDFW

What Scale of Management is Appropriate?

MLMA allows for scaled management approach based on fishery prioritization and complexity and degree of management change.

Management range continuum:

- ESR
- ESR & Rulemaking
- ESR & Basic FMP
- ESR & Complex FMP

Test Case for Basic Fishery Management Plan

Why develop a Basic FMP?

- Fishery is relatively simple
- Most of the work is already completed
 - ESR completed in 2019
 - Other elements of FMP have been developed as part of MSC application and can be implemented through a future rulemaking
- Less costly and workload intensive than complex FMP

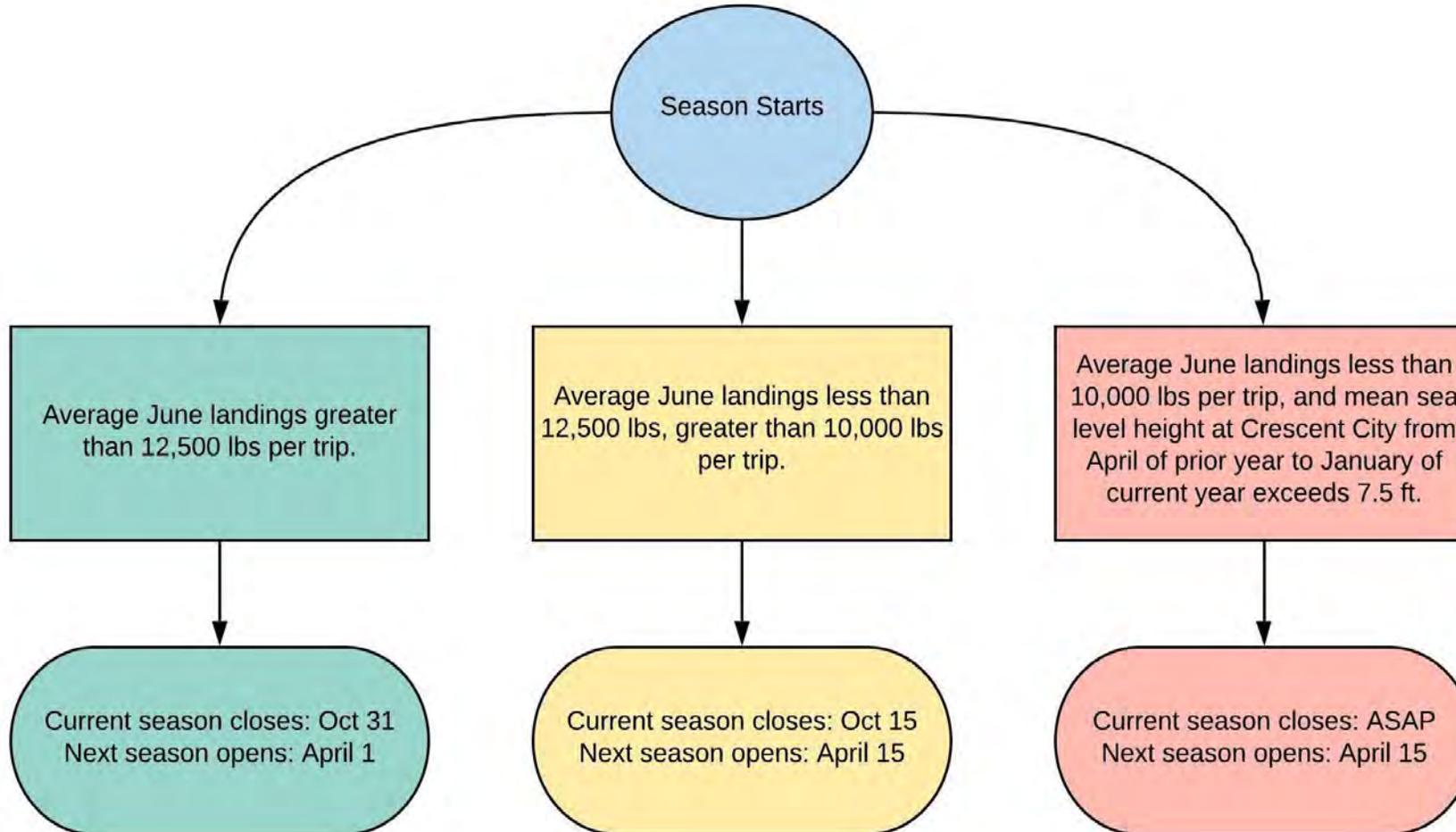
Requirements of Fishery Management Plans

Element	Location of Information
Fishery description	ESR
Essential Fishery Information	ESR
Conservation Measures	ESR & Future Rulemaking
Habitat	ESR & Federal Regulations
Bycatch/discards	ESR + Future Rulemaking (LED lights)
Overfishing and stock rebuilding	Future Rulemaking (Harvest Control Rule)
Amending FMP	Future Rulemaking

Future Rulemaking

- Harvest Control Rule
- LED lights

Harvest Control Rule



LED Lights

No LED lights used



LED lights used



Photos: ODFW

Progress to Date

- CDFW hosted several meetings with industry to solicit input on harvest control rule and mandatory use of LED lights
- Working with processors to collect biological samples
- Coordinating with ODFW
- Finalizing draft of Basic Fishery Management Plan

Considerations in Moving Forward

(Realities of Department Capacity)

- Determine additional outreach for Basic FMP (including CEQA, Tribal) and process (FGC, MRC)
- Staff capacity is limited, but industry is willing to fund a consultant to conduct an independent peer review process
- Develop regulatory documents for future rulemaking to adopt harvest control rule, LED lights and capacity changes

Proposed Approach

(Estimated Timeline = 1.5 yr)

- Additional stakeholder outreach on Basic FMP
 - Tribal outreach
 - Post draft on webpage and solicit input/comments
 - Provide response to comments
- Complete drafting of Basic FMP and initiate peer review process
- FGC adopt Basic FMP
- FGC adopt implementing rulemaking

QUESTIONS?



Brian L. Stone
President

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October 6, 2020

Melissa Miller-Henson, Executive Director
California Fish and Game Commission
1416 Ninth Street, Suite 1320
Sacramento, CA 95814

Dear Director Miller-Henson,

The Board of Harbor Commissioners is requesting your assistance to help to accelerate the Fish and Game Commission review and approval of the Fisheries Management Plan for Pink Shrimp which was completed by Department of Fish and Wildlife. The approval of this Plan by the Commission is the final step needed to complete the Marine Stewardship Council's (MSC) recognition of California's Pink Shrimp fishery as a Certified Sustainable Fishery.

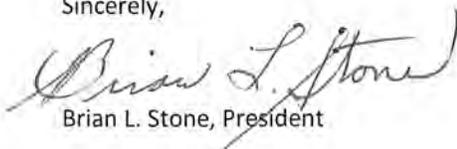
California is the only one of the three West Coast states in the Lower 48 to not be certified. Oregon has been certified for Pink Shrimp for thirteen years and Washington, which started the certification process at the same time as California, has been certified as a Sustainable Fishery for Pink Shrimp for four years.

Since the beginning of the Pink Shrimp season this year, 5.3 million pounds of California Pink Shrimp, shrimp that was caught in California waters, has been offloaded in Brookings, Oregon because that state has the MSC Sustainable Fishery certification for pink shrimp. Last season, 7 million pounds of the 9 million pounds of pink shrimp caught in California waters was offloaded in Oregon and Washington. Our California commercial fishers have to pay a 2.4% fee assessed on the total value of the catch offloaded in Oregon to the Oregon Trawl Commission. This financial burden is in addition to the cost of extra fuel, wear and tear on machinery and crew time necessary to offload in Oregon.

Please help expedite your review and approval of the Fisheries Management Plan so that our state fishery can begin to benefit from the Marine Stewardship Council's issuance of a Certified Sustainable Fishery designation for California Pink Shrimp. We desperately need this sustainable certification to help support our California Pink Shrimp Fishery.

Please feel free to contact me with any questions. Thank you for your help.

Sincerely,


Brian L. Stone, President

Ocean Protection Council (OPC) Agency Update **Received from OPC staff on July 16, 2020**

Marine Aquaculture Update

CDFW, OPC and FGC staff are continuing to coordinate on marine aquaculture following the FGC's receipt of the Aquaculture Information Report at the FGC's June 24th meeting. OPC will convene state agency leadership including but not limited to leaders from CDFW, FGC, California Coastal Commission, State Water Resources Control Board, California Department of Public Health, and California Department of Food and Agriculture between July and September 2020 to develop Aquaculture Principles for aligned state agency principles on marine aquaculture. These Aquaculture Principles will inform the more comprehensive statewide Aquaculture Action Plan, which will be recommended for funding by the OPC and developed in alignment with Target 4.2.1 of OPC's [Strategic Plan](#): "With the California Department of Fish and Wildlife and others, develop a statewide aquaculture action plan focused on marine algae and shellfish and land-based/recirculating tank operations of marine algae, shellfish, and finfish by 2023. The plan should identify areas of opportunity and avoidance to minimize impacts to habitat, biodiversity, and wild fisheries and should include minimum project criteria, including best practices for eliminating detrimental environmental impacts."

OPC June 19th Public Meeting Update

[OPC's public meeting on June 19th](#) included approved funding for various projects that advance OPC's [Strategic Plan](#) goals.

Statewide Kelp Recovery Research Program: Of particular note in connection with FGC priorities, OPC approved the disbursement of \$600,000 to California Sea Grant to create a [statewide kelp recovery research program](#). Together with \$1,200,000 in match funding from California Sea Grant, OPC funding will support solutions-oriented research projects aimed at restoring and protecting kelp ecosystems statewide, selected via a competitive call. A summary of the six individual projects recommended for approval as part of the Kelp Recovery Research Program is available [here](#).

Reducing the Risk of Entanglement in Fishing Gear and Gear Innovation Testing:

OPC also approved a total of \$2,900,000 across four projects that support [reducing the risk of whale and sea turtle entanglement in fishing gear](#). One project will initiate a competitive grant program, administered by Pacific States Marine Fisheries Commission in close partnership with OPC, to implement scientific research projects and collaborative partnerships that directly support OPC's [Strategy](#) for reducing the risk of entanglement. This project would aim to initiate the competitive process in fall 2020 and aim to support individual projects after selection in spring 2021. Individual projects could be supported until around January 2024 (allowing for projects of up to ~2.5 years). The second project will provide up to \$500,000 to the National Marine Sanctuary Foundation to support both pop-up and non-pop-up [gear innovations testing](#) within the Dungeness crab fishery during the 2020-2021 fishing season. The scientific project design will be refined before the start of the fishing season, with input solicited from the Dungeness Crab Fishing Gear Working Group.

COMMITTEE STAFF SUMMARY FOR MARCH 17, 2020 MRC*For background purposes only***6. REGULATIONS FOR COMMERCIAL HARVEST OF KELP AND ALGAE****Today's Item**Information Action

Discuss DFW-proposed regulation changes concerning commercial harvest of wild kelp and algae and consider potential committee recommendation.

Summary of Previous/Future Actions

- | | |
|---|--------------------------------------|
| • FGC approved 3-phase approach for wild kelp and algae regulation review | June 20, 2012; Mammoth Lakes |
| • FGC adopted Phase 1 kelp regulations | Nov 6, 2013; La Quinta |
| • MRC reviewed approach to next regulation phases | Nov 4, 2015; MRC, Ventura |
| • FGC approved revised 3-phase approach | Dec 9, 2015; San Diego |
| • DFW updated MRC on new Phase 2 regulation review | Nov 15, 2016; MRC, Los Alamitos |
| • DFW provided updates on regulation review | 2018-2019; MRC, various |
| • Today's discussion and potential recommendation | Mar 17, 2020; MRC, Santa Rosa |
| • Notice hearing | August 19-20, 2020; Fortuna |
| • Discussion/adoption hearing | October 14-15, 2020; Oakland |

Background

Kelp, an important biogenic habitat, is managed with other marine algae through DFW's kelp management program. In Jun 2012, FGC and DFW agreed to revise antiquated commercial kelp regulations over several years through a three-phase approach, to improve management and enforceability. Phase 1 was completed in 2013 and implemented in 2014; DFW commenced with Phase 2 in late 2016.

Phase 2 has focused on both regulatory clean-up and broader management and regulation overhaul in consultation with kelp and algae harvesters, which DFW highlighted through updates to MRC in Mar 2018 and Jul 2019.

During phase 2, DFW conducted direct outreach to kelp and algae harvesters, solicited feedback from stakeholders at MRC meetings, and engaged directly with individual tribes and tribal communities and through the FGC Tribal Committee. Concerns raised during public and tribal engagement focused, in part, on the extensive loss of bull kelp on the north coast, and how the recent impacts should be incorporated into DFW's kelp harvest management. DFW has integrated additional management proposals intended to be responsive to the ecosystem changes and public input received, which will be described at today's meeting.

COMMITTEE STAFF SUMMARY FOR MARCH 17, 2020 MRC

For background purposes only

In Nov 2019, DFW presented an overview of the types of regulatory changes proposed for the Phase 2 rulemaking and highlighted a potential rulemaking timeline for consideration. In Feb 2020, FGC approved an updated rulemaking timeline as proposed.

Today MRC will receive a presentation from DFW staff on specific proposed regulatory changes in seven management categories (Exhibit 1) and discuss possible recommendations.

Significant Public Comments

A non-governmental organization expressed support for the proposed statewide closure of bull kelp harvest; in conjunction with the bull kelp closure, it recommends that harvest provisions associated with bull kelp be removed and that administrative kelp beds within the bull kelp range be changed to a closed status to avoid public confusion (Exhibit 2).

Three edible seaweed harvesters do not believe they have had adequate time to fully engage in the regulation development process following DFW's harvester survey, and request 1) a delay in the rulemaking timeline until autumn*; 2) time to present at the Mar MRC meeting; 3) accommodation for participation via webinar; and 4) access to DFW survey results (Exhibit 3). (*Note that in Feb 2020, FGC adjusted the rulemaking timeline to Aug/Oct, which may satisfy this request.)

Recommendation

FGC staff: Consider public input and develop a recommendation to support advancing draft regulations to a rulemaking stage with proposed changes recommended by DFW.

Exhibits

1. DFW presentation
2. Email from Gillian Lyons, Pew Charitable Trusts, received Feb 18, 2020
3. Email from Terry D'Selkie, Ocean Harvest Sea Vegetables, Larry Knowles, Rising Tide Sea Vegetables, and James Jungwirth, Naturespirit Herbs, received Feb 14, 2020

Committee Direction/Recommendation

The Marine Resources Committee recommends that the Commission support proposed regulation measures for commercial kelp and algae harvest as recommended by the Department and discussed today.

OR

The Marine Resources Committee recommends that the Commission support proposed regulation measures for commercial kelp and algae harvest as recommended by the Department and discussed today, except _____.



October 30, 2020

Dear Pacific Herring Commercial Permit Holder:

Subject: 2020-21 Commercial Pacific Herring Quotas for All Areas and Sectors

Per Title 14 California Code of Regulations (CCR) Article 6, which implements the California Pacific Herring Fishery Management Plan (Herring FMP), the Director of the California Department of Fish and Wildlife sets annual fishing quotas for all areas and sectors of the Herring commercial fishery. These quotas are set in accordance with the management strategy described in Chapter 7 of the Herring FMP, which includes the assignment of each area to one of three management tiers. The Herring FMP's tiered management approach serves to scale management effort appropriately to the size of the fishery in a given management area.

The Herring FMP's management strategy also includes the use of a Harvest Control Rule (HCR) for quota adjustment in the San Francisco Bay Herring gill-net sector. This HCR determines gill-net quota for the upcoming season based on a Spawning Stock Biomass (SSB) estimate and includes a 15,000-ton SSB cutoff, below which quota is reduced to zero tons. During the 2019-20 spawning season, the SSB in San Francisco Bay was 6,427 tons, resulting in a zero-ton quota for the 2020-21 season.

As prescribed by the FMP, the 2020-21 commercial season quotas for San Francisco Bay (including the Herring Eggs on Kelp, or HEOK, sector), Tomales Bay, Crescent City, and Humboldt Bay have been established at the levels shown the the table below. For all fishing areas, regulations allow the take of Herring for both sac-roe and fresh fish markets using gill nets. A current fish receiver's license and a Herring Buyer's Permit are required to buy Herring and HEOK for commercial purposes, Title 14 CCR 163.5.

Management Area and Fishery Sector	Management Tier Assignment	Quota (short tons)
San Francisco Herring Eggs on Kelp (HEOK)	NA	14
San Francisco Bay Herring (Gill Net)	2	0
Tomales Bay Herring (Gill Net)	1	133

Management Area and Fishery Sector	Management Tier Assignment	Quota (short tons)
Crescent City Harbor Herring (Gill Net)	1	11
Humboldt Bay Herring (Gill Net)	1	11

Sincerely,



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Director

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California Fish and Game Commission Coastal Fishing Communities Project Update on Efforts Related to Initial Staff Recommendations

November 4, 2020

Note: For purposes of this document, Commission refers to the California Fish and Game Commission, MRC refers to the Commission's Marine Resources Committee, and Department refers to the California Department of Fish and Wildlife.

Project Summary and Timeline

Under Commission direction, the MRC's Coastal Fishing Communities Project has been underway since 2015. Commission staff held a series of eight meetings with stakeholders between 2016 and 2018 in coastal communities across the state, which were designed to inform MRC on the issues facing coastal fishing communities.

Commission staff synthesized input from the community meetings into key themes and provided a staff report to MRC in July 2018. Following a public comment period and additional discussion with MRC in November 2018, the Commission approved an MRC recommendation for staff to incorporate stakeholder comments into a revised staff report.

The revised staff report was presented to MRC in November 2019 and, in December 2019, the Commission approved an MRC recommendation to adopt the [Staff Synthesis Report on California Coastal Fishing Communities Meetings](#) as final. The 2019 report included ten staff recommendations as "initial concepts for potential development," herein referred to as "staff recommendations" or by acronym as "SRs". MRC directed staff to further evaluate and prioritize the SRs on which the Commission may choose to act.

In preparing to conduct more in-depth evaluations of the SRs, staff recognized that efforts have continued which are directly relevant to the recommendations; in a couple of areas staff is actively engaging and continuing to progress, while in some areas staff is staying abreast of related activities led by other organizations. This report has been compiled to record efforts made by staff, collaborators, and/or through external sources that are relevant to each SR as a starting point for the analytical evaluation to inform potential Commission direction. A separate update is provided for each SR.

Overview of Update Reports for SRs

The update report for each SR includes the original text of the SR with a corresponding overview largely taken from the 2019 staff synthesis report, a summary of progress to date, potential next steps, and any linkages to other SRs.

The summaries of progress to date include efforts within California of which Commission staff is aware, some at the Commission's direction and some external to the Commission, as well as useful external models. Where applicable, each summary of progress is divided into four focal areas: *Commission and Department Initiatives; Collaborations and Partner Efforts; Relevant External Actions and Models; and Education and Outreach.*

- *Commission and Department Initiatives:* This focal area includes any work conducted directly by Commission staff or the Department relevant to the project.
- *Collaborations and Partner Efforts:* This focal area includes actions taken by other agencies and organizations within California relevant to this project in consultation or collaboration with Commission staff. For purposes of these documents, collaborations are defined as efforts in which Commission staff was directly involved as a participant or part of a steering committee, and partners are defined as organizations with whom Commission staff has engaged in dialogue about work that could inform or advance the project, if not with direct staff involvement.
- *Relevant External Actions and Models:* This focal area includes actions by organizations both within and outside of California that are relevant to the recommendation and could inform MRC understanding, but with whom Commission staff has not directly engaged.
- *Education and Outreach:* This focal area includes any materials released or posted by Commission staff or any targeted stakeholder contact or conversation conducted relevant to the recommendation.

The potential next steps sections include known, specific, targeted actions that could be taken by Commission staff, MRC and/or FGC to move the SR in question forward.

The linkages sections detail ties to other SRs. For example, linkages could be supporting functions, direct ties that necessitate the two moving forward together, or similar goals that MRC or the Commission might elect to advance at the same time.

Project Updates by SR

For quick reference and ease of navigation, the ten SRs are listed here with a link to each update report.

[**SR 1: *Develop and adopt a policy and definition for coastal fishing communities***](#)

[**SR 2: *Review the Commission's policy on restricted access commercial fisheries***](#)

[**SR 3: *Approve specific, small-scale projects to test and evaluate proposed new approaches***](#)

[**SR 4: *Engage legislative staff to pursue adjustments to laws as ideas are refined, if warranted to support fishing community adaptability***](#)

[**SR 5: *Direct staff to increase engagement and coordination with sister agencies, when feasible, on management decisions affecting California coastal communities***](#)

[**SR 6: *Explore pathways for authorizing community-based adaptable fishery structures \(e.g., community permit banks or risk pools\)***](#)

[**SR 7: *Explore filling data needs through collaborative research and data collection***](#)

SR 8: Survey communities, commercial and recreational fishers, and processors about their priorities for Commission focus

SR 9: Explore a model of “fishing community sustainability plans” (CSPs) and possible development of a state fisheries-based module to add to existing CSPs.

SR 10: Continue to develop an understanding of climate change impacts on fisheries and fishing communities

Coastal Fishing Communities Project

Update on Efforts Related to Staff Recommendation #1

November 4, 2020

Recommendation: Develop and adopt a policy and definition for coastal fishing communities.

Overview: Consider developing a new policy related to coastal fishing communities for Commission adoption. A policy could help clarify how the Commission wishes to consider coastal fishing community needs in decision-making, and the information necessary to help support those decisions. A policy could help flesh out the vision for the role Commission decision-making can play in preserving coastal fishing communities in California. Developing a draft policy is best accomplished in collaboration with stakeholders, tribes and tribal communities, academics, the Department, and other government agencies and jurisdictions that influence the sustainability of coastal fishing communities through their actions.

Progress by Focal Area

Commission and Department Initiatives

- **Definition:** Commission staff worked with stakeholders to develop a working definition of “coastal fishing community”, which was adopted by MRC in November 2019 for use in the Coastal Fishing Communities Project as reported to the Commission in December 2019. The working definition is:

“A coastal fishing community is defined as a social, cultural, economic, and/or place-based group whose members are fishermen dependent upon or engaged in commercial, recreational, or subsistence fishing to meet the social or economic needs of the community; this includes, but is not limited to, businesses and organizations that depend on or support fishing by providing goods and services, including infrastructure. A fishing community may be a subset or member of larger or associated coastal communities which have an interest in and/or are dependent on healthy ocean ecosystems.”
- **Policy:** Previous Sea Grant fellows to the Commission have examined legislative context and stakeholder comments submitted in response to the SR, and worked to organize concepts for potential inclusion in a draft policy.

Relevant External Actions and Models

- **California Ocean Protection Council (OPC):** Through its [2020-2025 strategic plan](#), OPC has signaled that “Supporting Ocean Health through a Sustainable Blue Economy” is a key strategic priority (Goal 4); the objective to “Advance Sustainable Seafood and Thriving Fishing Communities (Objective 4.1) identifies the Commission and Department among the partner agencies to help achieve the objective.

- *Federal:* The Pacific Fishery Management Council's Climate and Communities Initiative, which has similar goals to the Commission's Coastal Fishing Communities Project, is guided by the Magnuson-Stevens Fishery Conservation and Management Act and the act's definition of fishing community.

Education and Outreach

- **Website:** The final staff synthesis report approved in December 2019 has been posted to the Commission's website, along with information from each of the public meetings held between 2016 and 2018.

Potential Next Steps

Draft a potential policy to bring to a future MRC meeting and/or workshop for review. In drafting any potential policy, rely on previous stakeholder comments on the staff report, consider the efforts of previous Sea Grant fellows, and reach out to interested stakeholders, tribes, and governmental agencies to solicit additional input on priorities for the contents of a policy.

Linkages to Other Staff Recommendations

This recommendation is linked to others in that it could provide a foundation to support various types of initiatives, depending on the content of the policy. For example, a policy could support interagency outreach in support of communities (SR 5), community self-sufficiency (SR 6), collaborative work (SR 7), or continued outreach and collaboration with communities (SR 8).

Coastal Fishing Communities Project

Update on Efforts Related to Staff Recommendation #2

November 4, 2020

Recommendation: Review the Commission's Policy on Restricted Access Commercial Fisheries.

Overview: Restricted access programs and the Commission's Policy on Restricted Access Commercial Fisheries (Restricted Access Policy) were cited by many community members as contributing barriers to entry and adapting fishing strategies and targets as local changes arise, including those associated with climate dynamics. Other community members defended current restricted access programs as effective management that has improved the resources, the economic viability of fishing, or both. This complex policy includes 21 individual sub-policies across 9 unique topic areas. The Commission could conduct a review of the conditions that led to the policy being developed, how the policy has or hasn't been applied to specific fisheries since it was adopted in 1999, how the policy has performed at meeting fishery objectives and whether any objectives have changed, any unintended consequences that have affected fishing communities, and whether current conditions warrant possible adjustments to the policy.

Progress by Focal Area

Collaborations and Partner Efforts

- *Evaluating scope:* Partner non-governmental organizations, in consultation with Commission and Department staff, conducted an initial evaluation of what a review of the Restricted Access Policy might entail related to scope, approach, and process. Commission staff is prepared to compile a staff scoping report based on the initial evaluation.
- *Information-gathering:* In consultation with Commission and Department staff, initial background information related to select commercial fisheries with existing restricted access permit structures is being compiled. The background information could be used to evaluate fishery performance, in consultation with fishery participants, relative to the goals each restricted access program was intended to achieve.

Relevant External Actions and Models

- *Sea Grant:* The [California Sea Grant Strategic Plan 2018-21](#) includes provisions for evaluating impacts of current fisheries and potential new policies on fishing communities and marine species. Commission staff has previously worked with Sea Grant Extension staff on fisheries analysis and may wish to pursue this collaboration further for restricted access fisheries.
- *Federal, Alaska region:* The NOAA Fisheries [Restricted Access Management \(RAM\)](#) Program is specifically for administering the logistics of federal limited-entry fisheries in

the Alaska region. While such a program for California is outside of current resource capacity, it serves as a different model for how restricted access can be managed and monitored.

- *Alaska fisheries:* The Alaska [Commercial Fisheries Entry Commission](#) was highlighted in public meetings as another model for limiting permit issuance that may be informative.

Education and Outreach

- *Stakeholders:* Primary outreach to date related to the Restricted Access Policy occurred during public review of the staff synthesis report; comment letters received from stakeholders included specific comments on the policy and should be considered in any review.

Potential Next Steps

Develop and present to MRC a staff report analyzing the potential breadth of and process for a policy review, referencing the initial evaluation developed by partner non-governmental organizations.

Linkages to Other Staff Recommendations

Restricted access fisheries were cited as an issue because they can create limitations to access and lack of flexibility. Access and flexibility could be addressed by the experimental fisheries program, see SR 3, though the extent of this program does not address the full breadth of the Restricted Access Policy. SR 6, related to community-owned quota programs, would likely require adjustments to the policy.

Coastal Fishing Communities Project

Update on Efforts Related to Staff Recommendation #3

November 4, 2020

Recommendation: Approve specific, small-scale projects to test and evaluate proposed new approaches.

Overview: Stakeholders have requested that the Commission allow for stakeholders and partners to develop small-scale projects to test new approaches, including departures from the Commission's Policy on Restricted Access Commercial Fisheries and current permit structures, acknowledging that permit holders are key stakeholders in helping to create, design and define these projects, in consultation with the Department. Consider projects supporting opportunities for small-scale fishing that can be designed to help to fill information gaps consistent with guidance from the MLMA master plan for fisheries.

Progress by Focal Area

Commission and Department Initiatives

- *Experimental fishing permit regulations:* The Fisheries Innovation Act of 2018 went into effect on January 1, 2019, repealing an experimental gear program and authorizing a new experimental fishing permit (EFP) program to be established upon adoption of regulations by the Commission. Once adopted, the EFP program will provide a new pathway for testing pilot projects for commercial and recreational fishing. A partner non-governmental organization sponsored the legislation and continues to support implementation efforts. The Commission divided implementation into two phases:
 - EFP Phase I: This rulemaking was designed specifically to continue uninterrupted an experimental brown box crab fishery that commenced in 2019; regulations adopted by the Commission went into effect in May 2020.
 - EFP Phase II: A rulemaking to establish a comprehensive EFP program, consistent with the new statute, is currently underway. The Department presented a proposed structure at the March and July 2020 MRC meetings; based on MRC recommendations, the Commission approved scheduling a rulemaking for the EFP program with a timeline to be determined pending regulatory staff capacity. A Department team is currently drafting regulations for the program; there has been substantial stakeholder engagement and input throughout the development and drafting process.

Collaborations and Partner Efforts

- *Workshop:* The report from a [July 2019 guided learning workshop](#) held by the California Ocean Science Trust and attended by Commission and Department staff summarized concerns and potential management strategies to assist with coastal fishing community resilience under climate change. Suggestions included exploring permit flexibility

(including establishing a working group) and strategies to take advantage of emerging fisheries. A second series of workshops has been funded and is in nascent stages, the results of which may help identify new permitting strategies to test as pilot projects.

Relevant External Actions and Models

- *California Ocean Protection Council:* The council's recently-released [strategic plan for 2020-2025](#) includes a target to "implement pilot projects statewide to increase fishing communities' resiliency and adaptation to climate impacts by 2025" (Target 4.1.2, for Objective 4.1 Advance Sustainable Seafood and Thriving Fishing Communities).
- *Federal management:* [NOAA Fisheries' Western Regional Action Plan](#) outlines regional objectives, tools and resources for federal fisheries management on the West Coast. The plan contains provisions for evaluating alternative harvest management strategies.
- *Federal Exempted Fishing Permit Program:* NOAA Fisheries' exempted fishing permits are similar to state-level EFPs, the application for which is on a region-by-region basis.
- *Literature review:* A [2016 NOAA Fisheries technical memo](#) reviewed several types of fisheries management from the literature, which could be applied to assist with climate resilience. One of the approaches was allowing alternative gear types, and stakeholders have expressed interest in using EFPs to test alternative gear types in California.

Education and Outreach

- *EFP program development workshop:* Department and Commission staff co-lead a [stakeholder workshop](#) on January 14, 2020 to solicit thoughts from stakeholders concerning program design and priorities for an EFP program in California.

Potential Next Steps

Continue to support Department action as it develops the framework to the EFP program.

Linkages to Other Staff Recommendations

This recommendation is inherently collaborative and requires a research element, which makes it a vehicle for exploring collaborative research as referenced in SR 7. This recommendation is also the first example of pursuing legislation to increase commercial fishing adaptability, relevant to SR 4.

Coastal Fishing Communities Project

Update on Efforts Related to Staff Recommendation #4

November 4, 2020

Recommendation: Engage legislative staff to pursue adjustments to laws as ideas are refined, if warranted to support fishing communities.

Overview: Recognizing that some possible actions may be outside of Commission authority to accomplish, direct staff to seek to partner with stakeholders, the Department, and non-governmental organizations to find appropriate issues and means of engaging with legislators and legislative staff.

Progress by Focal Area

Commission and Department Initiatives

- *Experimental fishing permit (EFP) program:* The EFP program (Phase II currently in development; see SR 3 update) is being developed specifically with the intent of allowing flexibility and adaptability in coastal commercial and recreational fishing, consistent with the enabling statute.

Collaborations and Partner Efforts

- *Legislative engagement:* Commission staff have been and intend to continue to engage in Joint Committee on Fisheries and Aquaculture hearings as a way of identifying potential areas of mutual interest that legislation may help support/facilitate. Commission staff also confers regularly with the committee's chief consultant.

Relevant External Actions and Models

- *Legislation:* The [Fisheries Omnibus Bill](#) (Senate Bill 1309; Chapter 985, Statutes of 2018) contains specific provisions for increasing flexibility for California halibut trawl vessel permit transfers. While small, this is an example of increasing flexibility within a state-managed restricted access fishery, reflecting legislative responsiveness to industry needs supported by both the Department and Commission.

Potential Next Steps

As needs emerge, determine opportunities for adjusting legislation to facilitate adaptations to meet coastal fishing community needs; these may be identified through conferring with stakeholders, the Department, non-governmental organizations and legislative staff where supported by the Commission.

Linkages to Other Staff Recommendations

This recommendation is linked to SR 5 (*coordinate with sister agencies*), as both are ways in which the Commission could collaborate on issues that are outside the scope of its authority. SR 3 (*approve small-scale projects*) will be facilitated through an EFP program, a result of legislation directly addressing issues relevant to this project.

Coastal Fishing Communities Project

Update on Efforts Related to Staff Recommendation #5

November 4, 2020

Recommendation: Direct staff to increase engagement and coordination with sister agencies when feasible on management decisions affecting California.

Overview: Commission-related actions in isolation cannot meet all needs of coastal fishing communities, and decisions made by different coastal management authorities can have a combined influence on the health of a coastal community. Community members have requested deeper Commission engagement with coastal management agencies to urge them to consider in their decision-making potential impacts to California's coastal fishing communities. Sister agencies that fishing community members emphasized include the Pacific Fishery Management Council (PFMC) related to federal fisheries management decisions for the West Coast, and the California Coastal Commission, related to coastal development permit approvals to facilitate awareness and coordination on relevant topics and projects.

Progress by Focal Area

Commission and Department Initiatives

- *Master plan for fisheries:* The Commission is inherently collaborative with the Department, with a strong partnership and reliance upon one another's expertise. The Department developed and the Commission adopted the [2018 master plan for fisheries](#). Chapter 11 ("Adapting to Climate Change") focuses on how climate change may impact California's fisheries and management strategies for preparing to maintain resilient ecological and socioeconomic systems; this is a lens through which the two agencies can continue to build collaborative projects relevant to coastal fishing communities.

Collaborations and Partner Efforts

- *California Coastal Commission (CCC):* Stakeholders have cited the [California Coastal Act](#) (Public Resources Code, Section 30000 et seq.) to justify protecting and prioritizing harbor infrastructure in coastal planning and development decisions. The statute has specific provisions for maintaining commercial fishing infrastructure in ports and harbors. In 2017, the Commission directed staff to draft and send a letter to CCC in response to requests from fishing community stakeholders who participated in the coastal fishing communities meetings. The letter urged CCC to consider fishing community infrastructure and economic needs when considering coastal development projects. Further collaboration between CCC and the Commission will be necessary to enhance clarity on shared objectives.
- *California Ocean Protection Council (OPC):* Several documents released or supported by OPC call for inter-agency collaboration to meet OPC's goals of sustainable fisheries and climate change mitigation. Specific documents include OPC's [2020-2025 strategic plan](#), and a [2017 report](#), *Readying California's Fisheries for Climate Change*.

- *California Ocean Science Trust (OST):* A [report from a July 2019 guided learning workshop](#) summarizes concerns and potential management strategies to assist with coastal fishing community resilience under climate change. Emphasis was placed on the need to collaborate, specifically at a local level.
- *Collaborative fishing communities inter-agency workgroup:* Prompted by a Pacific Fishery Management Council (PFMC) workshop (see federal efforts in next section), a committee of staff from the Commission, the Department, OPC, and OST convened and conferred about how best to engage each agency's specialized knowledge in climate change topics and how to leverage and influence federal momentum such that it could meet the needs of state-managed as well as federally-managed fisheries. The committee has not met regularly this year and could be reconvened as this project moves forward to keep all involved agencies up to date on each other's work and identify areas for potential collaboration.

Relevant External Actions and Models

- *Federal fisheries management - "climate and communities" initiative:* PFMC is one of the key agencies with whom stakeholders are most interested in seeing the Commission collaborate. Most relevant to this project is the [PFMC Climate and Communities Initiative](#), an ecosystem-based fishery management initiative contained within the [PFMC Fishery Ecosystem Plan](#).
 - In 2018, Commission staff participated in a [Climate and Communities Initiative Workshop](#), hosted by The Nature Conservancy (TNC) on behalf of PFMC, as part of the scoping for the initiative. The results of the workshop have helped shape and focus PFMC's pursuit of the climate and communities initiative. Commission staff have continued to meet with Department representatives to the PFMC to find ways to harmonize PFMC's and the Commission's efforts.
 - Most recent activities of the initiative include a January 2020 workshop, [Developing Future Scenarios for Climate Change in the California Current Ecosystem](#), co-sponsored by TNC and PFMC, which detailed a set of climate scenarios and potential fishery impacts that could be used in conversation with fishermen in community resilience planning. Potential impacts will be discussed at an upcoming series of regional workshops with West Coast commercial fisheries community members.

Potential Next Steps

Continue to track relevant issues in which other agencies are engaged and communicate where appropriate; continue to coordinate with the California interagency workgroup to facilitate information exchange and help further the Commission's goals in adjacent efforts.

Linkages to Other Staff Recommendations

This recommendation is linked to SR 4 (*Engage legislative staff*), as both are ways in which the Commission could collaborate on issues that are outside the scope of its authority. Coordination with sister agencies has occurred as part of actions taken on almost all other recommendations (SR 1-4, 6-8).

Coastal Fishing Communities Project

Update on Efforts Related to Staff Recommendation #6

November 4, 2020

Recommendation: Explore pathways for possible community-based adaptable fishery structures (e.g., community permit banks or risk pools) to be authorized.

Overview: Explore options for community-organized structures that provide for adaptable responses within the community and could include co-management responsibilities. Consult with partner organizations and possibly convene an experts' workshop. This recommendation may require legislative or regulatory frameworks to accommodate such avenues. An example of such a structure that could be used as a model is the Monterey Fisheries Trust.

Progress by Focal Area

Commission and Department Initiatives

- *Master plan for fisheries:* The Commission-approved [2018 master plan for fisheries](#) focuses on adapting fisheries to climate change in Chapter 11, and includes flexible permitting as a method of maintaining socioeconomic resilience, providing it is consistent with the Commission's restricted access policy. The chapter commits that "...As resources permit, the Department and Commission will work with stakeholders to conduct an analysis of permit transferability in California fisheries and the Commission's policy on restricted access fisheries. This analysis will include how permits are retired and new permits are issued, and the potential for gear switching. Permitting considerations will also be included in the development of new FMPs [fishery management plans]. A working group comprised of stakeholders, outside experts, and Department and Commission staff could help to interpret analyses and develop policy recommendations."

Collaborations and Partner Efforts

- *Guidance:* A 2017 guidance document from the California Ocean Science Trust (OST), [Readying California Fisheries for Climate Change](#), provides a high-level overview of climate change impacts on state-managed fisheries; the guidance includes "adaptable permitting structures" among its recommended management approaches for preparing for climate change impacts.
- *Resilience workshop:* In 2019, OST hosted a guided learning workshop on climate change and coastal fishing communities. Planned through an inter-agency steering committee including staff from the Commission, the Department, OST, and the California Ocean Protection Council, the workshop was designed to explore policy and management options to support coastal fishing community resilience under climate change. The [workshop report](#) discussed pursuing community-owned fishing opportunities (e.g., community permit banks, permit leasing).

- *Upcoming:* A follow-up to the 2019 workshop is being pursued to explore with fishing industry representatives and fisheries managers the feasibility of adaptable management structures. OST and two economists from the University of California at Davis recently received a grant to examine the economic impact of potential flexible permitting programs in California, and whether such programs would serve to help communities weather climate change impacts to fisheries; several workshops are anticipated.

Relevant External Actions and Models

- *Case study:* The [Monterey Bay Fisheries Trust](#) has a quota leasing program which supports local commercial fishing in Monterey Bay. This program serves as a good example for how other community-based quota banks could function in California.
- *Literature review:* A [2016 NMFS Tech Memo](#) recommends region-specific management approaches, including community quota ownerships.
- *Out-of-state examples:* The North Pacific Fishery Management Council has several [community quota programs](#) for federally-managed fisheries that are intended to provide fishing villages with access and to support economic development. The programs include a community development quota program (CDQ), a community quota entity program (CQE), and some species-specific cooperatives, which could serve as learning tools for developing community quotas at the state level.

Potential Next Steps

Assign staff to determine what would be required to authorize community-based adaptable fishery structures in order to determine authority and approach. Query fishing communities in California on the desire and level of support for community-based adaptable fishery structures.

Linkages to Other Staff Recommendations

Pursuing and implementing this recommendation would likely require adjustments to the restricted access program, relevant to SR 2. The experimental fisheries permit program, discussed under SR 3, could be leveraged along with this recommendation; for example, to test community structures, a single large permit (e.g., squid seine) might be allocated to several fishermen in a community by suspending the current one-vessel requirement.

Coastal Fishing Communities Project

Update on Efforts Related to Staff Recommendation #7

November 4, 2020

Recommendation: Explore filling data needs through collaborative research and data collection.

Overview: Coastal fishing community members have raised a concern that adaptive responses and new management strategies have not been pursued due to lack of data. Many fishermen have offered to support collaborative data gathering. The Commission could work with the Department on identifying data gaps and possible scientific information that could be gathered through collaborative research or experimental fishing between partner entities and fishermen. Such efforts might be coordinated through an app or a website. However, great care must be taken to create citizen science data collection systems that provide credible data. The Commission would have to rely on partners for labor costs.

Progress by Focal Area

Commission and Department Initiatives

- *Purple sea urchin:* In February 2020, the Commission adopted emergency regulations to authorize unlimited removal of purple sea urchin at Caspar Cove, Mendocino, by recreational divers. The Department is cooperating with divers and monitoring the cove to examine the impact of sea urchin removal on the ecosystem. The Commission is considering proposed regulations to continue the project for three years, and to add an additional project at Tanker Reef, Monterey. This provides an excellent model of research collaboration between recreational divers and the Department.
- *Experimental fisheries permit (EFP) program:* The EFP program currently being developed (see SR 3) will require participants to provide a research component for their experimental permits and collaborate with the Department to provide data.
- *Red abalone:* A fishery management plan for the recreational red abalone fishery is under development. At its July 2020 meeting, the MRC recommended, and the Commission subsequently adopted, a biological fishery option and a harvest control rule that includes a stakeholder data collection effort; This effort will simultaneously support fishery participation and provide fishery-dependent data on abalone populations. Interests in a similar program have been expressed by former commercial abalone divers in southern California.
- *Master plan for fisheries:* Chapter 11 of the [2018 master plan for fisheries](#) contemplates using stakeholder data in managing fisheries affected by climate change. Given that the Department is the Commission's main collaborator, continuing to pursue projects such as the three outlined above in concert would be appropriate and necessary.

Relevant External Actions and Models

- *Groundfish*: Within California, one well-established cooperative research effort is the California Collaborative Fisheries Research Program, which gathers hook-and-line data on federally-managed groundfish species in California's marine protected areas (MPAs). This program has received strong support and buy-in from the commercial passenger fishing vessel and recreational hook-and-line communities and could serve as a model for collaborative data-gathering.
- *Guidance*: A [2017 report](#) guidance document from the California Ocean Science Trust provided a high-level overview of climate change impacts on state-managed fisheries, with recommended management approaches on preparing for climate change impacts. Management recommendations include increasing collaboration within agencies and with fishermen.
- *Private/academic sector*: Humboldt State University researchers, Ecotrust, and Strategic Earth Consulting are currently [conducting a study](#) to assess port community well-being and socioeconomic conditions, which will contribute to long-term monitoring efforts to evaluate California's MPA performance. The study will include collecting qualitative data from fishing community leaders.

Education and Outreach

- *Workshop*: A [stakeholder workshop](#) on the proposed EFP program, co-lead by staff from the Commission, the Department, and The Nature Conservancy, was held on January 14, 2020. The workshop solicited thoughts from stakeholders concerning program design and their priorities for the program. This workshop is relevant as EFPs will require a collaborative research component.

Potential Next Steps

Commission staff could consult with the Department regarding outreach to key stakeholders for projects where data needs are not being met.

Linkages to Other Staff Recommendations

This recommendation is tied to SR 3, as the current approach to approving small-scale test projects (upon Commission approval of the EFP program) requires a research component.

Coastal Fishing Communities Project

Update on Efforts Related to Staff Recommendation #8

November 4, 2020

Recommendation: Survey communities, commercial and recreational fishers, and processors about their priorities for Commission focus.

Overview: This strategy could help refine understanding about the issues facing coastal fishing communities and their priorities. Some stakeholders have criticized this recommendation as being too similar to this Coastal Fishing Communities Project.

Progress by Focal Area

Collaborations and Partner Efforts

- *California Ocean Science Trust (OST):* OST facilitated a [July 2019 guided learning workshop](#) on climate change and coastal fishing communities to help inform the Commission's Coastal Fishing Communities Project. The steering committee for the workshop included staff from the Commission, Department, and California Ocean Protection Council (OPC), who also attended and participated in the workshop. The workshop report summarizes concerns and potential management strategies to assist with coastal fishing community resilience under climate change. Emphasis was placed on the need to collaborate with local port stakeholders—which would include commercial and recreational anglers and processors—and identified potential future steps, to include exploring adaptable fishing structures.

Relevant External Actions and Models

- *California Natural Resources Agency (CNRA):* CNRA's [2018 safeguarding California report](#) contains several next steps relevant to state-managed fisheries, including outreach to marine resource users.
- *Federal initiatives:* In January 2020 the Pacific Fishery Management Council (PFMC) held a workshop as a part of its [Climate and Communities](#) initiative. Developed in conjunction with The Nature Conservancy, the workshop detailed a set of climate scenarios and potential fishery impacts; the potential fishery impacts will be discussed at an upcoming series of regional workshops with West Coast commercial fisheries community members.
- *Private/academic sector:* Humboldt State University (HSU) researchers, Ecotrust, and Strategic Earth Consulting are currently [conducting a study](#) to assess port community well-being and socioeconomic conditions. The study will include outreach and surveying to collect qualitative data from fishing community leaders.

Education and Outreach

- *Experimental fishing permits*: One of the top priorities of community members identified in the 2016-2018 survey effort was a desire to pursue emerging fisheries and permit flexibility. The proposed experimental fishing permit (EFP) program currently under development provides an avenue to address those concerns. A [stakeholder workshop](#) was held in January 2020 and co-lead by Commission and Department staff and The Nature Conservancy to solicit thoughts from stakeholders concerning EFP program design and priorities. The EFP program is an example of what can result from being made aware of priority concerns for community members.
- *Apprenticeships*: Another top priority during the 2016-2018 surveying effort was addressing the “aging of the fleet” and the difficulty of entry for young participants. California Sea Grant has established a [commercial fishing apprenticeship program](#) consisting of workshops and on-the-job training designed to address the concern; the most recent workshop was in January 2020. This program is another example of what could be modeled to address priority concerns.

Potential Next Steps

Due to changes in fish businesses and markets associated with the COVID-19 pandemic, members of fishing communities and fisheries may have new perspectives to share relative to the surveying conducted by Commission staff from 2016 to 2018. Outreach to key stakeholders could be used to determine the best strategy for broader outreach to fishing communities and to gather ideas for survey questions if a survey is determined to be an effective and desired tool to use.

Linkages to Other Staff Recommendations

The most recent work surveying fishing communities was during the 2016-2018 meetings. Some of the primary concerns and priorities highlighted by fishermen during those meetings have been accounted for and addressed through other recommendations, notably concerns about the limitations of restricted access (SR 2), the ability to respond to emerging fisheries (EFP, SR 3), and adaptable permitting structures (SR 6). This recommendation is also tied to SR 10, as surveying communities would assist in building a better understanding of their needs in response to climate change.

Coastal Fishing Communities Project

Update on Efforts Related to Staff Recommendation #9

November 4, 2020

Recommendation: Explore a model of “fishing community sustainability plans” (CSPs) and possible development of a state fisheries-based module to add to existing CSPs.

Overview¹: CSPs are planning documents that require a coastal community to assess needs and provide concrete recommendations for improving its industry, its waterfront, and the sustainability of its ports. CSPs are cited in the [Magnuson-Stevens Fishery Conservation and Management Act](#) as a potential method to avoid negative impacts to small fishing communities from the catch share program by enabling communities to be eligible to purchase catch shares and participate in limited access fisheries when approved by regional fishery management councils. More broadly, CSPs enable communities to plan strategically and be more proactive in developing fishing community resilience for a sustainable future. Incorporating a state fisheries module could potentially be part of a future where ports are empowered to define how to support their own fishing community resilience and structure fisheries access according to their unique needs, and potentially become eligible to participate in state restricted access fisheries as communities.

Progress by Focal Area

Relevant External Actions and Models

- *Current CSPs in California:* As of June 2020, [Noyo Harbor](#), [Monterey](#), and [Morro Bay](#) have developed and adopted CSPs. In addition, a research group at Humboldt State University (HSU) worked with fishing community leaders in [Eureka](#) and [Shelter Cove](#) to develop CSPs for those communities; while these CSPs have not been approved by the Pacific Fishery Management Council (PFMC), they are being utilized actively in their ports to implement top priorities and [provide a roadmap for CSP planning](#).
- *Other community plans:* The California Natural Resources Agency’s [Safeguarding California Plan 2018](#) includes several next steps relevant to state-managed fisheries, including providing grants for community-based vulnerability assessments, specifically for marine fisheries and related socioeconomic groups. While not the same as a CSP, vulnerability assessments can help guide communities in choosing priorities for state CSPs.
- *Fishing community profiles:* PFMC is developing updated [profiles](#) of West Coast fishing communities which include economic, demographic, and fishery information. The profiles could be used as tools to define community needs and priorities for community sustainability planning.

¹ This overview has been updated since the completion of *2019 Staff Synthesis Report on California Coastal Fishing Communities Meetings, 2016-2018* to clarify the provisions for CSPs in federal fisheries under the Magnuson-Stevens Fishery Conservation and Management Act, and the potential new provisions for CSPs in state-managed fisheries.

- *Out of state:* Marine fisheries in Maine fall under the Department of Marine Resources (DMR). DMR has [several grant programs](#) to support its coastal communities, including the Shore and Harbor Planning Grants Program and the Coastal Communities Grants Program. Maine's programs could be a model for making funding available for communities to conduct assessment and planning activities.
- *Federal tools:* The [U.S. Climate Resilience Toolkit](#), a federal inter-agency initiative and part of the federal Climate Action Plan, provides tools and expertise for decision-makers to assess risk and build climate resilience. The toolkit includes a [Fisheries and Coastal Communities](#) page, which provides an overview of potential climate impacts to coastal fishing communities in the US, links to case studies of fishing communities adapting, and an example strategy for communities responding to change. These tools can help local communities define needs and strategies for sustainability planning.

Education and Outreach

- *Comments:* Directed conversations between the Commission and stakeholders resulted in comments on the 2019 staff synthesis report with specific requests to develop CSPs. Specifically, a letter from the Alliance of Communities for Sustainable Fisheries suggested developing a statewide CSP as part of a Commission fishing communities policy.

Potential Next Steps

Reach out to the organizations responsible for CSPs underway or already established in California. Presumably, experienced communities can provide valuable information on how communities begin to develop plans, identify any other interested communities of which they are aware, and provide insight regarding what, if any, updates to CSPs may be warranted specific to state-managed fisheries.

Linkages to Other Staff Recommendations

Like the permit banking options discussed in SR 6, this recommendation would require community-organized structures, and therefore requires community or local government initiatives. This recommendation would require the Commission to collaborate with local governments, relevant to SR 5.

Coastal Fishing Communities Project

Update on Efforts Related to Staff Recommendation #10

November 4, 2020

Recommendation: Continue to develop an understanding of climate change impacts on coastal fisheries and coastal fishing communities.

Overview: Science is still evolving regarding how fish populations and fisheries are affected by and respond to changing climate dynamics, including short-term, extreme ocean events. Developing successful fisheries management response strategies that meet both biological and socioeconomic needs is still nascent. Increased understanding of what is often referred to as “climate-responsive fisheries management” or adaptable management structures is needed.

Progress by Focal Area

Commission and Department Initiatives

- *Science Institute:* The Department’s Science Institute has [ongoing research and education projects relevant to climate change and management](#), including the [Climate College initiative](#), which could serve as resources to MRC and the Commission. The [2014 Climate College](#) focused on marine resources, and therefore may be particularly relevant.
- *Master plan:* Chapter 11 of the 2018 [master plan for fisheries](#) recommends conducting community vulnerability assessments to improve management of fisheries vulnerable to climate change, and would require building a community-specific understanding of climate impacts.
- *Urchins:* The Commission and Department have begun to undertake projects relevant to climate effects on marine resources, which will require information gathering. An example is the recent proposals for sea urchin removal as a management strategy for combating loss of kelp, a biogenic habitat; kelp loss has resulted from climate-driven ocean conditions exacerbated by abundant sea urchin populations.

Collaborations and Partner Efforts

- *Climate effects assessment:* Section 3.2 of [California’s Coast and Ocean Summary Report](#), from California’s Fourth Climate Change Assessment conducted in 2018, contains an in-depth overview of the known science surrounding fisheries and climate change in the state. The report includes summaries of scientific literature and case studies on specific effects of climate change that have already begun to affect California’s fisheries and fishing communities. Commission staff served on the advisory group for this effort.

Relevant External Actions and Models

- *California Ocean Science Trust (OST):* OST’s 2017 guidance document, [Readying California’s Fisheries for Climate Change](#), provides a high-level overview of climate

change impacts on state-managed fisheries, with recommended management approaches to guide preparations for managing fisheries impacted by climate change. The recommendations include many provisions for increasing understanding of climate change impacts.

- *California Sea Grant*: The goals of California Sea Grant's [2018-2021 strategic plan](#) include supporting ongoing research on fisheries relevant to climate and communities. Commission staff have previously worked with Sea Grant's extension staff on other parts of this fisheries research and the Commission may wish to pursue this collaboration further.
- *California Ocean Protection Council (OPC)*: The recently-released [OPC strategic plan for 2020-2025](#) includes objectives and targets relevant to building a better understanding of climate change, including research funding, reflecting alignment between goals of OPC and this Commission staff recommendation. For example, Objective 1.3 is to "Improve Understanding of Climate Impacts on California's Coast and Ocean", with an underlying target, 1.3.1, to "Identify and continue to fund and house needed climate-related data collection, research, and dissemination...".
- *Grants for vulnerability assessments*: The California Natural Resources Agency's [Safeguarding California Plan 2018](#) includes grants for community-based climate vulnerability assessments (CVAs) and, specifically, CVAs on marine fisheries and related socioeconomic groups, and integration of assessments into management and outreach to marine resource users.
- *Federal assessments*: The National Marine Fisheries Service (NMFS) is conducting [CVAs](#) on a species, system, and community basis. A Pacific Salmon CVA has been completed; a West Coast Fish Stock CVA and a Pacific Marine Mammal CVA are also in progress. These CVAs will build our understanding of how specific fisheries upon which communities depend may be impacted. A fishing community CVA for the eastern and gulf coasts of the U.S. provides a model for how fishing community social vulnerability can be assessed.
- *Federal tools*: PFMC held a January 2020 workshop in conjunction with The Nature Conservancy as a part of the [Climate and Communities Initiative](#), which detailed a set of climate scenarios and potential fishery impacts that could be used in conversations with fishermen in community resilience planning. These scenarios could be used in building our understanding of impacts.
- *Climate-specific tools*: The U.S. Climate Resilience Toolkit is an inter-agency initiative designed to help people find tools and expertise for building climate resilience. The [Fisheries and Coastal Communities page](#) provides an overview of potential climate impacts to coastal fishing communities in the U.S., links to case studies of fishing communities adapting, and an example strategy for communities responding to change. These tools can help local communities define needs and strategies for adaptive management and could help build our understanding of future impacts.
- *International symposium*: The 4th International Symposium on the Effects of Climate Change on the World's Oceans, held in 2018, included an exploration of climate-

induced changes impacting fisheries across the U.S. and internationally. Select sessions focused on approaches to [climate resiliency in fisheries management](#), which can be found in [Volume 76, issue five of the ICES journal of marine sciences](#). This wealth of global expertise could serve to inform local understanding of climate change impacts.

Potential Next Steps

Reach out to the research community to build a better understanding of available and ongoing research on climate change and marine fishing communities, and reach out to fishing community members to learn from their observations. As more is understood about climate change impacts on ocean ecosystems and fisheries in particular, adapt the strategies for the other SRs.

Linkages to Other Staff Recommendations

This recommendation could potentially tie into all other recommendations (SRs 1-9), as increasing our understanding of climate change impacts would inform priorities for implementing other recommendations.

California Fish and Game Commission Coastal Fishing Communities Project

DRAFT Analysis of Staff Recommendation 1 to “Develop and adopt a policy and definition for coastal fishing communities”

November 4, 2020 Draft

Background

In July 2020, staff presented the California Fish and Game Commission (Commission) Marine Resources Committee (MRC) with a draft standardized approach for evaluating and analyzing the staff recommendations contained in [2019 Staff Synthesis Report on Coastal Fishing Communities](#). The approach was presented as a potential structure for staff information-gathering and analysis of each recommendation to help inform MRC consideration of options for potential future action to recommend to the Commission. The approach presented to MRC forms the basis for this staff recommendation analysis, which focuses around four categories: **Basic informational needs, current regulatory and policy context, potential Commission role, and costs and benefits.**

An overview of SR 1, as reflected in the 2019 report, is:

Consider developing a new policy related to coastal fishing communities for Commission adoption. A policy could help clarify how the Commission wishes to consider coastal fishing community needs in decision-making, and the information necessary to help support those decisions. A policy could help flesh out the vision for the role Commission decision-making can play in preserving coastal fishing communities in California. Developing a draft policy is best accomplished in collaboration with stakeholders, tribes and tribal communities, academics, the Department, and other government agencies and jurisdictions that influence the sustainability of coastal fishing communities through their actions.

Analysis

I. **Basic Informational Needs**

Developing a definition of coastal fishing community and a Commission policy would be a near-term effort, relying on qualitative information derived from stakeholder input, existing laws and regulations, and relevant Commission direction as reflected in adopted management documents.

Information at hand: Input by stakeholders and fishing organization representatives on a potential policy was included in original comment letters on the 2019 staff report; the comments have been synthesized as an appendix to the report (table will be attached in final report). Needs for a policy, as identified by stakeholders, include but are not limited to recognizing loss of infrastructure, addressing access issues, and enumerating the pathways between biological and economic sustainability. Commission staff would ideally have more recent input on what stakeholders are interested in including in a policy, which would require additional stakeholder engagement.

Additionally, there are other models that could serve to assist in forming a policy. The Pacific Fishery Management Council (PFMC) Climate and Communities Initiative is an ecosystem-based fishery management initiative based on the Magnuson-Stevens Fishery Conservation

and Management Act's definition and guidance for fishing communities; the initiative has similar goals to the Commission's coastal fishing communities project and is currently active in California. Commission staff, along with staff from the California Department of Fish and Wildlife, the California Ocean Protection Council and the California Ocean Science Trust, as well as California fishermen and other partners, participated in a 2018 workshop as a part of the scoping for the initiative, and Commission staff have continued to confer with PFMC representatives to find ways to harmonize efforts. The most recent activities of the PFMC initiative include a January 2020 workshop in which participants developed a series of climate scenarios and potential fishery impacts. The potential impacts will be discussed at an upcoming series of regional workshops with members of commercial fisheries communities in Washington, Oregon, and California.

Information deficits: Relevant statutes (i.e., California Fish and Game Code, California Public Resources Code), including legislative findings, policies and mandates, need to be identified and compiled; sections of the MLMA and MLPA master plans need to be reviewed and evaluated for potential gaps in guidance pertaining to coastal fishing community and socioeconomic considerations when developing management actions.

Potential impacts: Pursuing this recommendation could provide a foundation to support development of other recommendations and various types of initiatives, depending on the content of the policy. For example, a policy could support interagency outreach in support of communities (recommendation 5), community self-sufficiency (recommendation 6), collaborative work (recommendation 7), or continued outreach and collaboration with communities (recommendation 8). Because this policy has the potential to impact any or all of the subsequent recommendations, staff recommends that initial efforts be taken relevant to this item while the prioritization and scoping process for other items is still underway.

II. Current Regulatory and Policy Context

MRC actions and context: In November 2019, MRC adopted a working definition of "coastal fishing community" for use in the Coastal Fishing Communities Project and recommended adoption to the Commission in December 2019. Adopting a definition is the first part of this recommendation. A policy has not yet been developed, though there have been some internal drafting discussions among Commission staff.

Commission actions and context: Prior relevant policies, including the restricted access policy and the formerly held commercial fishing and packing policy, could be used as references.

The commercial fishing and packing industries policy, adopted around 1993 and repealed in 2006, encouraged "...the development and expansion in all lawful ways of commercial fishing...consistent with the State's policy to provide for aesthetic, educational, scientific, and recreational uses of California's fisheries resources; the necessity of regulating the catch to sustain long term yields, and the development of distant water and overseas fisheries enterprises." Fostering and encouraging commercial fishing so explicitly, through the contemporary lens of the Marine Life Management Act, would be welcome to many stakeholders.

The restricted access policy is considerably longer than the now defunct commercial fishing and packing industries, but also includes language outlining that "...California's fisheries are a public trust resource. As such, they are to be protected, conserved, and managed for the

public benefit, which may include food production, commerce and trade, subsistence, cultural values, recreational opportunities, maintenance of viable ecosystems, and scientific research.” Stakeholders may rightfully argue that the “public benefit” is most directly relevant to them, as members of a community that relies on how a fishery is managed and, therefore, this ideal should be more explicitly shifted to their community needs. Any fishing communities policy will most likely affect and be affected by this restricted access policy, as well as non-restricted access fisheries, and outreach to the communities for any drafting process should include representatives from a diverse group of fisheries.

In addition to the Commission’s words, there is also the matter of its actions. In 2017, the Commission directed staff to draft and send a letter to the California Coastal Commission in response to requests from fishing community stakeholders who attended the 2016-2018 public coastal fishing communities meetings. The letter urged the California Coastal Commission to consider fishing community infrastructure and economic needs when approving coastal development projects. While such a request does not constitute regulation or policy, it is a prior act that implies values about coastal fishing communities which could be relevant to a new policy. The letter included language about the Commission’s desire to “strengthen the shared commitment of our partner coastal management agencies to maximize support for California’s coastal fishing communities” and to “preserve and balance California’s maritime heritage and economy and its coastal and ocean environments”, both of which are statements which could be tenants of a policy.

Statutory context: There is policy embedded in sections of the California Fish and Game Code and the California Public Resources Code that outlines, to a varying extent, the current stance of the institution towards fishing communities, though there is not a specific and explicit policy. Portions on conservation of aquatic resources, offshore fisheries that have become newly accessible, and assorted parts of the Marine Life Management Act include language about fishing community members or fishery participants. While it would take considerable text to describe all the relevant language from policy, there is support for growth of commercial fishery, protection for fishing infrastructure in ports, development of aquaculture, recognition of the importance of fisheries to economy and culture, and a desire to involve fishing community members in research and management concerning fishery resources. A compilation of select relevant parts of code relevant to fishing communities will be attached to the final report.

Further state context: The California Ocean Protection Council 2020-2025 strategic plan includes large, overarching goals and objectives that are relevant to coastal fishing communities. For example, goal one is to “safeguard coastal and marine ecosystems and communities in the face of climate change,” which inherently includes coastal fishing communities. Furthermore, objectives under goal four (“support ocean health through a sustainable blue economy”) include targets specifically brightlining coastal fishing communities, such as objective 4.1, to “advance sustainable seafood and thriving fishing communities.” While not statutory or regulatory language, the unique position of the California Ocean Protection Council means that its strategic plan indicates a political will and articulates a policy of the state, at least in line with the current administration. Therefore, the strategic plan is an important touchstone for issues currently considered administration priorities. The broader scope of the California Ocean Protection Council, especially its role as an interagency coordinator, may allow it to act with greater speed and breadth of role than the Commission in reference to coastal fishing communities.

Federal context: The Magnuson-Stevens Fishery Conservation and Management Act (Magnuson-Stevens Act) includes a federal definition of coastal fishing community, and includes a series of [national standards](#) for fishery management. National Standard eight defines the federal approach to fishery management relevant to the needs of fishing communities. The Magnuson-Stevens Act definition of a fishing community is “a community that is substantially dependent on or substantially engaged in the harvest or processing of fishery resources to meet social and economic needs, and includes fishing vessel owners, operators, and crew, and fish processors that are based in such communities. A fishing community is a social or economic group whose members reside in a specific location and share a common dependency on commercial, recreational, or subsistence fishing or on directly related fisheries-dependent services and industries (for example, boatyards, ice suppliers, tackle shops).” The standard states that any conservation and management measures must “take into account the importance of fishery resources to fishing communities by utilizing economic and social data that are based upon the best scientific information available in order to: (1) Provide for the sustained participation of such communities; and (2) To the extent practicable, minimize adverse economic impacts on such communities”, recognizing the social and economic importance of fisheries to communities affected by management measures.

III. Potential Commission Role

The Commission has authority to adopt policies to guide its actions and to guide the actions and set the expectations for what the Department brings to the Commission. Therefore, the Commission can take action on this item directly within its own authorities. However, this policy may be constrained by the fact that the Commission has only partial jurisdiction over commercial fishing in California, as some restricted access programs fall under the authority of the Department or the California State Legislature. A policy will necessarily have to be coordinated with the Department as it may affect the work of the Department.

IV. Costs and Benefits

Adaptability

Depending on the specific language of the policy, providing for adaptation could be built into the policy. If the Commission chose to include support for the principle of adaptive solutions for communities, that would provide more flexibility for those communities to pursue adaptable solutions, as well as strengthening their position to propose new approaches and providing support to leverage for pursuing other Staff Recommendations. However, this would require coordination with the Department and with stakeholders to ensure that any language on adaptability is feasible in terms of implementation and enforcement and usefulness for the fishing community.

Consistency

- How might this policy lead to potential changes to stable fishery management structures, such as impacts to an existing restricted access commercial fishery program? It may define new policy priorities and objectives that would lead to a review of existing management structures and programs to assess if the structures need to be adjusted in any way in response to the priorities/objectives within the new policy.
- Does this align with or possibly reflect a change to other existing Commission policies? It has the potential of placing greater emphasis on understanding the implications of

proposed management actions to not just the stock and the fishery, but also toward socioeconomic impacts at a finer scale – at the community and/or port level.

Accessibility

- Does this increase accessibility of a given fishery, and at what level (e.g., individual fishermen, new entrants, fishery-level, community- or geographic-level)? Has the potential to express policy for providing access at the levels described herein.
- How might this affect the species or fishing community involved? Does it increase engagement of fishing communities, in a manner that does not affect the sustainability of species harvested?

Manageability

A new policy would require the Department to consider the new policy when reviewing projects and developing recommendations, and take the lens of coastal fishing communities, which might increase the time and effort required for a review. Fishery information is not currently collected at the smaller scale that a fishing community or groups of fishing communities might necessitate. It could create a data gap that the Department would need to evaluate how to fill regarding collecting and reporting information at the relevant scale.

It would also be important to involve partners in drafting this policy, as any who do not feel represented in the process may take issue with the work of the Commission and Department related to the policy. Having a policy in place would demonstrate to commercial fishing communities that their current and future needs, and very value to preserve for the future, are recognized by the Commission, which might generate greater investment and engagement by communities to assist with management.

Affordability/Investment

Commission staff would be the lead on this effort, including the marine advisor, the Sea Grant fellow, and potentially the executive director or deputy executive director. The marine advisor and Sea Grant fellow would likely be responsible for initial drafting and for coordinating efforts to involve partners and stakeholders. The executive director and deputy executive director would be responsible for review and approval of materials and it would require a considerable investment from the marine advisor and sea grant fellow. Developing a policy would require multiple meetings with Department staff and partners, and one or two public workshops. The budget required primarily would be staff pay for time invested, for both FGC staff and Department staff.

Resilience

While a policy may not have a direct impact on resilience, it would require the Commission and the Department to give consideration to fishing community needs on project approvals, which may create more space for adaptive and economically beneficial programs to move forward in the fishing community. This additional consideration has the potential to improve both economic prospects and economic and ecological resilience in a broad, general way. However broad, evidence of general support would be useful for commercial fishing communities.

Conclusions

[To be developed]

From: Konstantin Karpov [REDACTED]

Sent: Friday, October 9, 2020 3:57 PM

To: FGC <FGC@fgc.ca.gov>

Cc: Peter Haaker <[REDACTED]>

Subject: Re: Oral comments regarding abalone closure extension on OCT 14 meeting - Susan Ashcroft

Hello Susan both Peter Haaker and I Plan on presenting oral comments totaling 6 minutes. We will plan to speak in sequence. Attached are three figures we wish to show during our presentations. Also we will be sending a longer letter to both Commissioners on Monday and hope they have a chance to read their before the meeting. See you then.

Yours Konstantin Karpov

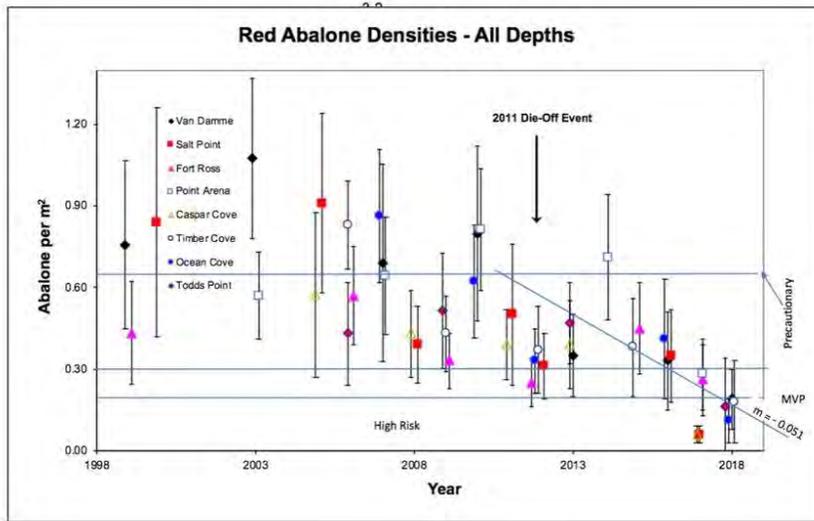


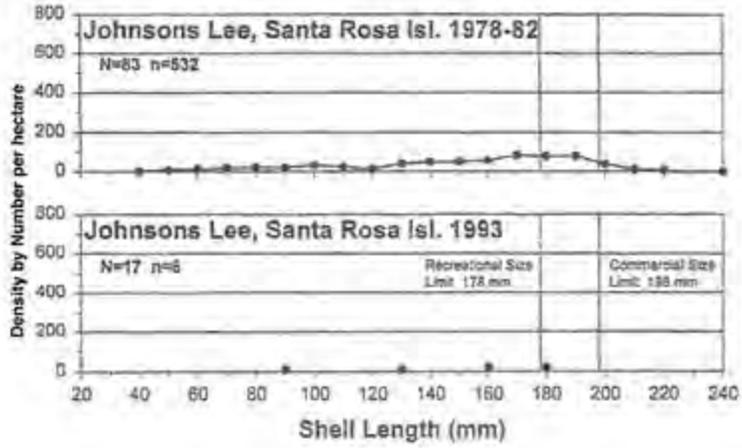
Figure 1.
CDFW abalone decline by density at northern California index sites. Slope of decline is average of all sites sampled, 2010-2018

$M = \text{slope of decline}$

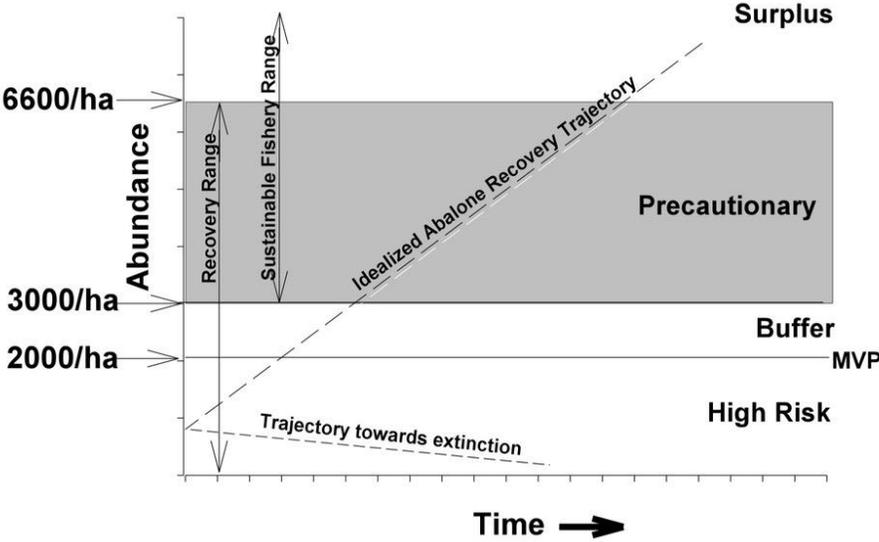


NOTE: 5 years to extirpation

Figure 2. Density by size of emergent red abalone, Johnsons Lee, Santa Rosa Island, 1978-1982 and in 1993. (Karpov et al., 1998) N = number of transects, n = number of abalone.



ARMP Approach



To: Commissioner Samantha Murray and Commissioner Peter Silva, Marine Resources Committee of the California Fish and Game Commission

From: Peter L Haaker

Re: Red abalone fishery closure extension and Fishery Management Plan (FMP)

Date: October 13, 2020

Dear Commissioners Murray and Silva,

I am writing about the Abalone Fishery Management Plan that's now under development, with some important background that I think you need to understand.

I worked for the Department for 37 years with a focus on California abalone for about half that time. I conducted monitoring and research and authored many papers and reports about abalone, including several of the species that eventually declined to endangered status. I was also a senior author of the Abalone Recovery and Management Plan. My group initiated and finally accomplished the listing white abalone on the Federal Endangered Species List.

When we proposed closure of the commercial and recreational abalone fisheries in central and southern California to the Fish and Game Commission in the 1990s, it was because the stocks had been reduced to levels below what was necessary to support a sustainable fishery.

After several years of research, we found that one species, white abalone, was in a more precarious biological situation. The population had dropped so low, there was a high probability that this species was in danger of extinction. The population of emergent and "legal" size animals was vastly reduced, and there was no evidence of any extant early life stages.

Populations of the other abalones, green, pink, black, and red, in central and southern were certainly depressed, but we were not worried about the continued existence of these species because there was still a viable crevice population, with many individuals surviving in the interstices of the rocky reef habitat. For instance, along the central California coast there was a large abalone fishery, which was eliminated by the recovery of the sea otter in the 1970s. Though there were not sufficient stocks of red abalone to have a fishery, many areas still supported crevice dwelling red abalone, which had continued to persist.

Indeed, this may have been the "basic" condition of red and other abalone populations throughout California for hundreds of thousands of years. When sea otters were hunted to near extinction in the mid-1850s, the primary predator of nearshore invertebrates was removed along the coast. This allowed the normally secretive, crevice dwelling abalone to grow and exist in higher numbers and larger sizes in more exposed areas. Thus, from about 1850 to the early 1900s, abalone stocks grew in size and number to eventually support a massive fishery that lasted for many decades.

In contrast, evidence we found in the early 1990s that white abalone was at high risk to become extinct eventually led to its placement on the Endangered Species Act in 2001. Subsequently, black abalone, which had been hit by an epidemic disease, was also listed as a federally endangered species.

With this background, I want to express my concern for the current condition of red abalone on the northern California coast. Like the green and pink abalone in southern California, I've long considered that this resource has been backed up and supported by a crevice population. However, recent events have suggested that this idea may no longer be operative. I am referring to recent deterioration of ocean conditions that adversely affect every life stage of abalones including crevice and emergent individuals. My concern includes all the abalones, but particularly the red abalone on the north coast.

It's important to underscore that many of the conditions that affect red abalone are in one way or another result from unusual elevated sea water temperature.

Red abalone is a "cold" water species. Its range is primarily in the cooler waters along the California coast north of Point Conception. It also occurs in the western most northern Channel Islands. The areas where red abalone are found are under the influence of the southerly moving cool California Current. Some limited populations of red abalone are also known to exist in areas of cold upwelling water in southern California. The physiological high temperature limit for these animals is around 68 F. At higher temperatures, red abalone will not do well reproductively. This thermal limit is probably one of the reasons why red abalone populations are generally smaller south of Point Conception.

Recently sea temperatures have been increasing and persisting for longer durations. These increased temperatures have extended beyond regular El Nino/La Nina warm water events, which temporarily reduce reproduction and settlement. In the past, we saw that red abalone could live through such thermal events and recover after the events passed. But now with projections for more extended marine heat waves, I am concerned for the welfare of red abalone. If the temperatures increase too much, lethal effects may occur.

Kelps, a primary food of abalone, also have an upper temperature range around 68 F. If the temperature increases above that value, kelps will die back, reducing food availability to abalone. Some kelps, particularly those on the northern California coast, are annuals, and may not be able to re-forest if the temperatures are elevated for extended periods. Thus, temperature affects food availability, too.

Juvenile, crevice dwelling abalone are not as dependent on kelps because they feed on bacterial films and smaller algae. It is not yet known how these other food sources do in elevated temperature regimes.

In addition, elevated sea temperature can create conditions conducive to sporadic red tide-like events, which affect and may be lethal to abalones and other invertebrates. Such occurrences have recently occurred at coves in Sonoma County, killing large numbers of red abalone. If large abalone were killed by red-tide like toxins, it's very likely the smaller crevice dwelling ones were killed as well. If this is the case, then local spawning events would be required to repopulate. But this, of course, would require that sea temperatures be conducive to allow for successful reproduction.

Indirect temperature effects can also affect red abalone. It is known that elevated sea temperatures were responsible for sea star wasting disease that drastically reduced the sea star populations along the coast. Sea stars prey on urchins that compete with abalone for food. Without sea stars, urchin populations can proliferate and actually take over and degrade former abalone habitat. Urchin barrens can develop and become persistent, depriving abalone of both food and space.

The bottom line is that elevated sea temperature can be quite problematic for red abalone. Though we don't yet know all the ramifications, all that we do know points to the high likelihood that red abalone may be imperiled along the north coast.

Another troubling problem likely to face abalones is ocean acidification, related to climate change. As CO₂ levels rise, there are increases in oceanic acidity that affect the ability of animals to extract calcium from the water. This particularly affects animals that utilize calcium in their protective shells, including abalones. For about a week after spawning, abalone larvae are platonic, and then begin to extract calcium from the sea to begin building a shell. Ocean acidification will make this process of development more difficult, as the acidity tends to keep the calcium in solution, and may make it more difficult for new cohorts of red abalone to recruit into the population.

The north coast red abalone fishery was long a successful fishery, probably self-sustaining, owing to a different and more protective set of regulations than those that governed the southern California fishery. The southern fishery allowed commercial divers to land tons of

abalone annually and allowed recreational divers to use SCUBA gear. Because oceanic conditions were better, and there was a larger recreational community, southern California abalone were no doubt subjected to higher overall fishing pressures.

In contrast, the northern fishery was conducted along a more or less remote coast with often severe oceanic conditions, including rough water and cold seas. No surface-supplied or SCUBA breathing devices were allowed. Because of these restrictions, there was a *de facto* deep-water reserve of abalone that was inaccessible to most skin divers. It was the deep reserves that provided for continual replacement of the shallow stocks that the divers could take. Because of these conditions, the northern fishery remained open many years longer than the southern.

Through time, several things changed in the northern fishery. Better access, more divers, and the increased market value of abalone caused an increase in the use of the resource, including poaching. In recent years, several changes to season and bag limits were necessary. While these regulation changes likely extended the fishery, it was ultimately the stressful oceanic conditions that were responsible for the closure of the fishery.

Judging from recent surveys of north coast red abalone, both deep and shallow (figure 1 CDW density survey - below), the condition of the red abalone population is now dire. The condition of the crevice population is unknown. To get an estimate of future conditions of the fishery, the numbers of abalone that have recently become emergent, but remain of sub-legal size can be used as an indicator. It is these individuals that will eventually, and hopefully, grow into a large enough population to once again support a fishery.

Our multi-year surveys of red abalone at Johnsons Lee, Santa Rosa Island are an example of how quickly an abalone fishery can collapse. Our studies began several years before the 1983 El Nino occurred and ended just as it was beginning. After the El Nino, we returned to Johnsons Lee to observe the red abalone population. We found that the population had declined, likely because the fishery had continued, but we did find abalone. This demonstrated that large abalone could survive a prolonged warm water episode. However, there was a severe reduction in growth of the remaining abalone, and there was no indication that reproduction and settlement had occurred. These conditions were likely the result of elevated sea water temperature that reduced food availability, depressed growth, and led to reproductive failure, which ultimately led to the population's decline collapse. (figure 2 - below).

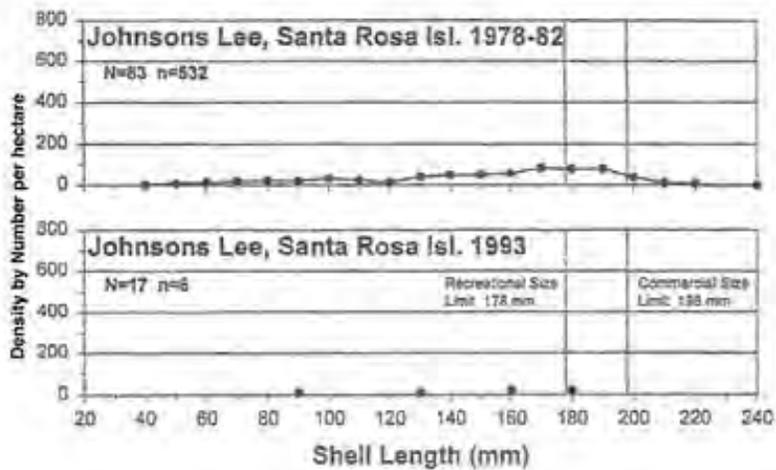
I am concerned that conditions on the north coast appear to be similar and could be highly problematic to the recovery of the red abalone brood stock.

In the absence of sufficient knowledge of the crevice population of red abalone, it would be unwise to allow for any kind of a new fishery, until it can be demonstrated that conditions exist to support the recovery of brood stock. Surveys designed to identify and document crevice and emergent red abalone populations should be added to current surveys. If this part of the population can be identified and shown to exist, I would be much more optimistic about the future of the resource, and the possibility of eventually considering another fishery. Until then, I think we need to keep with the precautionary approach of the ARMP and focus on abalone recovery.

Sincerely,

Peter L Haaker
Senior Marine Biologist
California Department of Fish & Game, Retired

Figure 2. Density by size of emergent red abalone, Johnsons Lee, Santa Rosa Island, 1978-1982 and in 1993. (Karpov et al., 1998) N = number of transects, n = number of abalone.



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ope of decline



:: 5 years to extirpation

From: Konstantin Karpov [REDACTED]
Sent: Thursday, October 22, 2020 1:55 PM
To: FGC <FGC@fgc.ca.gov>
Subject: follow-up letter letter to MRC and request for meeting.

Honorable Executive Director Mellisa Miller-Henson,

I am writing to provide a follow-up letter to Commissioners Samantha Murray and Peter Silva, expanding on my comments at the October 14 CFG Commission meeting regarding the extended fishery closure, and further clarifying concerns about the Red Abalone FMP now in development. My letter is intended to explain some of important science and background relevant to future management.

I would like an opportunity to meet in person with both commissioners via a Zoom meeting to discuss these critical matters further. Perhaps this time my colleague Peter Haaker could join us as well. Together, I think we can provide greater insight from our collective 60+ years as CDFW scientists outside of the political arena surrounding development of the FMP.

I realize both Commissioners are very busy, but both Peter and I are available on any day and time convenient to both Commissioners.

Sincerely Konstantin Karpov

To: Commissioners Samantha Murray and Peter Silva, Marine Resources Committee of the California Fish and Game Commission
From: Konstantin Karpov
Re: Red Abalone Fishery closure extension and Fishery Management Plan (FMP)
Date: Oct. 22, 2020

Honorable Commissioners Silva and Murray:

I am writing again to follow up regarding concerns with the Red Abalone FMP now in development. As a former CDFW abalone biologist, I have extensive firsthand experience and knowledge of red abalone populations, especially in northern California. I worked for the department for 30 years with a focus on abalone, so I have a close knowledge of abalone research and science. My qualifications include ten peer reviewed publications regarding abalone biology, authored along with many other highly respected abalone scientists. In 2007 I was awarded one the CDFW's highest honors, the Francis Clark Award, given to a staff member who exemplifies scientific excellence and dedication to conservation, as so famously demonstrated by Dr. Clark, who fended off commercial interests intent on liquidating California's sardines in the 1930s. I was also one of the senior authors of the state's Abalone Recovery and Management Plan.

I strongly support the MRC's recommendation to extend the interim fishery closure under Abalone Recovery and Management Plan (ARMP) rules for five years before implementing the proposed abalone Fishery Management Plan (FMP). I commend your support of a "precautionary" science-based approach to reopening this fishery, especially given that Department's most recent surveys show that northern California's red abalone population is already reduced to near extirpation (Fig. 1).

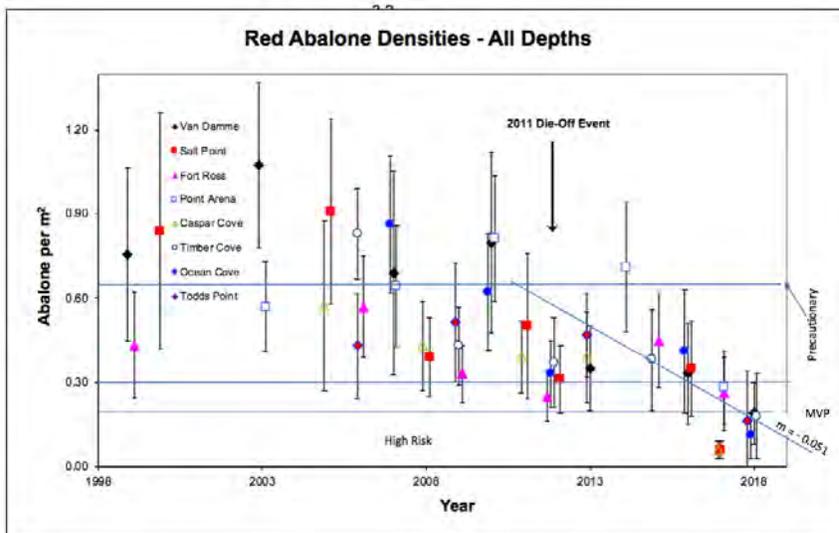


Figure 1.
CDFW abalone decline by density at northern California index sites. Slope of decline is average of all sites sampled, 2010-2018

$M = \text{slope of decline}$



NOTE: 5 years to extirpation

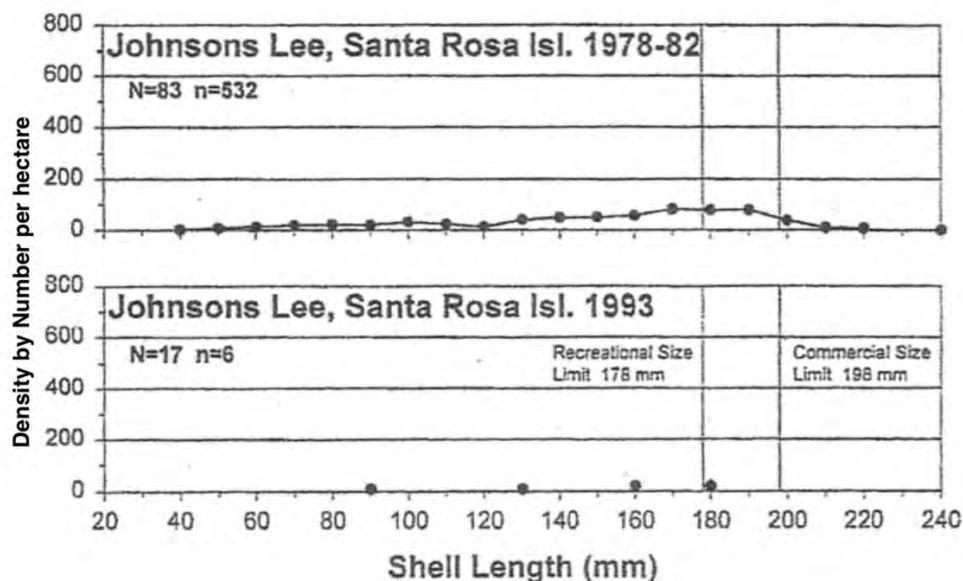
Concerns about new FMP

With regards to the new Red Abalone Fishery Management Plan now in development, I've had a chance to more carefully review the process and administrative rationale for the new FMP (*Summary of the Management Integration Process for the North Coast Recreational Abalone Fishery Management Plan*, April 2020) and want to express some specific concerns and explain some relevant background.

Limits of SPR modeling

As I've described to you before, I experienced the Department's failure to foresee the perilous decline of red abalone in southern California at Santa Rosa Island firsthand. In the late 1980s, egg and yield per recruit modeling (Tegner et al., 1989) incorrectly predicted that stocks there could persist in a sustainable fishery under then current size and take limits. In retrospect, we know that modeling had some major failings: it assumed continuous recruitment, it did not recognize that the population had already fallen below a minimum viable population (MVP), and thus it did not account for the Allee effect. Unfortunately, the red abalone stocks at that site were ultimately extirpated (Fig. 2).

Figure 2, Density by size of emergent red abalone, Johnsons Lee, Santa Rosa Island, 1978-1982 and in 1993. (Karpov et al., 1998) N = number of transects, n = number of abalone.



I am concerned that we are seeing an alarmingly parallel situation developing today in the north. In northern California, even when red abalone were abundant in the 1980s and 1990s, data strongly suggests that red abalone recruitment was episodic, with major recruitment events occurring only infrequently, on a decadal scale (Haaker et al. 2003; Karpov et al., 2001; Kashiwada and Taniguchi, 2007).

Now in northern California, as was the case at Santa Rosa Island, cumulative impacts have already reduced red abalone abundances to below MVP (<0.2 abalone per m²), a tipping point at

which the Allee effect becomes critical (Fig. 1). As you well know, the “Perfect Storm,” a mix of El Niño, disease, urchin takeover, plus marine heat waves, has led to kelp loss and abalone starvation that has profoundly affected recruitment success. With elevated water temperature, red abalone may not even be able to produce viable gametes (Vilchis et al., 2005). The impacts of these environmental stressors on red abalone is likely to be great, but it cannot be readily modeled.

For these reasons, I am concerned about the modeling in the new FMP. In a data poor environment where density surveys are not being conducted, the SPR model assumes, without evidence, that recruitment failure occurs only one out of every four years (Harford et al. 2019). The model also does not address the Allee effect of recruitment failure at densities below MVP at shallow depths (<8.4m). In addition, the model ignores the importance of historically abundant deeper water refuge populations that are now reduced to ~ 0 per m² at sites sampled in 2018. Finally, current Department scientists have raised concerns that the model is highly inaccurate for tracking recovery at current low densities.¹

Recognizing the unique biology of abalone, density surveys remain the most reliable measure of population recovery, especially at the low populations that we now have and are projected to have into the future. Density sampling has occurred at 10 locations that have represented 52% of fishing effort (Kalvass and Geibel, 2006). These surveys have been conducted at locations in both shallow and deep water, representing a critical spatial aspect of red abalone persistence. Finally, several of the locations have a long history of sampling (>30 years) and provide added statistical power of a time series (Parker et al., 1988) (Fig. 1).

Limitations of new FMP's SPR model

- Assumes continuous abalone recruitment success, ignoring published science
- Does not provide a “red flag” if stock abundance is declining towards extirpation
- Not sufficiently sensitive at low population density (<MVP)
- Cannot account for unpredictable environmental stresses

Advantages of ARMP target densities with a defined MVP

- Trigger density is science based, precludes political pressure
- Provides a “red flag” for failing recovery
- Identifies effectiveness of restoration efforts
- Monitors recovery at both shallow and deep depths

Insufficient attention to critical spatial dynamics

In addition, I am concerned that the new FMP does not recognize a critical local spatial dynamic that has long benefitted northern California’s red abalone fishery, but that is now gone: the persistent de-facto deep-water refuge of a dense abalone population that consistently replenished

¹ Dr. Laura Rogers-Bennett described the model of time to recovery as inaccurate as a predictive tool at current low densities. Recreational Abalone Advisory Committee (RAAC) July 11 2020 and July 27 2020 MRC meetings

the shallower areas through both onshore movement and additional recruitment. A fishery of free-diving without SCUBA provided a refuge from take below 8.4 meters for ~30 % of red abalone population. The proximity of this deep-water refuge at high abundances was instrumental in sustaining a multi-million pounds per year recreational fishery for so long (Karpov et al., 1998). Understanding this spatial structure will be important for future recovery of abalone.

Recommendations to improve the ARMP prior to implementing the FMP

Given these limits, I'd like to make the following recommendations:

Include density triggers above MVP

I strongly urge you to recommend that any new Red Abalone FMP include density triggers above the minimum viable population (MVP) of 0.3 abalone per m² BEFORE considering the opening of any fishery (This critical density trigger is described in on page 54 of the *Summary of the Management Integration Process for the North Coast Recreational Abalone Fishery Management Plan*). Only a density trigger can provide a strong enough “red flag” if stock abundance is declining toward extirpation.

Account for both shallow and deep-water conditions

In addition, during the first 5 years of closure, I urge the MRC to recommend that the Department rely on density surveys at both deep *and* shallow depths at sites with historical time series for “red flag” monitoring prior to considering either a de minimis or open fishery. Do not rely on yet-to-be-proven, size-based SPR as a comparable “red flag” when densities are at or below MVP. Only after densities reach > 0.3 per m² at both deep and shallow depths should fishery considerations be resumed. I recommend that existing survey sites with a record of data over time be targeted for continued sampling as a proxy for coast wide density sampling that can then be approached using SPR modeling combined with continued density surveys.

Improve “Exceptional Circumstances” plan

We are now clearly in a period of “Exceptional Circumstances” (Appendix G of the FMP). Only if and when we recover from the current period of “Exceptional Circumstances” and target densities are achieved, should a refined and revised Red Abalone FMP proceed. I urge the MRC to direct the Department to create a more robust delineation and plan for “Exceptional Circumstances.” It seems that there has been too little attention given to this critically important aspect of the FMP, especially given emerging science about projections for how climate change will affect the marine ecosystems of the California current, with extended marine heatwaves and increased likelihood of disease. I urge the MRC to recognize that climate change demands more precaution not less and to convene a group of independent scientists, preferably with expertise in how climate change will affect the marine environment and who are independent of the Department politics and the FMP process, be involved in this effort.

Need for new peer review before FMP implementation

Given the exceptional circumstances we are now in, with densities approaching extirpation a DFW sampled sites, I strongly urge the MRC to recommend another independent peer review for the red abalone FMP before it is implemented. Previous peer reviews were incomplete and occurred before the mass die off of kelp and abalone. I think this would be prudent given the lack

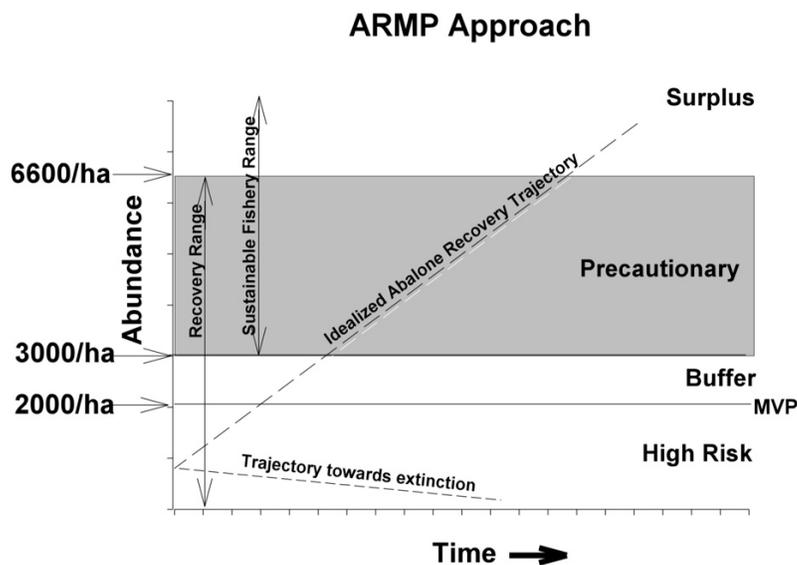
of empirical evidence supporting a de minimis fishery in the absence of density data showing real recovery.

Need greater focus on Red Abalone Recovery

In addition, I am concerned about the Department's focus on developing a fishery management plan for animals that are so depleted. Instead, the Department should be focused on recovery of these animals. Along with a precautionary approach to the fishery, recovery was an important priority of the ARMP (Fig. 3). The Commission and Department should be leading the fishing community to support recovery options instead of the fishing community pressing the Department to find ways to keep fishing on these animals that clearly can no longer support fishing pressure.

As kelp recovers, the Department should work to reestablish and recover deep-water (at depths >8.4 m) abalone populations at key *deep water refugia* sites known to have micro-topography that will likely support strong kelp recovery. (These are most of the historic sampling sites with a time series of data.) I urge the Department to engage partners to recreate these brood stocks by translocation of shallow red abalone, combined with purple urchin management using predator introduction, active relocation, or lethal removal, whichever is deemed most effective. This is where the greatest effort should be invested if we want to have a fishery in the future.

Figure 3. The 2005 Abalone Recovery and Management Plan (ARMP) takes a precautionary approach, prioritizing recovery and long-term persistence of abalone.



Consider rotational closures or pulse fishery model

If in the future conditions allow for implementation of the FMP (following recovery to minimum trigger density of 0.3 per m²), I urge the Department to consider the option of rotating area closures or a pulse fishery model (Sluczanowski, 1984). SPR surveys could then be tested empirically at specific fished locations for later confirmation using ongoing density surveys. This

approach would be more prudent and science-based, rather than to risk area-wide extirpation of red abalone by prematurely relying on the unproven SPR model.

Science based sampling during closure

Given low and declining numbers of red abalone, science-based lethal collection should be minimized and justified by agency scientists. Size, sex, and condition can all be sampled non-lethally. In a major tag and recapture study in 1971 to 1975, almost 4,000 abalone were captured, tagged, sexed, and measured for length and weight at Point Cabrillo Reserve and then returned to the same reef area from where they had been removed (Ault and DeMartini 1987). Abalone, when placed on their back, will attempt to right themselves exposing their gonads in the process. Pelican buoys were used to mark and return animals safely to scars and optimal crevice habitat with minimum disruption. Collaborative non-lethal sampling with volunteers and Department scientists should be encouraged in place of unneeded lethal taking of abalone to provide for meats thought necessary to encourage fisher participation.

Conclusion

I believe it should now be the Commission's and the Department's highest goals to recover red abalone and to keep them from becoming part of the "Sixth Extinction." I urge the MRC to take a big picture look at the present circumstances and to 1) Recommend that the Department re-focus efforts on Recovery rather than fishery management, 2) Recommend Retaining Minimum Viability Criteria and triggers in the FMP, and 3) Recommending that the Department strengthen the Exceptional Circumstances portion of the FMP.

I would like to meet with both of you, possibly including my colleague Peter Haaker, to further clarify some insights as the senior authors of the ARMP.

Sincerely,

Konstantin Karpov

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**California Fish and Game Commission
Marine Resources Committee (MRC) Work Plan
Scheduled Topics and Timeline for Items Referred to MRC
Updated November 2, 2020**

TOPIC	CATEGORY	JUL 2020	NOV 2020	MAR 2021
Planning Documents & Fishery Management Plans (FMPs)				
MLMA Master Plan for Fisheries - Implementation Updates	Master Plan Implementation	X	X	X
Red Abalone FMP / ARMP Update	FMP	X/R	X	X
California Halibut FMP (TBD)	FMP			
Regulations				
Experimental Fishing Permit Program, Phase II	Fisheries	X/R		
Kelp and Algae Commercial Harvest	Kelp		X	
Update on and possible review of California Spiny Lobster FMP implementing regulations (<i>added Feb 2019; timing TBD</i>)	FMP			
Maintenance of Preexisting Structures Within Marine Protected Areas	Marine Protected Areas	X	X/R	
California Grunion Recreational Fishing Regulations	Fisheries	X	X/R	
Aquaculture				
Aquaculture Program Planning (Information Report, Action Plan)	Planning Document		X	
Aquaculture State Water Bottom Leases: Existing & Future Lease Considerations	Current leases/planning			
Moratorium on New Aquaculture Lease Applications	New Leases		X/R	
Aquaculture Lease Best Management Practices (BMP) Plan Requirements (HOLD, TBD)	Regulations			
Emerging/Developing Management Issues				
Kelp Restoration and Recovery Tracking	Kelp			X
Invasive Non-native Kelp and Algae Species (<i>added Oct 2020</i>)	Kelp / Invasive Species		X	
Recreational Swordfish Fishing Regulations	Regulations			
Special Projects				
California's Coastal Fishing Communities	MRC Special Project	X	X	X

KEY:

X Discussion scheduled

X/R Recommendation developed; topic generally moved to FGC

California Fish and Game Commission: Perpetual Timetable for Anticipated Regulatory Actions

Updated October 20, 2020

Regulatory Change Category	Title 14 Section(s)	TC	MRC	FGC	FGC	FGC	WRC	FGC	FGC	MRC	TC	FGC	FGC	FGC	WRC	FGC	FGC	MRC	TC	FGC	FGC	WRC	FGC	FGC
		Webinar/Teleconference Nov 16, 2020	Webinar/Teleconference Nov 10, 2020	Webinar/Teleconference Dec 9, 2020	Webinar/Teleconference Dec 10, 2020	Webinar/Teleconference Jan 12, 2021	Webinar/Teleconference Jan 12, 2021	Webinar/Teleconference Feb 10, 2021	Webinar/Teleconference Feb 11, 2021	Webinar/Teleconference Mar 16, 2021	Webinar/Teleconference Apr 13, 2021	Webinar/Teleconference Apr 14, 2021	Webinar/Teleconference Apr 15, 2021	Webinar/Teleconference May 11, 2021	Webinar/Teleconference May 11, 2021	Webinar/Teleconference Jun 16, 2021	Webinar/Teleconference Jun 17, 2021	Webinar/Teleconference Jul 20, 2021	Webinar/Teleconference Aug 17, 2021	Webinar/Teleconference Aug 18, 2021	Webinar/Teleconference Aug 19, 2021	Webinar/Teleconference Sep 16, 2021	Webinar/Teleconference Oct 13, 2021	Webinar/Teleconference Oct 14, 2021
Central Valley Sport Fishing (Annual)	7.40(b)(4), (43), (66), (80)			N				D				D		A					E 8/1					
Klamath River Basin Sport Fishing (Annual)	7.40(b)(50)			N				D				D		A					E 8/1					
Waterfowl (Annual)	502			N				D				A						E 7/1						
Mammal Hunting - Deer and Antelope tag adjustments, and big game license tag drawing	360, 363, 708.19			N		D		A				E 4/1												
Marine Protected Areas (MPAs), Marine Managed Areas (MMAs), and Special Closures	632								N													E 10/1		
Commercial Pacific Herring Eggs on Kelp (Fishery Management Plan Implementation)	163, 164			E 11/30																				
Groundfish	27.30, 27.35, 27.45, 28.27, 28.28, 28.54, 28.55, 28.65, 150.16						E 1/1																	
Simplification of Statewide Inland Fishing Regulations ³	3.00, 4.00, 5.00, 5.41, 5.84, 5.86, 5.89, 7.00, 7.40, 7.50, 8.10									E 3/1														
Recreational Crab Marine Life Protection Measures	29.80, 29.85, 701			A						E 3/1														
Recreational Take of Red Abalone	29.15			A								E 4/1												
Recreational take of Sea Urchin at Caspar Cove and Taker Reef regulations ⁵	29.06			A						E 3/1														
Recreational Purple Sea Urchin emergency regulations (120 day extension)	29.06								EE 1/12															

Rulemaking Schedule to be Determined	Title 14 Section(s)	TC	MRC	FGC	FGC	FGC	WRC	FGC	FGC	MRC	TC	FGC	FGC	FGC	WRC	FGC	FGC	MRC	TC	FGC	FGC	WRC	FGC	FGC
		Webinar/Teleconference Nov 16, 2020	Webinar/Teleconference Nov 10, 2020	Webinar/Teleconference Dec 9, 2020	Webinar/Teleconference Dec 10, 2020	Webinar/Teleconference Jan 12, 2021	Webinar/Teleconference Jan 12, 2021	Webinar/Teleconference Feb 10, 2021	Webinar/Teleconference Feb 11, 2021	Webinar/Teleconference Mar 16, 2021	Webinar/Teleconference Apr 13, 2021	Webinar/Teleconference Apr 14, 2021	Webinar/Teleconference Apr 15, 2021	Webinar/Teleconference May 11, 2021	Webinar/Teleconference May 11, 2021	Webinar/Teleconference Jun 16, 2021	Webinar/Teleconference Jun 17, 2021	Webinar/Teleconference Jul 20, 2021	Webinar/Teleconference Aug 17, 2021	Webinar/Teleconference Aug 18, 2021	Webinar/Teleconference Aug 19, 2021	Webinar/Teleconference Sep 16, 2021	Webinar/Teleconference Oct 13, 2021	Webinar/Teleconference Oct 14, 2021
CA Grunion (FGC Petition #2019-014)	TBD																							
Commercial Kelp and Algae Harvest Management	165, 165.5, 705																							
Santa Cruz Harbor Salmon Fishing (FGC Petition #2016-018)	TBD																							
European Green Crab (FGC Petition #2017-006)	TBD																							
Wildlife Areas/Public Lands ⁴	TBD																							
Experimental Fishing Permit (EFP) Program Phase II)	TBD																							
Possess Game / Process Into Food	TBD																							
American Zoological Association / Zoo and Aquarium Association	671.1																							
Night Hunting in Gray Wolf Range (FGC Petition #2015-010)	474																							
Shellfish Aquaculture Best Management Practices	TBD																							
Commercial Pink Shrimp Trawl	120, 120.1, 120.2																							
Ridgeback Prawn Incidental Take Allowance	120(e)																							

KEY
 FGC = California Fish and Game Commission MRC = FGC Marine Resources Committee WRC = FGC Wildlife Resources Committee TC = FGC Tribal Committee
 EM = Emergency EE = Emergency Expires E = Anticipated Effective Date (RED "X" = expedited OAL review)
 N = Notice Hearing D = Discussion Hearing A = Adoption Hearing V = Vetting R = Committee Recommendation
 3 = Includes FGC Petition #2018-008 4 = Includes FGC Petition #2018-003 5 = Includes FGC Petition #2020-001